

Supporting Information

CdTe quantum dots-encapsulated ZnO nanorods for high-efficient photoelectrochemical degradation of phenols

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Experimental section

Synthesis of tris-1,10-phenanthroline complexes of cobalt(II), [Co(Phen)₃]Cl₂.

Tris-1, 10-phenanthroline complexes of cobalt(II), [Co(Phen)₃]Cl₂, was prepared by mixing CoCl₂ and 1,10-phenanthroline (Phen) in a 1:3 molar ratio.

Synthesis of MPA-capped CdTe QDs. The MPA-capped CdTe QDs were prepared according to a previously published procedure from our laboratory.^{S1} 0.986 g of Cd(ClO₄)₂·6H₂O was dissolved in 125 mL of distilled water, and then 0.43 mL of MPA was added. The pH value was adjusted to 11.2 by adding 2 M NaOH. The solution was placed in a three-neck flask and deaerated by N₂ gas for about 30 min. H₂Te gas generated by the mixture of 0.2 g of Al₂Te₃ and 15 mL of 0.5 M H₂SO₄ in another three-neck flask was directed into the prepared solution. The mixture was then refluxed at 105 °C for 48 h in N₂ gas. The absorbance shoulder of MPA-capped CdTe QDs was located at 545 nm, while the photoluminescence peak of QDs was situated at 595 nm (Fig. S1). The size and concentration of CdTe QDs in the solution were estimated from the UV-Vis spectrum to be around 3.24 nm and 6.99×10^{-5} M, respectively, according to the Peng's empirical equation.^{S2}

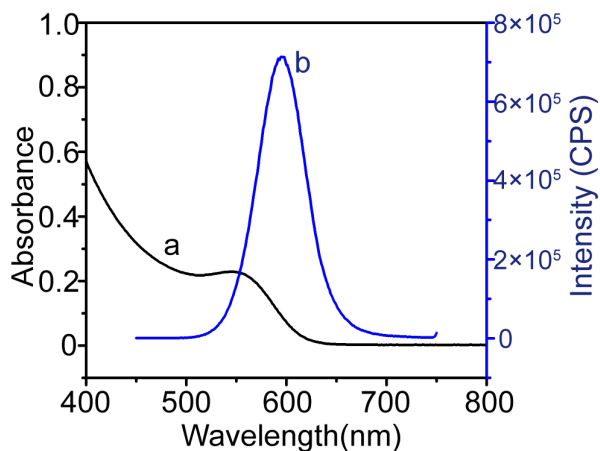


Figure S1. UV-Vis (black curve, a) and luminescence spectra (blue curve, b) of MPA-capped CdTe QDs in solution.

Determination of phenol concentration. The concentration of phenol was determined based on the Emerson's method.^{S3} At each checkpoint, 500 μ L of solution from the reactor was added to a 25 mL volumetric flask. Then, 500 μ L of

4-aminoantipyrine solution (2%), 500 μ L of potassium ferricyanide (8%) and 250 μ L of ammonia were added to the volumetric flask and diluted with deionized water to volume and mixed. The volumetric flask was incubated in the dark for 10 min. The absorbance at 510 nm was recorded.

Identification of the reaction intermediates. The reaction intermediates were analyzed by gas chromatography coupled with mass spectrometry (GC-MS, Agilent 7890A and 5975C, Agilent Tech.) using a 30 cm HP-5 column (Hewlett-Packard). In order to detect the intermediate products, GC-MS measurements were performed as follows. After reactions were stopped, the solution (15mL) was dried by freezing-drying. The resulting material was dissolved by diethyl ether and precipitate was separated by centrifugation. The pellucid solution was dried and derivatized by 50 μ L MSTFA and 0.5 μ L TMIS at 62 $^{\circ}$ C for 70 min.^{S4} Then 1 μ L was analyzed on the GC-MS. GC-MS analyses were carried out on a GC ultra gas chromatograph (7890A, Agilent Technologies) using a 30 cm HP-5 column, coupled with a mass spectrometer (5975C, Agilent Technologies). The injector port was set for split operation at 280 $^{\circ}$ C. The temperature program of the column was as follows: at 50 $^{\circ}$ C, hold time = 2 min; from 50 to 150 $^{\circ}$ C, rate = 30 $^{\circ}$ C/min. Because the background produced by derivatization was complex, 15 mL 0.07 M Na₂SO₄ solution was used to do the blank test using the same method.

Part 1 Preparation and characterization of CdTe QD encapsulated ZnO nanorod arrays.

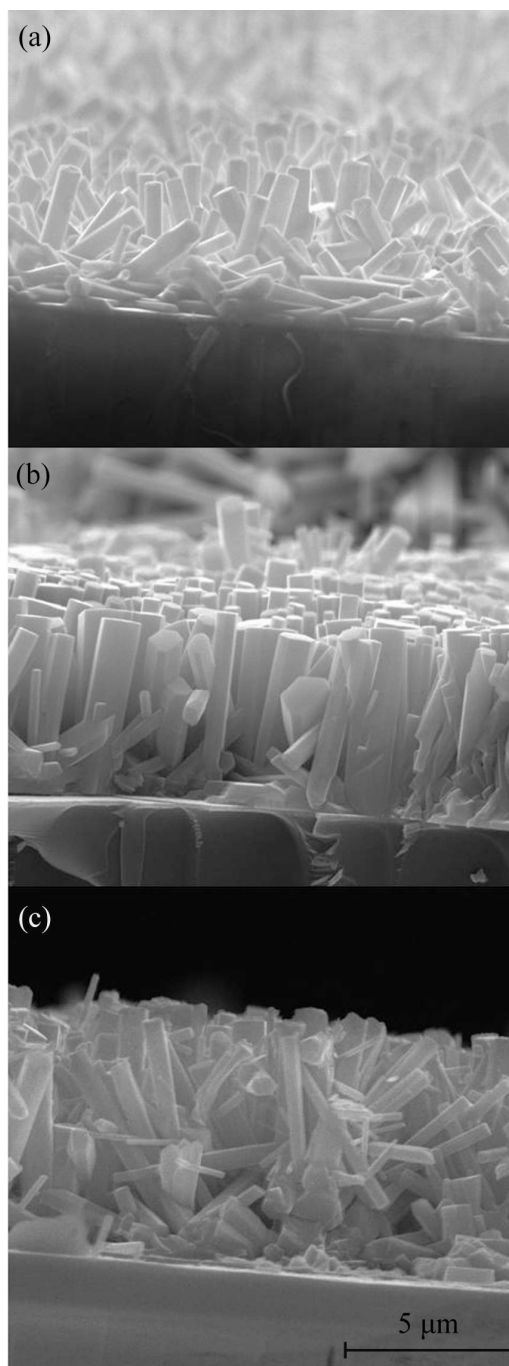


Figure S2. SEM images of ZnO nanorod arrays with different growth times: (a) 5 h, (b) 10 h, and (c) 15 h. The length of these nanorods is 3.4, 5.2 and 8.1 μm when the reaction time is 5, 10 and 15 h, respectively.

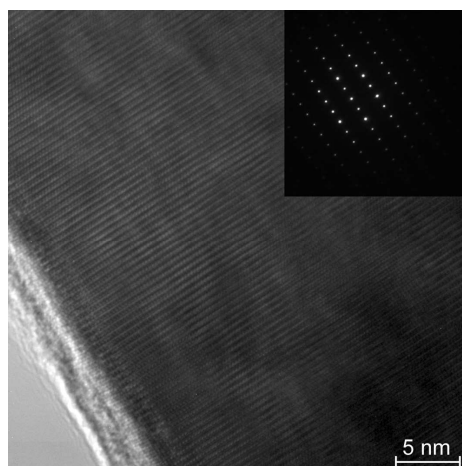


Figure S3. HRTEM image and corresponding diffraction pattern (inset) of ZnO nanorod arrays fabricated on ITO substrates by a hydrothermal method with pH 8.4 at 90 °C for 2.5 h, which exhibits the single crystalline structure with the [0001] growth direction.

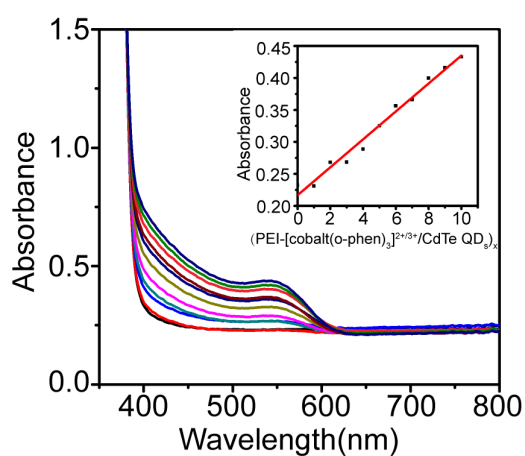


Figure S4. UV-Vis spectra of growing $(\text{PEI}[\text{cobalt}(\text{o-phen})_3]^{2+/3+}/\text{CdTe QD})_x$ multilayers (from bottom to up: $x = 0-10$). Inset shows a plot of 550 nm versus the number of bilayers, x . Regular layer growth is revealed by a linear dependence of the absorbance determined at 550 nm versus x , and the absorbance increase for one bilayer of $\text{PEI}[\text{cobalt}(\text{o-phen})_3]^{2+/3+}/\text{CdTe QDs}$ at 550 nm is 0.021 ± 0.0003 .

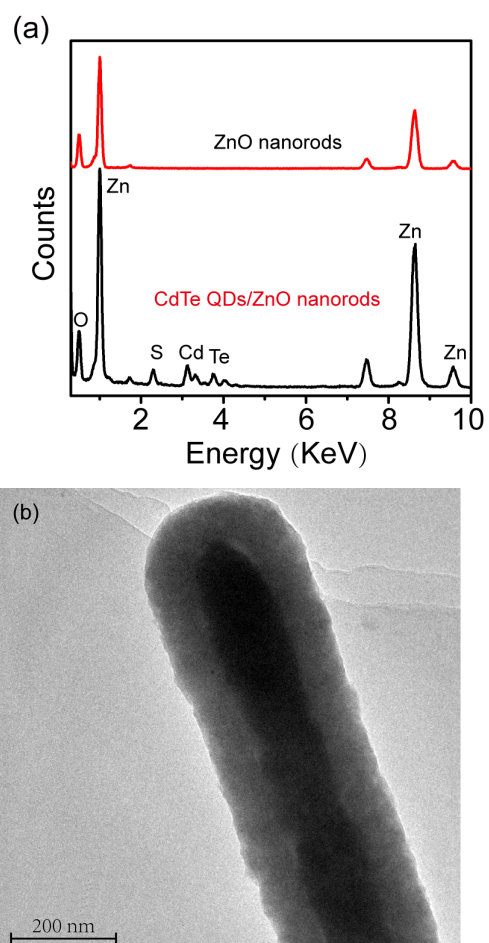


Figure S5. (a) EDX spectra of ZnO nanorod arrays and 20 bilayers of $[\text{cobalt}(\text{o-phen})_3]^{2+/3+}$ -PEI/CdTe QD encapsulated ZnO nanorod arrays. (b) TEM image of 20 bilayers of $[\text{cobalt}(\text{o-phen})_3]^{2+/3+}$ -PEI/CdTe QD encapsulated ZnO nanorod arrays.

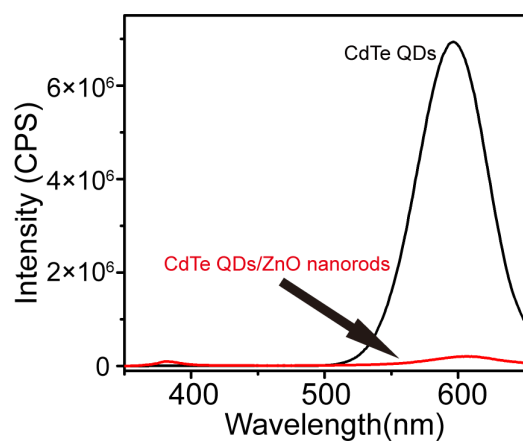


Figure S6. Luminescence spectra of MPA-capped CdTe QDs in solution (black curve) and ZnO nanorods grown on the ITO electrode after coated with 20 [cobalt(o-phen)₃]^{2+/3+}-PEI/CdTe QD bilayers (red curve).

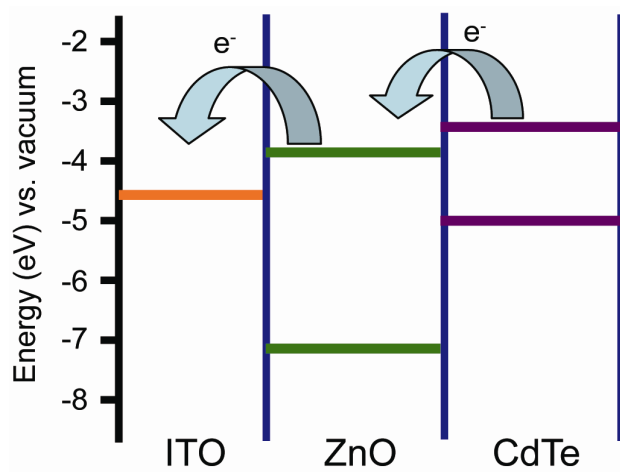


Figure S7. Schematic diagram of charge transfer at the CdTe-ZnO interface.

Part 2. Photoelectrocatalytic performance of CdTe QD encapsulated ZnO nanorod arrays for phenol degradation

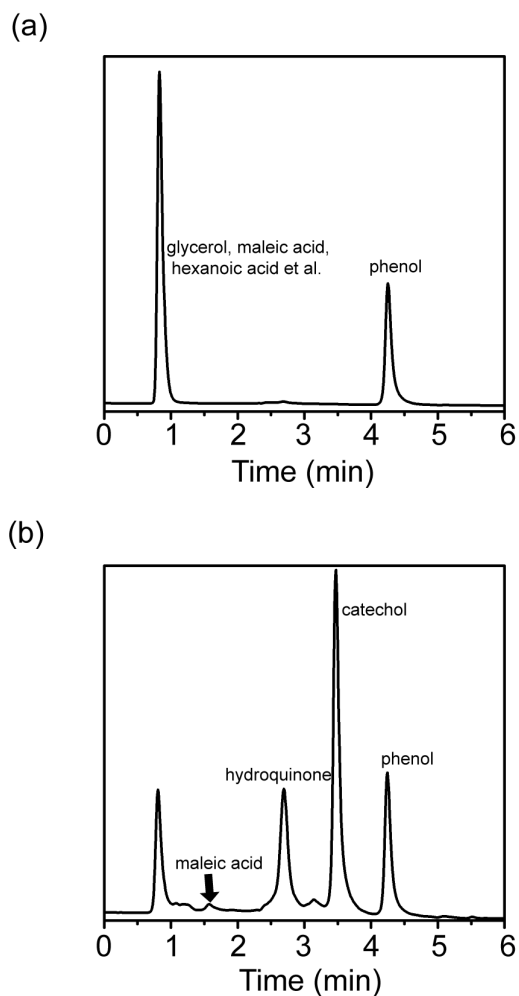


Figure S8. HPLC chromatogram of (a) photoelectrocatalytic degradation (reaction time: 2.5 h) and (b) photocatalytic degradation (reaction time: 5 h) of phenol using CdTe QD encapsulated ZnO nanorod arrays.

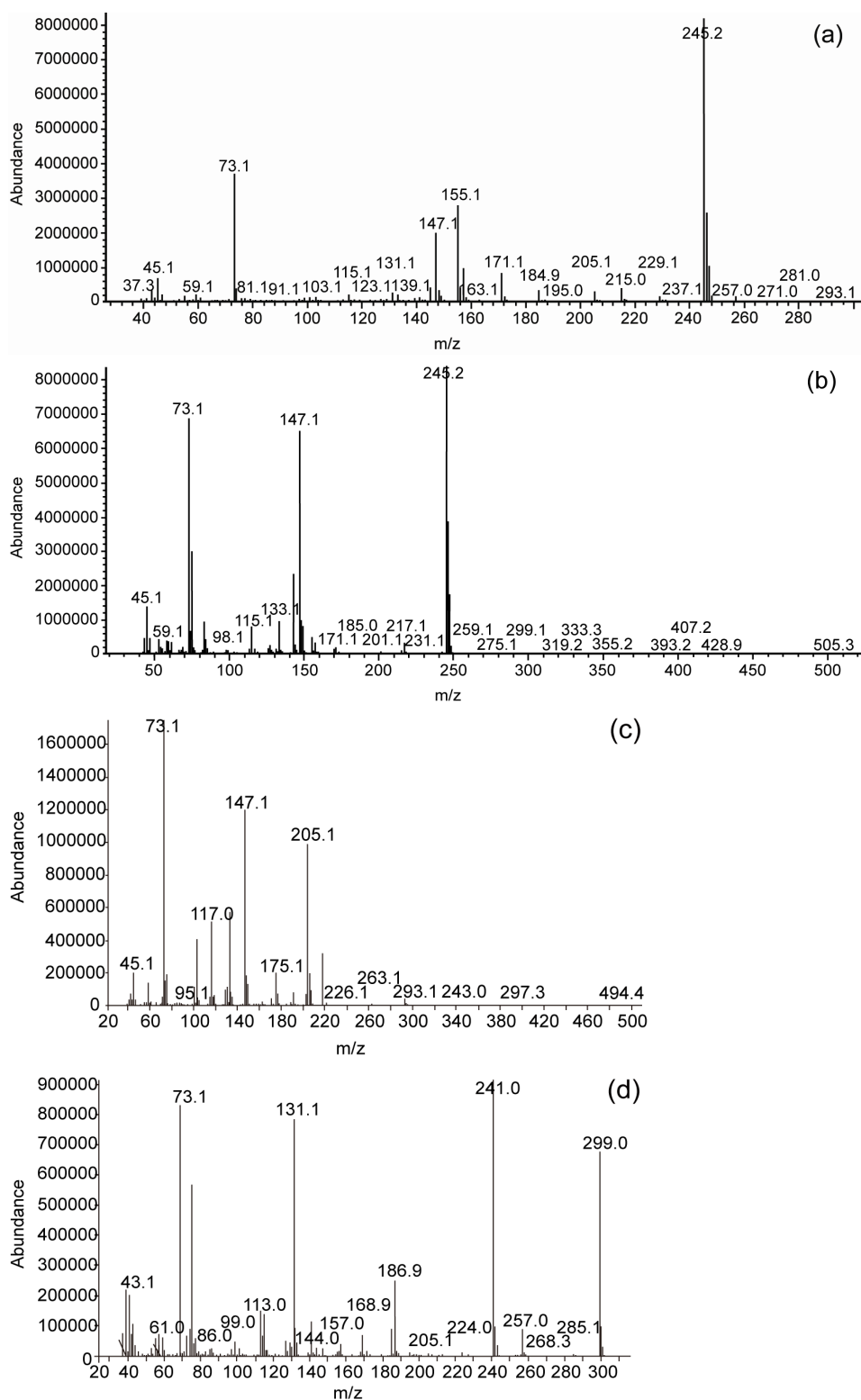


Figure S9. Mass spectrum of some products after derivatization by MSTFA and TMIS: (a) hydroxy-propyl acid, (b) maleic acid, (c) glycerol, and (d) hexanoic acid, resulting from phenol degradation by CdTe QD encapsulated ZnO nanorod arrays.

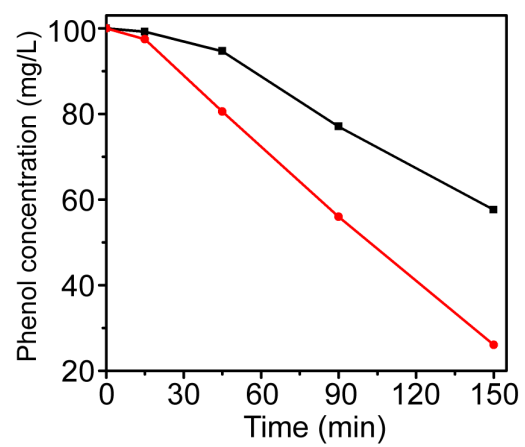


Figure S10. Phenol degradation in the photoelectrocatalytic process under oxygen (●) or nitrogen atmosphere (■).

Table S1 Comparison of removal efficiency of phenol by chemical oxidation, biological methods, photocatalytic degradation and electrochemically assisted photocatalysis.

Catalyst	Reaction conditions	Removal efficiency	TOC removal	Intermediates	Ref.
Fenton					
Fe ²⁺ and H ₂ O ₂	[Fe ²⁺] ₀ =1 mg/L, [H ₂ O ₂] ₀ =500 mg/L, pH ₀ =3	Completely degraded within 4 h for 50 mL of 100 mg L ⁻¹ phenol solution	6% within 4 h	Catechol, hydroquinone, p-benzoquinone, oxalic acid, formic acid, maleic acid, et al.	S5
mesoporous Co ₃ O ₄ /MnO ₂ nanoparticles	Catalyst loading of 0.1 g/L, oxone loading = 0.5 g/L	Completely degraded within 100 min for 500 mL of 25 mg L ⁻¹ phenol solution	60-70% within 1.5h	4-hydroxybenzoic acid, 1,2-dihydroxybenzene, p-benzoquinone	S6
Biodegradation					
<i>P. australis</i> -strain IT-4 and <i>P. australis</i> plant	10 ⁵ colony forming units (CFU) (g dry sediment) ⁻¹ , 28 °C, 16:8 h light/dark	94.8% within 21 d for 100 mg [kg dry sediment] ⁻¹ 4- <i>tert</i> -OP		2,2,4-trimethyl-1-pentanol, hydroquinone, 2-octylhydroquinone	S7
Coculture of PG-02 and PG-08	2.0-2.3×10 ⁶ colony forming units (CFU) mL ⁻¹ , 25 °C	Completely degraded within 29 h for 100 mL of 250 mg L ⁻¹ phenol solution			S8
Photodegradation					
Pt/I-TiO ₂	400 W dysprosium lamp, visible light, Catalyst loading of 1.0 g/L	62% within 2.5h for 1.5 L of 23.6 mg L ⁻¹ phenol solution	Ca. 38% within 2.5h	Quinone, oxalic acid, formic acid	S9
Cu ₂ O/TiO ₂ nanosheets	λ>420 nm, Catalyst loading of 2.5 g/L	Ca. 77% within 2.5h for 40 mL of 10 mg L ⁻¹ phenol solution			S10
Electrochemically assisted photocatalysis					
Nanocrystalline TiO ₂ electrode	130 mW/cm ² , +0.7 or 1.1 V, 9 cm ²		Ca. 40% within 2.5h for 10 mL of 100 mg L ⁻¹ phenol solution	Hydroquinone and benzoquinone	S11
p-SiNW/TiO ₂ cathode	λ>422 nm, 94.5 mW/cm ² , -2.0 V (vs Ag/AgCl), 3 cm ²	80.7% within 100 min for 30 mL of 10 mg L ⁻¹ phenol solution	Up to 62% in 6 h	benzoquinone	S12
Bifunctionalized dye-sensitized	420 nm<λ< 500 nm, 4 mW/cm ²	Ca. 90% within 2h for 14 mL of 10 mg		Hydroquinone and benzoquinone	S13

TiO ₂ film	+0.5 V, 2 cm ²	L ⁻¹ phenol solution							
Bi ₂ WO ₆ films	500-W Xe lamp,	about 78% in 3 h for	about 66% in						S14
with nanoleaflike structures	+0.6 V (vs SCE), 3.75 cm ²	15 mL of 10 mg L ⁻¹ phenol solution	3 h						
CdTe Quantum Dots-encapsulated ZnO Nanorods	$\lambda > 400$ nm, 130 mW/cm ² , +1.0 V (vs SCE), 4.15 cm ²	75% within 2.5h for 50 mL of 100 mg L ⁻¹ phenol solution	Up to 53.2% within 2.5h	hydroxy-propyl acid, glycerol, maleic acid and hexanoic acid	This study				

Reference:

- (S1) Li, X.; Zhou, Y.; Zheng, Z.; Yue, X.; Dai, Z.; Liu, S.; Tang, Z. Glucose Biosensor Based on Nanocomposite Films of CdTe Quantum Dots and Glucose Oxidase. *Langmuir* **2009**, *25*, 6580-6586.
- (S2) Yu, W. W.; Wang, Y. A.; Peng, X. G. Formation and Stability of Size-, Shape-, and Structure-Controlled CdTe Nanocrystals: Ligand Effects on Monomers and Nanocrystal. *Chem. Mater.* **2003**, *15*, 4300-4308.
- (S3) Peng, C.; Jiang, B.; Liu, Q.; Guo, Z.; Xu, Z.; Huang, Q.; Xu, H.; Tai, R.; Fan, C. Graphene-templated Formation of Two-dimensional Lepidocrocite Nanostructures for High-efficiency Catalytic Degradation of Phenols. *Energy Environ. Sci.* **2011**, *4*, 2035-2040.
- (S4) Guo, Z. F.; Ma, R. X.; Li, G. J. Degradation of Phenol by Nanomaterial TiO₂ in Wastewater. *Chem. Eng. J.* **2006**, *119*, 55-59.
- (S5) Zazo, J. A.; Casas, J. A.; Molina, C. B.; Quintanilla A.; Rodriguez, J. J. Evolution of Ecotoxicity upon Fenton's Oxidation of Phenol in Water. *Environ. Sci. Technol.* **2007**, *41*, 7164-7170.
- (S6) Liang, H. W.; Sun, H. Q.; Patel, A.; Shukla, P.; Zhu, Z. H.; Wang, S. B. Excellent Performance of Mesoporous Co₃O₄/MnO₂ Nanoparticles in Heterogeneous Activation of Peroxymonosulfate for Phenol Degradation in Aqueous Solutions. *Appl. Catal. B: Environ.* **2012**, *127*, 330-335.
- (S7) Toyama, T.; Murashita, M.; Kobayashi, K.; Kikuchi, S.; Sei, K.; Tanaka, Y.; Ike M.; Mori, K. Acceleration of Nonylphenol and 4-tert-Octylphenol Degradation in Sediment by *Phragmites australis* and Associated Rhizosphere Bacteria. *Environ. Sci. Technol.* **2011**, *45*, 6524-6530.
- (S8) Jiang, H. L.; Tay, J. H.; Maszenan, A. M.; Tay, S. T. Enhanced Phenol Biodegradation and

- Aerobic Granulation by Two Coaggregating Bacterial Strains. *Environ. Sci. Technol.* **2006**, *40*, 6137-6142.
- (S9) He, Z. Q.; Xie, L.; Tu, J. J.; Song, S.; Liu, W. P.; Liu Z. W.; Fan, J. Q.; Chen, J. M. Visible Light-Induced Degradation of Phenol over Iodine-Doped Titanium Dioxide Modified with Platinum: Role of Platinum and the Reaction Mechanism. *J. Phys. Chem. C* **2010**, *114*, 526-532.
- (S10) Liu, L. C.; Gu, X. R.; Sun, C. Z.; Li, H.; Deng, Y.; Gao, F.; Dong, L. In situ loading of ultra-small Cu₂O particles on TiO₂ nanosheets to enhance the visible-light photoactivity. *Nanoscale* **2012**, *4*, 6351-6359.
- (S11) Oliveira, H. G.; Nery, D. C.; Longo, C. Effect of Applied Potential on Photocatalytic Phenol Degradation Using Nanocrystalline TiO₂ Electrodes. *Appl. Catal. B: Environ.* **2010**, *93*, 205-211.
- (S12) Yu, H. T.; Li, X. Y.; Quan, X.; Chen, S.; Zhang, Y. B. Effective Utilization of Visible Light (Including $\lambda > 600$ nm) in Phenol Degradation with p-Silicon Nanowire/TiO₂ Core/Shell Heterojunction Array Cathode. *Environ. Sci. Technol.* **2009**, *43*, 7849-7855.
- (S13) Qin, G. H.; Wu, Q. P.; Sun, Z.; Wang, Y.; Luo, J. Z.; Xue, S. Enhanced Photoelectrocatalytic Degradation of Phenols with Bifunctionalized Dye-sensitized TiO₂ film. *J. Hazard. Mater.* **2012**, *199-200*, 226-232.
- (S14) Sun, S. M.; Wang, W. Z.; Zhang, L. Efficient Contaminant Removal by Bi₂WO₆ Films with Nanoleaflike Structures through a Photoelectrocatalytic Process. *J. Phys. Chem. C* **2012**, *116*, 19413-19418.