# **Responses of organic phosphorus fractionation to**

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# 17 S1 Material and Methods

### 18 S1.1 Organic phosphorus fractionation

The fractionation of the individual organic phosphorus (OP) pools is based on the reactivity 19 with the given extractants. The optimized OP fractionation procedure, detailed in the 20 literature,<sup>1</sup> can be simply summarized as follows: The labile pool is extracted with 0.5M 21 NaHCO<sub>3</sub> at pH 8.5. The moderately labile pool is extracted with 1.0 M HCl (HCl-OP), 22 followed by 0.5 M NaOH. The NaOH extract is acidified with concentrated HCl to separate 23 the nonlabile fraction (humic acid fraction, HAOP) from the moderately labile fraction 24 (fulvic acid fraction, FAOP). Finally, the highly resistant (Residual OP, Re-OP), nonlabile 25 fraction is determined by ashing the residue from the NaOH extraction at 550°C for 1 h, 26 followed by dissolution in 1.0 M H<sub>2</sub>SO<sub>4</sub>. In all cases, P concentration in the extracts was 27 determined colorimetrically by the phospho-molybdate method of Murphy and Riley.<sup>2</sup> Acid 28 or alkaline extracts are neutralized prior to P determinations. OP in the extracts is calculated 29 from the difference between total P (TP) and inorganic P (IP). TP in all extracts was measured 30 after an aliquot was digested with  $K_2S_2O_8 + H_2SO_4$ .<sup>3</sup> 31

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1.0 g sediments	
50ml 0.5M NaHCO <sub>3</sub> (pH=8.5), shake16h, centrifuge, filter	Aliquot Labile P <sub>i</sub> Aliquot K <sub>2</sub> S <sub>2</sub> O <sub>4</sub> /H <sub>2</sub> SO <sub>2</sub> digestion → Total Labile P
	Labile OP=Total Labile P-Labile F
50ml 1.0M HCI, shake 3h, centrifuge,filter	Aliquot → Moderately Labile P <sub>i</sub> → K <sub>2</sub> S <sub>2</sub> O <sub>g</sub> /H <sub>2</sub> SO <sub>s</sub> digestion → Total P
	HCI-OP=Total P-Moderately Labile F
¥ Rinsed with deionized water 5 min, centrifuge	➤ Discard supernatant Fulvic Acid OP
50ml 0.5M NaOH, shake 16h,centrifuge, filter	Aliquot Aliquot Aliquot $K_{,S_2O_g/H_2SO_digestion}$ Aliquot $K_{,S_2O_g/H_2SO_digestion}$ Aliquot $K_{,S_2O_g/H_2SO_digestion}$ Total P
	Humic Acid OP=Total P-Fulvic Acid O Moderately Labile OP=HCI-OP+Fulvic Acid O
Rinsed with deionized water 5 min, centrifuge	Aliquot  Discard supornatant
Ļ	Nonlabile OP=Humic Acid OP+Residual O
Ash at 550°C, 1h	Dissolve in 50ml 1.0M → Residual OP

Total P (TP) in sediments was measured by treating at 500 °C (2 h), followed by HCl extraction. Inorganic P (IP) was determined by direct extraction with 1 M HCl (16 h). OP concentrations in sediments were calculated as the difference between TP and IP (Aspila et al., 1976). The recovery (%) of OP was measured as the ratio of the OP concentration determined by extraction to that determined by ignition. The results showed that the recovery of TOP ranged from 90.25% to 110.93%, with an average of 99.81% (Table S1 and S2), suggesting that total OP in sediments was satisfactorily extracted with this procedure.

-	Conc	entration (m	g/kg)	Recovery (%)						
_	ТР	IP	ТОР	ТР	IP	ТОР				
HLH1 (n=3)	498.01	427.38	70.63	104.55	104.93	102.39				
HLH2 (n=3)	707.39	536.54	170.86	100.93	99.86	104.23				
HLH3 (n=3)	654.63	501.11	153.51	98.80	98.89	98.71				
HLH5 (n=5)	738.72	549.54	189.18	101.85	100.34	103.36				
HLH6 (n=3)	701.66	534.25	167.41	101.80	101.73	102.02				
HLH7 (n=3)	821.87	603.24	218.63	96.35	96.44	96.28				
HLH9 (n=3)	709.49	531.19	178.31	100.71	100.92	100.34				
HLH10 (n=3)	698.37	490.17	208.20	102.89	103.83	101.16				
HLH11 (n=3)	823.55	676.27	147.27	99.99	99.93	100.65				
HLH12 (n=3)	390.86	312.00	78.86	99.69	99.86	99.71				
HLH13 (n=3)	324.84	244.57	80.27	100.19	99.97	100.88				
HLH14 (n=5)	563.35	447.28	116.07	97.85	97.39	98.31				
HLH15 (n=3)	643.53	525.61	117.93	98.71	98.55	99.45				
HLH16 (n=3)	705.83	548.08	157.75	100.43	100.24	101.30				
HLH17 (n=3)	750.67	605.17	145.51	96.00	95.53	97.86				
HLH19 (n=3)	853.36	645.09	208.27	97.85	98.59	95.82				
HLH20 (n=3)	758.25	585.55	172.71	99.74	100.32	98.21				
HLH21 (n=5)	703.52	511.17	192.34	97.46	98.82	96.09				
HLH22 (n=3)	690.89	535.57	155.32	99.35	99.82	98.39				
HLH23 (n=3)	756.61	626.66	129.95	95.16	94.11	100.53				
HLH24 (n=3)	440.88	344.05	96.84	98.15	97.96	98.85				
HLH25 (n=5)	711.20	609.76	101.44	100.96	100.04	101.88				
HLH26 (n=3)	770.53	621.77	148.76	97.43	97.35	98.32				
HLH27 (n=5)	832.35	628.78	203.57	98.16	95.76	100.55				
HLH28 (n=3)	677.36	525.05	152.31	103.94	103.16	106.67				
HLH29 (n=3)	788.12	612.75	175.37	92.58	92.26	93.77				
HLH31 (n=3)	404.91	299.13	105.78	100.90	101.14	100.68				
HLH32 (n=3)	362.85	230.40	132.46	92.37	90.33	95.80				
HLH33 (n=3)	377.69	288.58	89.11	104.83	104.16	107.20				

Table S1 Concentration and recovery of TP, IP and TOP in surface sediments



	Conce	entrations (m	ng/kg)		Recovery (%)	
Depth(cm)	ТР	IP	ТОР	ТР	IP	ТОР
0-2	611.43	350.54	260.89	94.23	94.69	93.62
2-4	628.14	381.08	247.07	91.74	90.87	93.07
4-6	598.19	379.31	218.87	90.37	89.53	91.83
6-8	610.60	374.27	236.33	92.35	93.03	91.29
8-10	570.16	364.50	205.66	101.45	96.24	110.70
10-12	599.82	323.16	276.66	99.26	101.45	96.70
12-14	590.12	345.14	244.97	101.80	97.54	107.79
14-16	605.68	348.01	257.67	97.82	100.37	94.37
16-18	550.18	326.81	223.37	94.28	97.04	90.25
18-20	546.16	346.36	199.80	105.37	107.21	102.18
20-22	598.26	397.57	200.69	98.30	97.76	99.38
22-24	574.14	397.45	176.70	98.99	93.68	110.93
24-26	480.41	331.31	149.11	105.15	103.91	107.92
26-28	451.51	310.03	141.48	96.56	99.24	90.69
28-30	550.10	363.41	186.69	108.97	108.14	110.58
30-32	582.50	390.93	191.56	100.35	97.00	107.20
32-34	581.40	382.74	198.66	101.21	98.37	106.68
34-36	582.22	397.84	184.37	96.46	97.39	94.46
36-38	617.99	384.15	233.84	97.19	100.12	92.39
38-40	649.47	429.90	219.57	96.26	97.03	94.75
40-42	676.63	421.68	254.95	93.91	95.76	90.87
42-44	549.31	360.09	189.23	107.91	108.03	107.67
44-46	685.81	428.54	257.28	92.76	91.85	94.27
46-48	531.07	349.47	181.60	106.92	105.42	109.82
48-50	520.61	322.21	198.40	109.49	109.89	108.85
50-52	621.80	405.63	216.18	96.87	90.13	109.51
52-54	657.88	402.73	255.15	93.61	91.96	96.23
54-56	625.61	388.55	237.05	97.15	93.91	102.45
56-58	566.48	348.41	218.07	101.74	97.46	108.57
59-60	562.44	364.09	198.35	100.69	98.37	104.94
60-62	464.83	348.88	115.95	103.99	105.99	97.95
62-64	517.99	395.19	122.80	92.72	93.02	91.76
64-66	502.13	386.75	115.38	97.76	96.34	102.53
66-68	531.80	392.23	139.57	93.68	91.74	99.14
68-70	576.63	416.84	159.80	90.88	90.79	91.12
70-72	542.90	409.24	133.66	93.18	93.42	92.45
72-74	560.58	415.83	144.75	95.37	95.90	93.86
74-76	531.04	405.32	125.71	95.76	96.05	94.81
76-78	521.06	395.19	125.87	100.32	99.95	101.47
78-80	549.10	412.12	136.98	101.16	100.69	102.56

 Table S2 Concentration and recovery of TP, IP and TOP in sediment core

Furthermore, surface sediments samples from 5 sites (HLH5, HLH14, HLH21, HLH25, 63 HLH27) were selected, capturing the span in levels of different OP fractions, to check the 64 precision of the laboratory methods. Average values and standard deviation (SD) from the 65 analysis of 5 replicate samples for TOP, and for the five fractions determined directly by the 66 OP fractionation procedures are given in Table S3. TOP is determined as the differences 67 between the TP and TIP, which are both relatively large values compared to the OP. Thus the 68 69 precision in the TOP data was expected to be especially poor, though the average relative 70 standard deviation (RSD) turned out to be 3.3%. This is within the range achieved for the 71 other OP fractions (2.3-4.8%).

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73 Table S3 Concentration of OP fractions in sediment from HLH (average ± SD) (mg/kg)

	LOP	HCI-OP	FAOP	HAOP	ReOP	ТОР
HLH5 (n=5)	$1.88 \pm 0.05$	40.01±1.23	48.19±2.20	10.37±0.54	94.42±2.75	194.86±6.31
HLH14 (n=5)	5.00±0.20	13.28±0.43	18.98±0.88	5.71±0.18	69.92±2.07	112.89±3.32
HLH21 (n=5)	11.30±0.21	15.43±0.63	40.36±2.50	18.16±0.68	98.55±3.09	183.79±6.78
HLH25 (n=5)	2.19±0.03	24.85±0.76	22.17±0.66	12.20±0.49	41.87±1.60	103.23±2.90
HLH27 (n=5)	19.03±0.29	7.02±0.18	75.50±4.20	10.46±0.46	90.92±3.14	202.93±8.04

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All analytical results are expressed as their average value of three sample replicate measurements. The data were subjected to standard analysis of variance appropriate to a factorial randomized block design. Wherever appropriate, the treatment average were compared at 5% level of significance using least significant difference. The SD's for each OP fraction are given in Table S4. The precision and accuracy of the fractionation procedure have also been described in preceding studies.<sup>1,4</sup>

As can be seen from Table S1-S4, the recovery and SD suggest that OP determination in sediments was satisfactorily extracted with this procedure and the results were relatively accurate for OP concentrations in lake sediments.

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	LOP	HCI-OP	FAOP	HAOP	ReOP
HLH1 (n=3)	0.12	0.47	0.18	0.36	2.28
HLH2 (n=3)	0.07	2.38	1.72	1.15	2.87
HLH3 (n=3)	0.17	0.74	2.11	0.83	3.11
HLH5 (n=5)	0.05	1.23	2.20	0.54	2.75
HLH6 (n=3)	0.13	0.44	2.19	0.96	4.21
HLH7 (n=3)	0.30	0.99	2.43	0.62	5.42
HLH9 (n=3)	0.15	0.85	1.94	0.96	4.39
HLH10 (n=3)	0.14	1.38	1.90	0.99	5.39
HLH11 (n=3)	0.04	2.47	1.26	0.43	2.61
HLH12 (n=3)	0.08	0.65	0.33	0.06	2.56
HLH13 (n=3)	0.04	0.75	0.23	0.29	2.49
HLH14 (n=5)	0.20	0.43	0.88	0.18	2.07
HLH15 (n=3)	0.05	0.35	1.74	0.61	2.64
HLH16 (n=3)	0.14	0.75	1.88	0.77	3.86
HLH17 (n=3)	0.07	0.27	2.26	0.91	3.06
HLH19 (n=3)	0.37	0.74	2.66	0.62	4.80
HLH20 (n=3)	0.19	1.95	2.02	0.47	3.15
HLH21 (n=5)	0.21	0.63	2.50	0.68	3.09
HLH22 (n=3)	0.33	0.27	1.93	1.14	3.38
HLH23 (n=3)	0.29	1.66	0.87	0.31	2.88
HLH24 (n=3)	0.06	0.66	0.42	0.10	3.28
HLH25 (n=5)	0.03	0.76	0.66	0.49	1.60
HLH26 (n=3)	0.20	2.38	1.21	0.52	2.41
HLH27(n=5)	0.29	0.18	4.20	0.46	3.14
HLH28 (n=3)	0.31	0.11	1.99	0.64	4.52
HLH29 (n=3)	0.27	1.12	2.65	0.88	2.57
HLH31 (n=3)	0.02	0.55	0.35	0.18	3.92
HLH32 (n=3)	0.16	2.08	0.28	0.12	3.26
HLH33 (n=3)	0.36	0.77	0.43	0.10	2.81

 Table S4 Standard derivation of different OP fractions in the sediments (mg/kg)

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### 94 S1.2 Other Experiments and methods

95 Total inorganic carbonate (TIC) in the sediments was measured by volumetric titration with HCl (1:4, v/v). Organic oxidation by dichromate was applied to measure OM content in 96 the sediments,<sup>5,6</sup> and TOC was derived by dividing the OM content by 1.724. Allo-OC was 97 estimated using a binary model which has previously been successfully applied in Lake 98 Daihai in Inner Mongolia plateau.<sup>7</sup> Total nitrogen in the sediments was analyzed by Kjeldahl 99 method using a BÜCHI B-324 distillation unit. The concentrations of Fe, Al, Ca, Mg, Sr and 100 Ba in sediments were determined using atomic absorption spectrometer (Perkin Elmer, 101 AA800). Previous studies had established the depth-age relationship for the core of HL06, 102 using fallout radionuclide <sup>210</sup>Pb and <sup>137</sup>Cs method,<sup>8</sup> which were performed at the State Key 103

Laboratory of Lake Science and Environment at the Nanjing Institute of Geography &Limnology, Chinese Academy of Science.

## 106 S2 Results

#### 107 S2.1 Regression models

Organic matter in lacustrine sediments includes detrital terrestrial and aquatic organic 108 109 materials. There are two general sources: organic matter produced in the aquatic system via 110 photosynthesis (autochthonous), and terrestrial organic matter from the catchment basins 111 delivered via tributaries (allochthonous). Typical autochthonous sources of organic matter are 112 phytoplankton, algae, and macrophytes, while typical allochthonous sources of organic matter are plant detritus, soil organic matter and soil humus. The content of Allo-OC in 113 114 sediments is mainly controlled by the climate and environmental conditions in the catchment basin. The greater the amount of annual precipitation (MAP), the more surface or sub-lateral 115 116 runoff and thicker and denser vegetation cover is likely to be found. Both factor lead in turn to a high flux of OM from the terrestrial environment into surface waters and thus to a high 117 amount of allochthonous organic carbon (Allo-OC) in the lake sediments<sup>7</sup>. Hence the content 118 of Allo-OC in lacustrine sediments may reflect the water supply rate to the lake. Likewise, its 119 120 variations within the sedimentary column may correspond to changes in lake level. This 121 relationship is further enhanced by (1) greater input of nutrients when the water discharge is high<sup>9</sup> and (2) better conservation of organic matter when the lake level is high, favoring 122 anoxic conditions near or above the water-sediment interface.<sup>10, 11</sup> Accordingly, levels of 123 Allo-OC in sediments are an index of past environment conditions in the watershed, enabling 124 us to reconstruct historical changes in lake levels.<sup>7, 11, 12</sup> 125

A significant correlation (p<0.0099) was found between Allo-OC and the lake level of 126 127 Lake Hulun during the last 50 years (Figure S2(a)). Using this regression model the historical 128 changes in lake level of Lake Hulun were reconstructed according to the variations in the 129 Allo-OC pools in the sediment core. Salinity is a key indicator of changes in water balance, which in turn is related to the lake levels:<sup>13</sup> A significant correlation (p<0.0000) was thus 130 found between salinity and lake level in Lake Hulun over the past 50 years (Figure S2(b)). 131 Historical changes in salinity could thereby be reconstructed using the regression model 132 according to the above modeled historical lake level. The runoff from the lake basin was 133 reconstructed based on the significant correlation found between HCl-OP and runoff during 134 the last 50 years (Figure S2(c)). 135

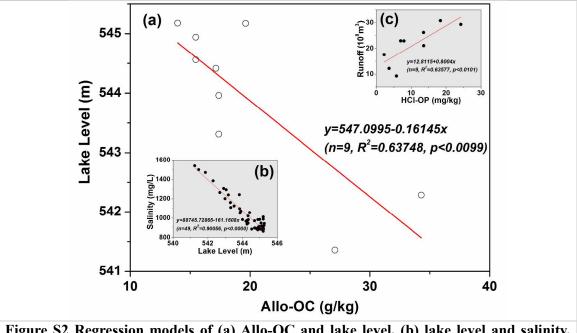
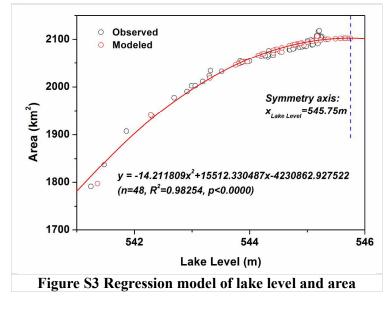


Figure S2 Regression models of (a) Allo-OC and lake level, (b) lake level and salinity, and (c) HCI-OP and runoff



An empirical quadratic function model, linking lake level to lake area, was based on the large recorded changes in lake surface area during the last 50 years (Figure S3). The axis of symmetry for the model was  $x_{Lake \ Level}$ =545.75m, indicating the maximum lake level of Lake Hulun. At this lake level the water flow out of Lake Hulun through the Dalanolom River to the Hailaer River (see Study area) balances the inflow of water from its watershed.



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### 143 S2.2 Redundancy analysis

144 Redundancy analysis (RDA), which is a constraint ordination technique, was applied to

understand the relationships between the OP fractions and environmental data. RDA is a 145 linear direct gradient analysis that uses responsive variables (species) and explanatory 146 variables (environmental parameter) in a single integrated analysis. The concentrations and 147 distribution of OP pools in surface sediments from the 13 selected lakes listed in Table S1 148 149 were used as response variables. The geographical latitude (N), the mean annual temperature (MAT) and the mean annual precipitation (MAP) of the lakes, along with the OM 150 151 concentration in surface sediments as well as the pH and the total salinity (TS) in overlying 152 water were used as the explanatory parameters. Before the RDA, we tested whether the 153 dataset was based on a linear or unimodal ordination model. Detrended constrained canonical analysis (DCCA) generated a length value for the longest gradient below 3, suggesting it was 154 suitable for linear direct gradient analysis.<sup>14</sup> The forward selection function combined with 155 Monte Carlo permutation tests was used to find the significant environmental parameters 156 157 (p < 0.05) to explain the variation in the two datasets.

To interpret the correlations between one of the OP fractions and one of the environmental parameters, Figure 2(a) may be read by projecting the arrow tips of the OP fraction perpendicular to the extended arrow of the environmental parameter (e.g. OM, Lat, MAP). The nearer the projection lies in the direction of the environmental parameter arrow, the higher the correlation. No correlation exists when the projection lies close to the origin of the ordination plot. Projections with opposite directions reveal negative correlations, with increasing strength with increasing parameter loading along the axis.

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#### 166 **S2.3 OP pools in surface sediments**

OP levels in surface sediments from Lake Hulun and the other 12 studied lakes were listed in Table S5. As can be seen from Table S5, the concentration ranges of total extracted OP (TOP) were 72.2-210 mg/kg with the means of 146 mg/kg in surface sediments from HLH. The ranges of MLOP and NLOP were 14.7-92.6 and 54.0-133 mg/kg in surface sediments from HLH with the relative contribution of 19-57% and 42-80% to TOP, respectively. The average concentration of LOP found in surface sediments from HLH was 4.54 mg/kg with the average percentages of 3.0%.

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		140	10 55		ML		scum		NL				I IAKC	s in Cn	ша	Total	ТОР
	Lakes	LOP		HCl-		FAG	)P	HA		Re-0	ОР	ML	OP	NL	OP	Extracted OP	Recovery
		mg/kg	%	mg/kg	%	mg/kg	%	mg/kg	%	mg/kg	%	mg/kg	%	mg/kg	%	mg/kg	%
	Min	0.55	0.52	2.34	1.44	4.14	5.12	1.33	1.70	41.87	32.36	14.67	19.26	54.07	41.55	72.15	
Hulun	Max	19.03	9.38	54.86	37.03	75.50	37.27	24.60	16.00	112.16	76.67	92.59	56.70	132.65	80.22	210.42	93.77-107.20
(n=29)	Mean	4.54	3.04	22.49	15.97	34.90	22.05	12.28	8.10	72.05	50.83	57.39	38.03	84.33	58.94	146.26	
	QL1	16.64	4.80	43.54	12.56	21.34	6.16	12.63	3.64	252.53	72.84	64.88	18.71	265.16	76.49	346.68	
Qilu	QL3	9.01	2.54	48.66	13.74	21.24	6.00	7.50	2.12	267.68	75.60	69.90	19.74	275.18	77.71	354.09	100.33-109.31
	QL5	8.28	2.80	51.99	17.59	13.58	4.60	9.52	3.22	212.12	71.79	65.57	22.19	221.64	75.01	295.49	
Erhai		8.70	3.76	15.30	6.61	57.20	24.73	17.80	7.70	132.30	57.20	72.50	31.34	150.10	64.89	231.30	98.9
Hongfeng		18.30	5.96	52.30	17.02	44.00	14.32	28.50	9.27	164.20	53.43	96.30	31.34	192.70	62.71	307.30	101.1
Charachart	СНЗ	2.02	1.44	15.66	11.20	48.87	34.94	16.39	11.72	56.93	40.70	64.53	46.14	73.32	52.42	139.87	100.91-105.03
Chenghai	CH5	3.03	2.01	12.63	8.37	44.68	29.61	32.63	21.63	57.92	38.39	57.31	37.98	90.55	60.01	150.89	
Baihua		17.50	4.27	123.40	30.14	47.30	11.55	35.50	8.67	185.70	45.36	170.70	41.70	221.20	54.03	409.40	96.7
T	L2	34.76	8.85	47.60	12.11	11.84	3.01	58.67	14.93	240.10	61.10	59.44	15.13	298.77	76.03	392.97	94 71 100 26
Lugu	L8	43.16	11.43	57.12	15.12	42.61	11.28	35.51	9.40	199.26	52.76	99.73	26.41	234.77	62.16	377.66	84.71-100.36
	Gonghu bay	8.60	5.97	41.10	28.54	27.00	18.75	23.90	16.60	43.40	30.14	68.10	47.29	67.30	46.74	144.00	99.5
Taihu	Meiliang bay	11.00	5.63	63.90	32.69	29.60	15.14	21.30	10.90	69.70	35.65	93.50	47.83	91.00	46.55	195.50	100.3
	Yuantouzhu bay	9.30	6.02	42.60	27.55	30.50	19.73	26.00	16.82	46.20	29.88	73.10	47.28	72.20	46.70	154.60	94.3
Chashu	C1	16.18	7.96	41.75	20.54	47.01	23.13	13.22	6.50	85.10	41.87	88.76	43.67	98.32	48.37	203.26	97.89-99.01
Chaohu	C14	13.83	8.70	65.21	41.03	17.40	10.95	16.39	10.31	46.12	29.02	82.61	51.97	62.51	39.33	158.95	97.89-99.01
Oinghai	Q4	4.87	2.80	24.72	14.22	24.15	13.89	48.33	27.80	71.78	41.29	48.87	28.11	120.11	69.09	173.85	04 49 00 54
Qinghai	Q7	11.20	6.08	17.22	9.35	18.94	10.28	52.65	28.59	84.16	45.70	36.16	19.63	136.81	74.28	184.17	94.48-99.56
Wuliangsu	ıhai	2.53	1.94	28.27	21.72	29.72	22.84	12.42	9.54	57.21	43.95	58.00	44.56	69.63	53.50	130.16	92.83-105.51
Jingpo		5.50	2.27	98.00	40.45	60.00	24.77	37.50	15.48	41.25	17.03	158.00	65.22	78.75	32.51	242.25	104.27
Wudaliand	chi	43.00	9.93	155.50	35.91	100.00	23.09	20.75	4.79	113.75	26.27	255.50	59.01	134.50	31.06	433.00	97.43

Table S5 OP pools in surface sediments from Lake Hulun and other lakes in China

**Data sources:** Data of OP fractions in Lake Qilu, Chenghai, Lugu, Chaohu, Qinghai, Wudalianchi and Jingpo from Hou et al,<sup>4</sup> OP fractions in Lake Erhai, Hongfeng, Baihua and Taihu (bays of Gonghu, Meiliang and Yuantouzhu) from Zhang ea al,<sup>1</sup> OP fractions in Lake Hulun and Wuliangsuhai were determined in this work. 

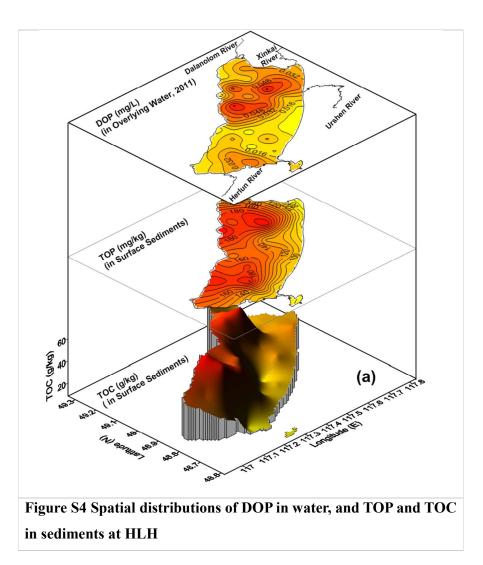
## 181 S3 Discussion

#### 182 **3.1 Spatial distribution of OP**

The spatial distribution of TOC and TOP in the HLH surface sediments showed clear 183 decreasing trends from the west to the east lake districts (Figure S4). Geomorphologically, 184 the relatively hilly terrains protected by only a semi-arid grassland with less than 40% 185 186 vegetation cover on the west bank of HLH allow for enhanced soil erosion during rainy 187 season. As a result, a large import of suspended matter, associated OM and nutrients from the HLH west bank are anticipated. This loading of eroded organic matter and nutrients, 188 189 transported during periods of overland flow, is believed to be one of the main nutrient sources 190 of HLH. Furthermore, the prevailing wind direction is from the 150 km long northwestern shore line to the southeast. The amounts of loess and dry grass deposited into HLH are 191 estimated to be  $6.44 \times 10^4$  t and 3350 t each year. Previous studies have documented that silty 192 clay is the dominating particle size found in the sediments on the northwest lake bottom, 193 while it is mainly sandy deposit on the south and east lake bottom.<sup>15</sup> Therefore, on the one 194 hand, there are more inputs of loess and dry grass driven by the prevailing wind in the west 195 lake districts; on the other hand, the inputs will be conserved effectively due to the finer grain 196 197 size in these lake districts. The decreasing trends of TOC and TOP from west to east in the 198 surface sediments from HLH is thus likely due to the combined effect of larger influx of 199 eroded soil material and loess along with dry grass in the west than in the east.

Actually, dissolved OP (DOP) constitutes a major fraction of dissolved phosphorus in 200 surface waters, comprising up to 75% of total dissolved P (TDP).<sup>16-18</sup> This fraction is believed 201 202 to represent an important potential reservoir for the more bioavailable inorganic phosphorous. 203 Over the past decades, DOP has therefore been increasingly recognized as a critical driver for aquatic biological production and thus ecosystem structure and function.<sup>19, 20</sup> DOP in water 204 above the sediment in HLH accounted for as much as 24-54% of TDP, implying that this 205 fraction may play an major role as a potential P source for the primary producers.<sup>21, 22</sup> The 206 level of P levels in water column is mainly governed by the balance between allochthonous P 207 208 inputs from the watershed and internal loading from the sediments on the one side and assimilation followed by sedimentation and adsorption across the interface of 209 water-sediments, as well as runoff on the other. The spatial distribution pattern of DOP differs 210 from the TOC and TOP (Figure S4). The higest DOP levels were found in areas close to the 211 lake inlets, suggesting that the main sources of DOP to the lake is allochthonous, through the 212 213 main tributary streams, rather than internal cycling from the sediments. Also the midwestern part of the lake had elevated levels of DOP. DOP in aquatic environments may come from the secretions and excretions of living bacteria, plants, and animals or from the decomposition of dead organisms.<sup>23</sup> The elevated DOP levels in the Midwest can thus be explained by the measured higher biomass levels of phytoplankton in this part of the lake than in the southeast.<sup>24</sup> Therefore, both terrestrial inputs and biological factors are found to jointly govern the distribution pattern of DOP in overlying water from HLH.







## 222 3.2 Impacts of Fe, Al and Ca

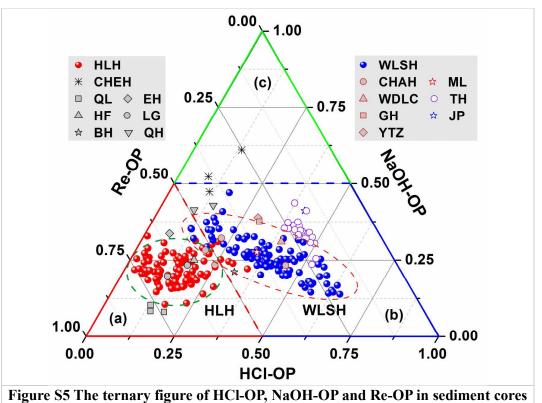
P extracted by 1.0 M HCl includes P bound to oxides/hydroxides of Fe and Al.<sup>25, 26</sup> Accordingly, acid extractable OP (HCl-OP) are derived principally from the adsorption of OP to Fe/Al oxides and hydroxides or from the formation of OP salts co-precipitated with calcium ions.<sup>27</sup> Ferric oxide can strongly adsorb OP on their surfaces in soils<sup>28</sup> and sediments,<sup>29</sup> even can play a governing role in most acidic soils and sediments. According to

our previous work, the average concentrations of Fe-P and Al-P were 1.54 and 2.04 mg/kg in 228 sediment cores from HLH, respectively; while the pools of Fe-P and Al-P were much higher 229 in the sediments from Lake WLSH, i.e. 11.10 and 7.15 mg/kg, respectively.<sup>30</sup> Therefore, both 230 Fe-P and Al-P significantly correlated with HCl-OP and MLOP in the WLSH sediment cores, 231 respectively; while no significant correlations were found in HLH. It suggested that Fe and Al 232 oxyhydroxides play a more important role in OP accumulation and fractionation in Lake 233 WLSH than in Lake Hulun<sup>30</sup> due to that the amount of iron oxide in the substrate governs the 234 extent of OP adsorption.<sup>31, 32</sup> 235

Calcium (Ca) is usually more abundant than Fe and Al in aquatic environment. This is 236 why P extracted by HCl in sediments is mainly found to be bound to Ca (mainly apatite).<sup>25, 26</sup> 237 P bound to Ca is fractionated to ACa-P (authigenic carbonate fluorapatite + biogenic apatite + 238 CaCO<sub>3</sub>-associated P) and De-P (detrital apatite P).<sup>33</sup> Calcium ions and its minerals can also 239 complex with OP in soils or sediments<sup>34, 35</sup> For instance, precipitation of calcium salts with 240 calcite occurs at even very low concentrations of mvo-inositol hexakisphosphate<sup>36</sup> due to the 241 simultaneous occurrence of adsorption and precipitation of insoluble calcium-phytate.<sup>37</sup> This 242 may explain why the OP content is positively correlated to the calcium content.<sup>34</sup> As can be 243 seen from Figure S5, the samples from WLSH and HLH were clustered in each their groups. 244 245 Almost all of the points of Re-OP and HCl-OP in HLH sediment cores were found in the ranges of 50-70% and ~25%, respectively. As for HLH, Re-OP and HCl-OP in the ranges of 246 20-50% and 12-60%, respectively.<sup>30</sup> Furthermore, as well as HLH, the points of QL (Lake 247 Oilu),<sup>4</sup> EH (Lake Erhai),<sup>1</sup> HF (Lake Hongfeng)<sup>1</sup> and LG (Lake Lugu)<sup>4</sup> presented in Area (a) 248 (Figure S5). Comparatively, most lakes in Southwestern China Plateau were far away from 249 local cities, where weathering and erosion were the main processes in the carbonate 250 watershed.<sup>38</sup> It indicated that calcium ions and their minerals played a key role in regulating 251 OP fractionations in these lakes due to co-precipitation with calcite.<sup>30, 39</sup> 252

As can be seen from Figure S5, as well as WLSH, the points standing for Lake Chaohu 253 (CHAH),<sup>40-42</sup> Lake Gonghu (GH),<sup>43</sup> Lake Taihu (TH),<sup>40, 42</sup> Lake Meiliang (ML),<sup>43</sup> Lake 254 Yuantouzhu (YTZ),<sup>44, 45</sup> Lake Jingpo (JP) and Lake Wudalianchi (WDLC)<sup>46</sup> distributed in 255 256 Area (b) (Figure S5). These lakes including WLSH shared the same characteristics that excess nutrients, especially P, were input by wastewater discharge and agricultural runoff 257 from the neighboring cities, which subsequently enhanced the association and adsorption of P 258 on Fe/Al oxyhydroxides.<sup>40</sup> Previous publications showed that the fraction of P bound to Al 259 and Fe oxyhydroxides was a major fraction, even the dominant fraction, in sediments from 260 CHAH,<sup>40-42</sup> GH,<sup>43</sup> TH,<sup>40, 42-45</sup> JP and WDLC.<sup>46</sup> In light of the OP extraction scheme, HCl-OP 261

includes OP fractions adsorbed to Fe and Al oxyhydroxides or OP salts co-precipitated with
 calcium ions. As a result of the mentioned above, it suggested that OP fractionations were
 jointly regulated by Ca, Fe and Al and their minerals in sediments (Figure S5).<sup>30</sup>



265

The vertical distribution of TP and TOC in the core of HLH15 is shown in Figure S6. Significant correlations between TOP and TP were found both in the surface sediments and in the sediment core from Lake Hulun (Figure S7). Mean annual temperature (MAT) and ratios of E/P (evaporation/precipitation) and their trends in the Lake Hulun basin from 1960's till present are depicted in Figure S8.

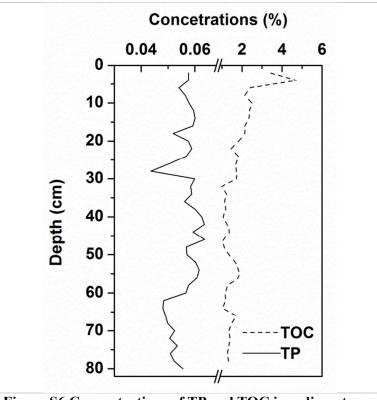
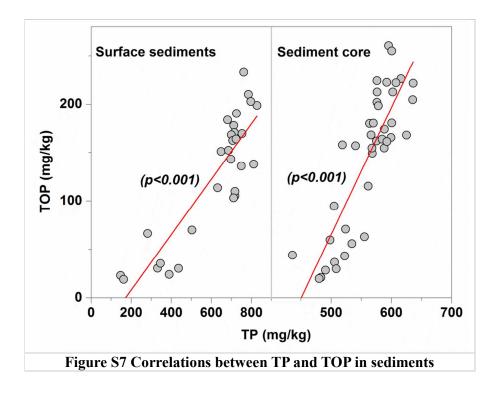
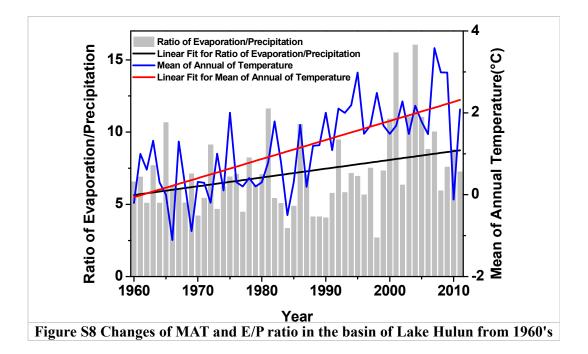


Figure S6 Concentrations of TP and TOC in sediment core





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