

Supporting Information

Ammonia-mediated bromate inhibition during ozonation promotes the toxicity due to organic byproduct transformation

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Text S1 Chemicals and reagents

For chemical analysis, ultrapure water (18.2 M Ω ·cm) was prepared by a water purifier system (Milli-Q Integral). Ammonium chloride labeled with ¹⁵N (¹⁵NH₄Cl, $\geq 99\%$), ammonium chloride (NH₄Cl, $\geq 99.5\%$), potassium bromate (KBrO₃, $\geq 99.8\%$), potassium bromide (KBr, $\geq 99\%$), and potassium nitrate (KNO₃, $\geq 99\%$), were purchased from Sigma-Aldrich (St. Louis, MO, USA). All the solvents including methanol, acetone, and dichloromethane were of chromatographic grade and bought from MallinckrodtBaker Inc. (Phillipsburg, NJ, USA). Tannic acid (analytical reagent), gallic acid (99%), salicylic acid (99%), and vanillin (99%) were purchased from Aladdin (Shanghai, China). Suwannee River natural organic matter (SRNOM, 2R101N) was purchased from International Humic Substances Society.

For toxicity assays, phenol (99%), 4-nitroquinoline *N*-oxide (4-NQO, $\geq 98\%$), Hoechst 33342 ($\geq 98\%$), and paraformaldehyde (95%) were obtained from Sigma-Aldrich (St. Louis, MO, USA). Triton X-100 (reagent grade) was purchased from Solarbio (Beijing, China). Dimethyl sulfoxide (DMSO, $\geq 99.7\%$), phosphate-buffered saline (Hyclone), and a mixture of Dulbecco's Modified Eagle Medium and F-12 (DMEM/F-12 1:1 medium) were bought from Thermo Fisher Scientific (Waltham, MA, USA). Bovine serum albumin (BSA) was purchased from Amresco (Solon, OH, USA). The kit for cytotoxicity assay (Cell Counting Kit-8, CCK-8) was obtained from Dojindo (Kumamoto, Japan). Primary antibody phospho-Histone H2AX and secondary antibody Alexa Fluor 647 conjugate for genotoxicity assay were purchased from Cell Signaling Technology (Boston, MA, USA).

Text S2 Br-N-DBP analysis and calculation of their cytotoxicity index and genotoxicity index

Haloacetonitriles were analyzed by GC-ECD (7890B, Agilent Technologies, USA) equipped with a DB-5MS capillary column (30 m × 0.25 mm × 0.25 μm, Agilent, USA). Injections of 1 μL of the extracts were introduced via a splitless injector using nitrogen as the carrier gas. The injector and detector temperatures were 230 and 250 °C, respectively. The initial column temperature was 35 °C (maintained for 3 min), which was increased to 70 °C at a rate of 20 °C/min (maintained for 1 min), then 90 °C at a rate of 5 °C/min (maintained for 1 min), then 160 °C at a rate of 20 °C/min (maintained for 1 min), and finally 220 °C at a rate of 40 °C/min (maintained for 3 min).

Halonitromethanes were analyzed by GC-ECD (7890B, Agilent Technologies, USA) equipped with a DB-5MS capillary column (30 m × 0.25 mm × 0.25 μm, Agilent, USA). Injections of 1 μL of the extracts were introduced via a splitless injector using nitrogen as the carrier gas. The injector and detector temperatures were 117 and 250 °C, respectively. The initial column temperature was 35 °C (maintained for 6 min), which was increased to 190 °C at a rate of 20 °C/min (maintained for 5 min). The relatively low injector temperature was used to prevent the thermal decomposition of tribromonitromethanes ([Hong et al., 2013](#); [Plewa et al., 2004](#)).

Haloacetamides were analyzed by GC-ECD equipped with a Rtx-1701 capillary column (30 m × 0.25 mm × 0.25 μm, RESTEK, USA). Injections of 1 μL of the extracts were introduced via a splitless injector using nitrogen as the carrier gas. The injector and detector temperatures were 230 and 250 °C, respectively. The initial column

temperature was 35 °C (maintained for 3 min), which was increased to 90 °C at a rate of 20 °C/min (maintained for 1 min), then 160 °C at a rate of 20 °C/min (maintained for 1 min), and finally 220 °C at a rate of 40 °C/min (maintained for 5 min).

Contributions of these Br-N-DBPs to cytotoxicity and genotoxicity were estimated using the cytotoxicity index (CTI) and genotoxicity index (GTI), according to the method of [Chuang et al. \(2019\)](#) using previously reported LC₅₀ (for cytotoxicity) and EC₅₀ (for genotoxicity) values based on CHO cells ([Wagner and Plewa, 2017](#)). CTI and GTI were calculated using Eqs. (1) and (2). A smaller toxicity index means less contribution of the corresponding DBP to toxicity.

$$\text{CTI} = \text{DBP concentration}/\text{LC}_{50} \quad (1)$$

$$\text{GTI} = \text{DBP concentration}/\text{EC}_{50} \quad (2)$$

Text S3 Cell culture

The CHO-k1 cell line purchased from the American Type Culture Collection was used to perform the toxicity assay. The culture medium DMEM/F12 (1:1) was added with streptomycin (0.1 mg/mL), penicillin G (100 unit/mL), and fetal bovine serum (10%). Cells culture dishes were placed in an incubator with saturated humidity and 5% CO₂ gas at 37 °C. Cells undergoing 2–5 passages were used for the toxicity assay. Cell passaging was conducted every 48 h.

Text S4 Cytotoxicity assay and cytotoxicity equivalent quantification

To perform cytotoxicity assay, 1×10⁴ cells per well were seeded in a sterile 96-

well plate (3599, Corning, USA), and cultured for 12 h before the cells were exposed to organic byproducts. The dried organic extracts were firstly dissolved in DMEM/F12 containing 0.5% DMSO (v:v), then further diluted to 6-8 different concentration factors by DMEM/F12 containing 0.5% DMSO. Cell cultures were exposed to the solution for 48 h. DMEM/F12 containing 0.5% DMSO was used as the negative control, while phenol dissolved in DMEM/F12 containing 0.5% DMSO was used as the reference compound. All the 96-well plates for toxicity assay were covered with the sterilized seal films to avoid the cross contamination. After 48 h exposure, the culture medium was discarded. Cells were washed with 100 μ L phosphate buffered solution (PBS). Then 100 μ L of the reagent in Cell Counting Kit-8 (CCK-8) dissolved in DMEM/F12 was added into each well of the 96-well plates. Cells were then cultured in the incubator (37 $^{\circ}$ C) for another 2 h, then the absorbance of each well at 450 nm was measured using the microplate reader SpectraMax i3 (Molecular Devices, USA). Each test was performed in 6-12 replicates.

The cell viability (CV) for a sample or reference compound was calculated according to Eq. (1):

$$CV=(A_S-A_B)/(A_N-A_B), \quad (1)$$

where CV is the cell viability for a sample or phenol against the negative control. The A_S is the absorbance of the sample or phenol at 450 nm. A_B is the absorbance of the blank control (only added the Cell Counting Kit-8 reagent dissolved in DMEM/F12) at 450 nm. The A_n is the absorbance of the negative control at 450 nm.

The CV values at different concentration factors of each sample were used to

create a concentration–effect curve. Similarly, concentration–effect curve of phenol was obtained with the CV values at different concentrations. The concentration at which the CV value was 50% from the regression analysis was defined as the lethal concentration of 50% (LC₅₀). The cytotoxicity equivalent of a sample was calculated according to Eq. (2):

$$\text{Cytotoxicity equivalent} = \text{LC}_{\text{Phenol},50} / \text{LC}_{\text{Sample},50} \quad (2)$$

where the cytotoxicity equivalent unit is mg–phenol/L. LC_{Phenol,50} is the LC₅₀ of the reference compound phenol. LC_{sample,50} is the LC₅₀ of a sample.

Text S5 DNA DSB assay and genotoxicity equivalent quantification

To perform the genotoxicity assay, 5×10^3 cells were seeded in each well of sterile 96-well plate (3603, Corning, USA), pre-cultured for 12 h, and exposed to organic byproducts dissolved in DMEM/F12 containing 0.5% DMSO at 6-8 concentrations. DMEM/F12 containing 0.5% DMSO was used as the negative control, while 4-nitroquinoline *N*-oxide (4-NQO) dissolved in DMEM/F12 containing 0.5% DMSO was used as the reference compound. All the 96-well plates for toxicity assay were covered with the sterilized seal films to avoid the cross contamination. The 6-12 replicates of each test were performed. After exposure for 24 h, cells were fixed with paraformaldehyde, permeated with Triton-100, and blocked with bovine serum albumin. Cells were incubated with the primary antibody phospho-histone H2AX and then stained with the second antibody Alexa Fluor® 647 conjugate together with Hoechst 33258. After the staining, images of the cells were obtained using an HCA system (ImageXpress® Micro, Molecular Devices, USA) with a 40× objective lens. The

pH2AX foci were obtained from the CY5 channel and nucleus DNA was obtained from the DAPI channel. In each well of the plate, the number of pH2AX per cell was calculated by the total number of pH2AX foci over the number of nucleus.

The pH2AX induction rate of a sample was calculated according to Eq. (3):

$$IR = (\text{pH2AX foci per cell})_s / (\text{pH2AX foci per cell})_n \quad (3)$$

where IR is the pH2AX foci induction ratio for a sample or 4-NQO against the negative control. The $(\text{pH2AX foci per cell})_s$ is the number of pH2AX foci induced by a sample or 4-NQO. The $(\text{pH2AX foci per cell})_n$ is the number of pH2AX foci induced by the negative control. The pH2AX foci induction ratios of samples at different concentration factors with cell viabilities >70% (Plewa et al., 2010) were used to obtain concentration–effect curves by regression analysis. The genotoxicity equivalent of a sample was calculated according to Eq. (4):

$$\text{Genotoxicity equivalent} = \frac{(IR_{1.5})_{4-NQO}}{(IR_{1.5})_{\text{Sample}}} \quad (4)$$

where the unit of genotoxicity equivalent is $\mu\text{g-4-NQO/L}$, $(IR_{1.5})_{4-NQO}$ is the concentration of the 4-NQO solution that leads to the 1.5-fold induction ratio over the negative control, and $(IR_{1.5})_{\text{Sample}}$ is the concentration of a sample that leads to the 1.5 fold induction ratio over the negative control.

Text S6 Rough calculation of the extent of HOBr/OBr- reduced by NH_4^+ or H_2O_2

It is difficult to model the whole process of SRNOM/ O_3 /Br $^-$ / NH_4^+ or SRNOM/ O_3 /Br $^-$ / H_2O_2 because some critical information is unknown. For example, the rate constant of HOBr and SRNOM ($k_{\text{HOBr, SRNOM}}$) was not reported. Moreover, the $k_{\text{HOBr,}}$

SRNOM would vary during the ozonation process with the change of SRNOM property. Besides, the self-decomposition mechanism of NHBr_2 , the rate constant of NH_2Br and SRNOM were also unknown. However, based on reasonable assumptions we can roughly understand to what extent HOBr/OBr^- concentration would be reduced by the addition of NH_4^+ or H_2O_2 .

(1) In the $\text{SRNOM}/\text{Br}^-/\text{O}_3/\text{NH}_4^+$ system, taken the reaction condition of O_3 5 mg/L, Br^- 0.5 mg/L, NH_3 0.5 mg-N/L, NOM 5 mg-C/L as example.

The initial concentrations were $[\text{O}_3]=5$ mg/L, 1.04×10^{-4} M; $[\text{Br}^-]=0.5$ mg/L, 6.25×10^{-6} M; $[\text{NH}_3]_{\text{T}}=0.5$ mg-N/L, 3.57×10^{-5} M; $[\text{NOM}]=5$ mg-C/L, 4.17×10^{-4} M-C; $[\text{H}_2\text{O}_2]_{\text{T}}=17$ mg/L, 5×10^{-4} M. The distribution of HOBr and OBr^- could be calculated as:



HOBr would react with both O_3 , $\cdot\text{OH}$, SRNOM and NH_3 . The maximum concentration of O_3 and $\cdot\text{OH}$ were assumed and the rate of each elementary reaction was expressed in Table S2. From Table S2 (reactions 1-6), even under the assumed maximum concentration of O_3 and $\cdot\text{OH}$, reactions of HOBr/OBr^- and $\text{O}_3/\cdot\text{OH}$ (reactions 2, 3, 4 and 5) were negligible, while the reaction 1 (HOBr and SRNOM) and reaction 6 (HOBr and NH_3) would be prevailing.

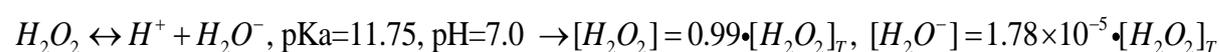
The ratio of HOBr reacting with SRNOM was estimated as $186/(186+2800)\cdot 100\%=6\%$, while the ratio of HOBr reacting with NH_3 was estimated as $2800/(186+2800)\cdot 100\%=94\%$, namely 94% HOBr would be transformed to NH_2Br .

The formed NH_2Br further reacts with O_3 , $\cdot\text{OH}$, HOBr and SRNOM. Due to the

unknown mechanism and rate constant, it was not discussed here.

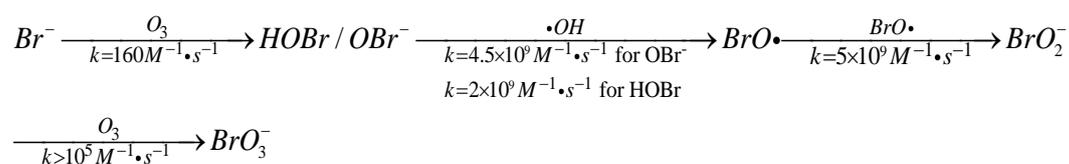
(2) In the SRNOM/Br⁻/O₃/H₂O₂ system, taken the reaction condition of O₃ 5 mg/L, Br⁻ 0.5 mg/L, H₂O₂ 17 mg /L, NOM 5 mg-C/L as example.

The initial concentrations were [O₃]=5 mg/L, 1.04×10⁻⁴ M; [Br⁻]=0.5 mg/L, 6.25×10⁻⁶ M; [NOM]=5 mg-C/L, 4.17×10⁻⁴ M; [H₂O₂]_T=17 mg/L, 5×10⁻⁴ M. The distribution of H₂O₂ and H₂O⁻ could be calculated as:

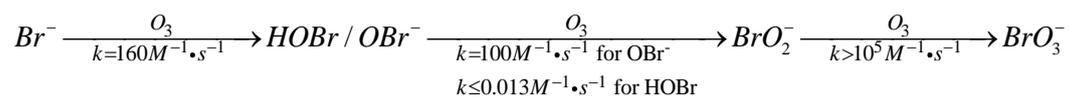


HOBr would react with both O₃, ·OH, SRNOM and H₂O₂. From [Table 1](#), among the involved reactions of 1-5, 18, 19, only reaction 1 (HOBr and SRNOM), reaction 18 (OBr⁻ and H₂O₂) and reaction 19 (HOBr and H₂O⁻) would be prevailing. The ratio of HOBr reacting with SRNOM was estimated as 186/(186+650+6.8)*100%=22%, while the ratio of HOBr reacting with H₂O₂ was estimated as (650+6.8)/(186+650+6.8)*100%=78%, namely 78% HOBr would be suppressed by H₂O₂.

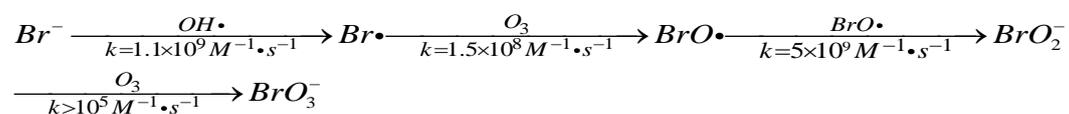
Pathway I (Direct/Indirect pathway):



Pathway II (Direct pathway):



Pathway III (Indirect/ Direct pathway):



Scheme S1 Simplified pathways for bromate formation during ozonation

(Adapted from Buxton et al. (1988), Haag and Hoigné, (1983), Hofmann and

Andrews (2006), Song et al., (1996), von Gunten and Hoigne (1994))

Table S1 Reactions and rate constants during ozonation of water containing bromide
in the presence of ammonia or hydrogen peroxide

No.	Reaction	k (25 °C)	Reference
R1	$HOBr + NH_3 \rightarrow NH_2Br$	$k_1=8 \times 10^7 \text{ M}^{-1} \cdot \text{s}^{-1}$	Haag et al., 1984
R2	$NH_2Br + OBr^- + OH^- \rightarrow N_2 + Br^- + H_2O$	$k_2=0.1 \sim 4.7 \times 10^{-4} \text{ s}^{-1}$	Hofmann and Andrews, 2001
R3	$NH_2Br + HOBr \rightarrow NHBBr_2$	$k_3=7.0 \times 10^5 \text{ M}^{-1} \cdot \text{s}^{-1}$	Haag and Lietzke, 1980
R4	$NHBBr_2 + HOBr \rightarrow NBr_3$	$k_4=2.5 \times 10^4 \text{ M}^{-1} \cdot \text{s}^{-1}$	Haag and Lietzke, 1980
R5	$NH_2Br + O_3 \rightarrow Br^- + NO_3^-$	$k_5=40 \text{ M}^{-1} \cdot \text{s}^{-1}$	von Gunten and Hoigne, 1994
R6	$NHBBr_2 + O_3 \rightarrow Br^- + NO_3^-$	$k_6=10 \text{ M}^{-1} \cdot \text{s}^{-1}$	von Gunten and Hoigne, 1994
R7	$OBr^- + H_2O_2 \rightarrow Br^- + H_2O + O_2$	$k_7=1.3 \times 10^6 \text{ M}^{-1} \cdot \text{s}^{-1}$	von Gunten and Oliveras, 1997
R8	$HOBr + HO_2^- \rightarrow Br^- + H_2O + O_2$	$k_8=7.6 \times 10^8 \text{ M}^{-1} \cdot \text{s}^{-1}$	von Gunten and Oliveras, 1997
R9	$HOBr + NOM \rightarrow \text{products}$	$k_9=4.56 \times 10^5 \text{ M}^{-1} \cdot \text{s}^{-1}$ ^a	Duirk and Valentine, 2007

Note:

^a The k value was obtained from a river water located in Iowa, the U.S.A., not SRNOM (Duirk and Valentine, 2007).

Table S2 Assumed reaction rate during ozonation of water containing bromide in the presence of ammonia or hydrogen peroxide

N o.	Reaction	k	Assumed maximum conc. (M)	Rate expression	Assumed maximum reaction rate	Referenc e
1	$HOBBr + NOM \rightarrow$ <i>products</i>	4.56×10^5 $M^{-1} \cdot s^{-1}$ a	[NOM] 4.17×10^{-4}	$k[NOM][HOB$ $r]$	$1.86 \times 10^2 \cdot$ [HOB $r]_T$	Duirk and Valentine, 2007
2	$HOBBr + O_3 \rightarrow$ $BrO_2^- + O_2 + H^+$	0.013 $M^{-1} \cdot s^{-1}$	[O ₃] 1.04×10^{-4}	$k[O_3][HOBBr]$	$1.32 \times 10^{-6} \cdot$ [HOB $r]_T$	Haag and Lietzke, 1980
3	$OBBr^- + O_3 \rightarrow$ $BrO_2^- + O_2$	100 $M^{-1} \cdot s^{-1}$	[O ₃] 1.04×10^{-4}	$k[O_3][OBBr^-]$	$2.08 \times 10^{-4} \cdot$ [HOB $r]_T$	Haag and Lietzke, 1980
4	$HOBBr + \cdot OH \rightarrow$ $BrO \cdot + H_2O$	2×10^9 $M^{-1} \cdot s^{-1}$	[·OH] 1.0×10^{-11} b	$k[\cdot OH][HOBBr]$	$1.96 \times 10^{-2} \cdot$ [HOB $r]_T$	Buxton et al., 1988
5	$OBBr^- + \cdot OH \rightarrow$ $BrO \cdot + OH^-$	4.5×10^9 $M^{-1} \cdot s^{-1}$	[·OH] 10^{-11}	$k[\cdot OH][OBBr^-]$	$9.0 \times 10^{-4} \cdot$ [HOB $r]_T$	Buxton et al., 1988
6	$HOBBr + NH_3 \rightarrow$ NH_2Br	8×10^7 $M^{-1} \cdot s^{-1}$ c	[NH ₃] 3.57×10^{-5}	$k[NH_3][HOBBr]$	$2.80 \times 10^3 \cdot$ [HOB $r]_T$	Wajon and Morris, 1982
7	$NH_3 + 4O_3 \rightarrow$ $H^+ + NO_3^- + H_2O + 4O_2$	5 $M^{-1} \cdot s^{-1}$	[NH ₃] 3.57×10^{-5} [O ₃] 1.04×10^{-4}	$k[NOM][HOB$ $r]$	3.64×10^{-9} $M \cdot s^{-1}$	Haag et al., 1984
8	$NH_2Br + O_3 \rightarrow$ $Br^- + NO_3^-$	40 $M^{-1} \cdot s^{-1}$	\	\	\	von Gunten and Hoigne, 1994
9	$NH_2Br + NOM \rightarrow$ <i>products</i>	<i>unknown</i>	\	\	\	\
10	$NH_2Br + OBr^- + OH^- \rightarrow$ $N_2 + Br^- + H_2O$	1×10^{-5} s^{-1}	\	\	\	Hofmann and Andrews, 2001
11	$NH_2Br + HOBBr \rightarrow$ $NHBr_2$	$k_3 = 7.0 \times 10^5$ $M^{-1} \cdot s^{-1}$	\	\	\	Haag and Lietzke, 1980

12	$NHBr_2 + NOM \rightarrow$ products	unknown	\	\	\	\
13	$NHBr_2$ decay, unknown	unknown	\	\	\	Hofmann, 2000
14	$NHBr_2 + O_3 \rightarrow$ $Br^- + NO_3^-$	10 $M^{-1} \cdot s^{-1}$	\	\	\	von Gunten and Hoigne, 1994
15	$NHBr_2 + HOBr \rightarrow$ NBr_3	2.5×10^4 $M^{-1} \cdot s^{-1}$	\	\	\	Haag and Lietzke, 1980
16	$2NBr_3 + 3H_2O \rightarrow$ $N_2 + 3HOBr + 3H^+ + 3Br^-$	unknown	\	\	\	Hofmann, 2000
17	$NBr_3 + NOM \rightarrow$ products	unknown	\	\	\	\
18	$OBr^- + H_2O_2 \rightarrow$ $Br^- + H_2O + O_2$	1.3×10^6 $M^{-1} \cdot s^{-1}$	$[H_2O_2]$ 5×10^{-4}	$k[H_2O_2][HOBr]$ r]	$6.5 \times 10^2 \cdot$ $[HOBr]_T$	von Gunten and Oliveras, 1997
19	$HOBr + HO_2^- \rightarrow$ $Br^- + H_2O + O_2$	7.6×10^8 $M^{-1} \cdot s^{-1}$	$[H_2O^-]$ 8.9×10^{-9}	$k[H_2O^-][HOBr]$]	$6.8 \cdot$ $[HOBr]_T$	von Gunten and Oliveras, 1997

Note:

^a The k value was obtained from a river water located in Iowa, the U.S.A., not SRNOM (Duirk and Valentine, 2007).

^b The steady state concentration of $\cdot OH$ was assumed according to the data from a SRNOM/ O_3 system (DOC 4 mg-C/L, O_3 4 mg/L, reaction time 180 s) reported previously (Wang et al., 2020), of which the condition was similar to this study. The $\cdot OH$ exposure ($\int [\cdot OH] dt$) and the steady state concentration of $\cdot OH$ ($[\cdot OH]_{ss}$) could be calculated according to equation (1) and (2):

$$\int [\cdot OH] dt = R_{\cdot OH, O_3} \cdot \Delta [O_3]_t \quad (1)$$

$$[\cdot OH]_{ss} = \int [\cdot OH] dt / t \quad (2)$$

Where $R_{\cdot OH, O_3}$ is the $\int [\cdot OH] dt$ per unit of O_3 consumed (5.10×10^{-6} s); $\Delta [O_3]_t$ is the consumed O_3 (4 mg/L, 8.33×10^{-5} M); t is the reaction time (180 s). The $[\cdot OH]_{ss}$ is therefore calculated as 2.36×10^{-5}

¹² M. Here the maximum concentration of [$\cdot\text{OH}$] was assumed as 1.0×10^{-11} M. Moreover, the assumed maximum concentration of 1.0×10^{-11} M would also be applicable in the $\text{O}_3/\text{H}_2\text{O}_2$ system (Rosenfeldt et al., 2006).

^c The k value was obtained using the concentration of total ammonia (Wajon and Morris, 1982). Therefore, the distribution of NH_4^+ and NH_3 was not considered.

Table S3 Concentrations of Br-N-DBPs formed during ozonation and their cytotoxicity index (CTI) and genotoxicity index (GTI)^a

DBPs	Cytotoxicity LC ₅₀ (M) ^b	Genotoxicity EC ₅₀ (M) ^c	Concentration (M)		CTI		GTI	
			Br ⁻	Br ⁻ and NH ₄ ⁺	Br ⁻	Br ⁻ and NH ₄ ⁺	Br ⁻	Br ⁻ and NH ₄ ⁺
MBAN	3.21× 10 ⁻⁶	3.85× 10 ⁻⁵	1.62× 10 ⁻⁹	9.58× 10 ⁻¹⁰	5.05× 10 ⁻⁴	2.98× 10 ⁻⁴	4.21× 10 ⁻⁵	4.21× 10 ⁻⁵
DBAN	2.85× 10 ⁻⁶	4.71× 10 ⁻⁵	6.28× 10 ⁻¹⁰	\	2.20× 10 ⁻⁴	\	1.33× 10 ⁻⁵	\
DBNM	6.09× 10 ⁻⁶	2.62× 10 ⁻⁵	\	9.36× 10 ⁻¹⁰	\	1.53× 10 ⁻⁴	\	7.06× 10 ⁻⁵
TBNM	8.57× 10 ⁻⁶	6.99× 10 ⁻⁵	1.11× 10 ⁻⁹	1.61× 10 ⁻⁹	1.30× 10 ⁻⁴	1.88× 10 ⁻⁴	1.59× 10 ⁻⁵	2.30× 10 ⁻⁵
MBAc Am	1.89× 10 ⁻⁶	3.68× 10 ⁻⁵	\	2.39× 10 ⁻⁹	\	1.26× 10 ⁻³	\	6.49× 10 ⁻⁵
DBAc Am	1.22× 10 ⁻⁵	7.44× 10 ⁻⁴	\	6.22× 10 ⁻¹⁰	\	5.10× 10 ⁻⁵	\	8.36× 10 ⁻⁷
TBAc Am	3.14× 10 ⁻⁶	3.25× 10 ⁻⁵	\	4.22× 10 ⁻¹⁰	\	1.34× 10 ⁻⁴	\	1.30× 10 ⁻⁵
Sum	\	\	\	\	8.55× 10 ⁻⁴	2.08× 10 ⁻³	7.13× 10 ⁻⁵	2.14× 10 ⁻⁴

^a The experiment conditions are O₃ 5 mg/L, Br⁻ 0.5 mg/L, NH₄⁺-N 0.5 mg/L.

^b LC₅₀ values are cited from [Wagner and Plewa \(2017\)](#).

^c EC₅₀ values are cited from [Wagner and Plewa \(2017\)](#).

Table S4 Cytotoxicity index (CTI) and genotoxicity index (GTI) of Br-N-DBPs at the concentration factor of the measured LC₅₀ or IR_{1.5} in this study

Experimental conditions	O ₃ 5 mg/L, Br ⁻ 0.5 mg/L		O ₃ 5 mg/L, Br ⁻ 0.5 mg/L, NH ₄ ⁺ -N 0.5 mg/L	
	Cytotoxicity	Genotoxicity	Cytotoxicity	Genotoxicity
Measured LC ₅₀ or IR _{1.5} in this study (folds)	297	134	189	78
CTI or GTI at the concentration factor of 1	8.55×10 ⁻⁴	7.13×10 ⁻⁵	2.08×10 ⁻³	2.14×10 ⁻⁴
CTI or GTI at the concentration factor of the measured LC ₅₀ or IR _{1.5}	0.25	0.010	0.39	0.016

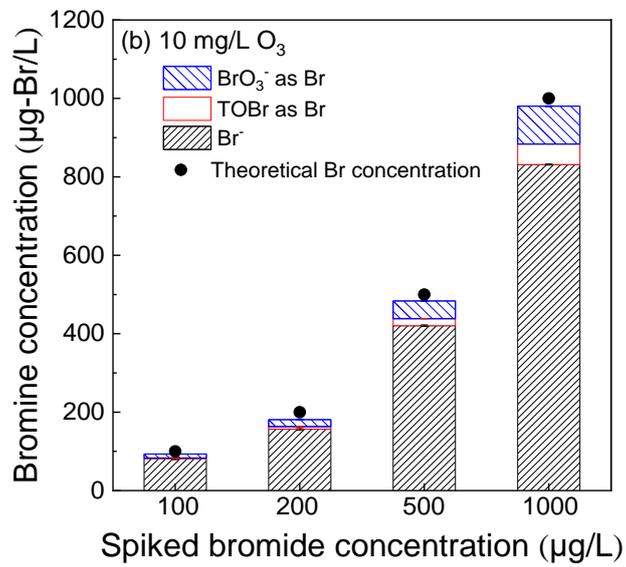
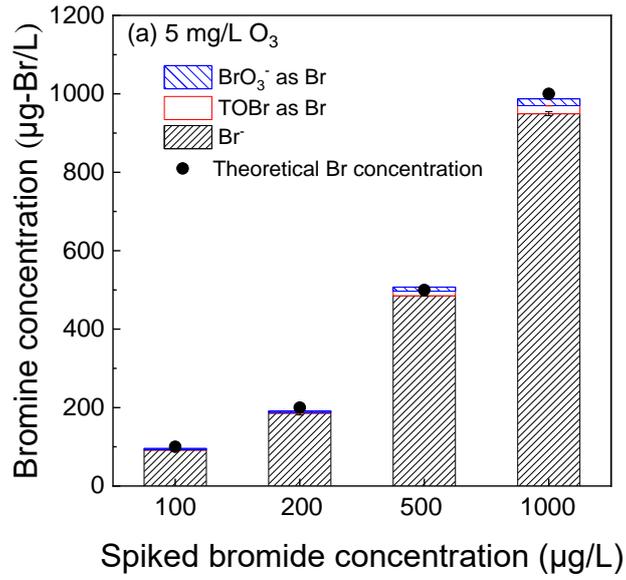


Figure S1 Mass balance of bromine during ozonation: **(a)** Ozone dose 5 mg/L;
(b) Ozone dose 10 mg/L

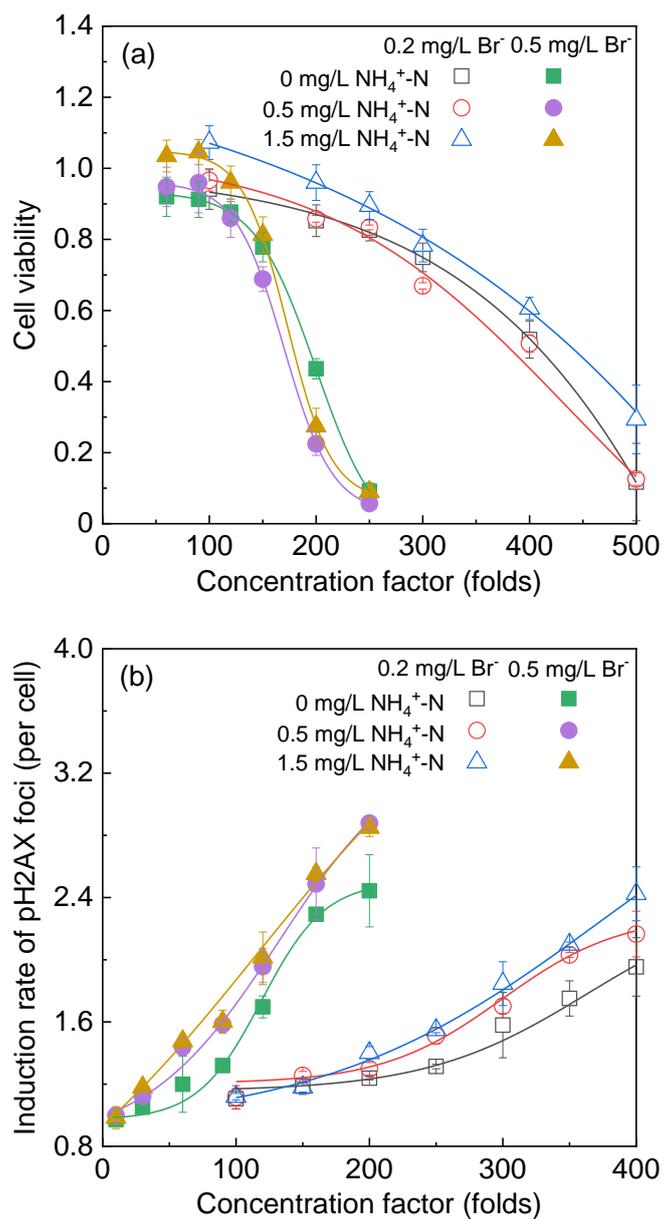


Figure S2 Influence of ammonia on the toxicity of organic byproducts formed during ozonation in the presence of bromide: **(a)** Concentration-effect curves in the cytotoxicity assay (ozone dose 10 mg/L); **(b)** Concentration-effect curves in the genotoxicity assay (ozone dose 10 mg/L)

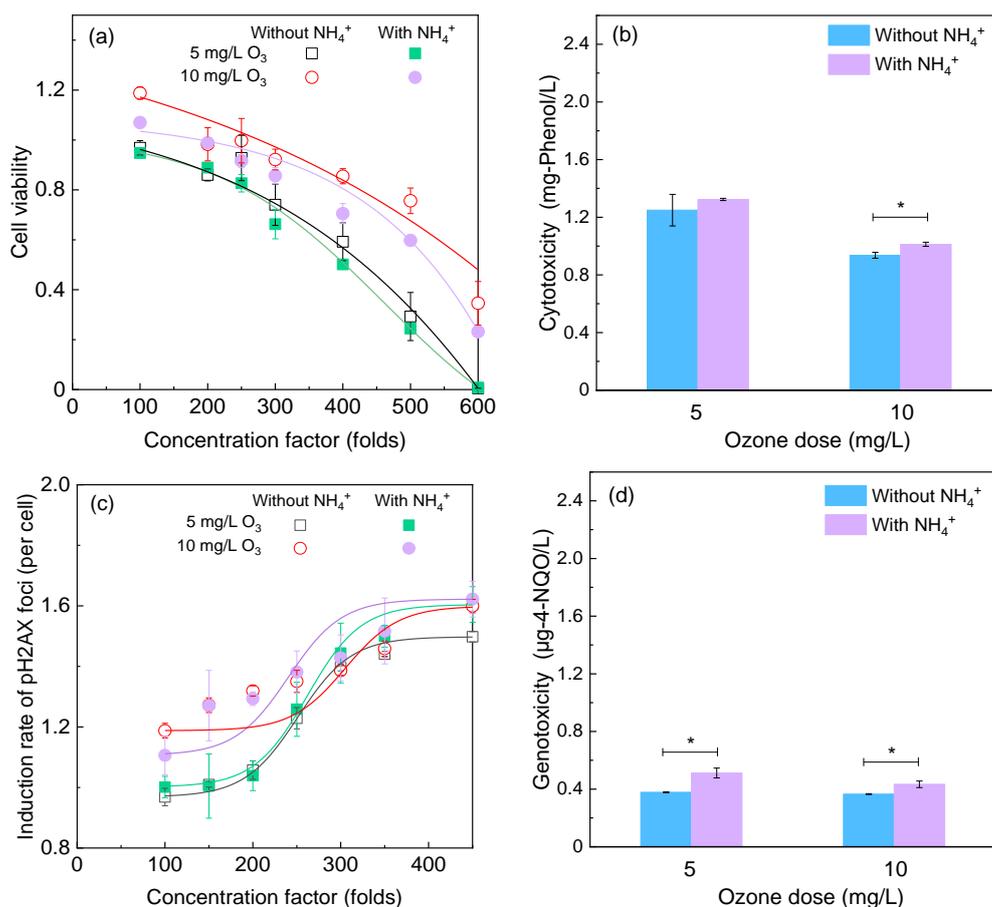


Figure S3 Influence of ammonia on the toxicity during ozonation in the absence of bromide ($\text{NH}_4^+\text{-N}$ 1.5 mg/L): **(a)** Concentration-effect curves in the cytotoxicity assay; **(b)** Cytotoxicity equivalents (* indicates significantly increased cytotoxicity from the addition of ammonia compared to that without adding ammonia; one-way ANOVA, $p < 0.05$); **(c)** Concentration-effect curves in the genotoxicity assay; **(d)** Genotoxicity equivalents (* indicates significantly increased genotoxicity from the addition of ammonia compared to that without adding ammonia; one-way ANOVA, $p < 0.05$).

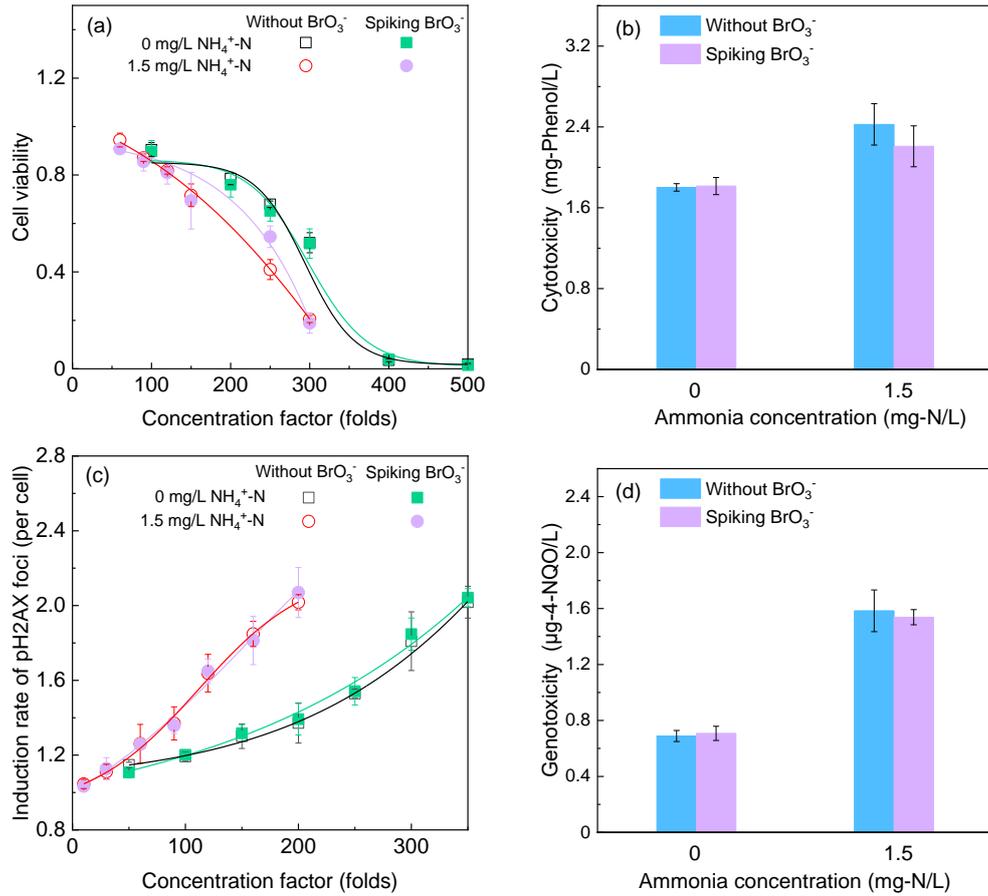


Figure S4 Influence of the formed bromate on the toxicity of organic byproducts. (5 mg/L O_3 , 0.5 mg/L Br. The formed bromate was measured and then spiked into the organic extracts by the same concentration factor as organic extracts. The formed bromate concentration were 16 $\mu\text{g/L}$ (not adding ammonia) and 12 $\mu\text{g/L}$ (adding 1.5 mg/L $\text{NH}_4^+\text{-N}$): **(a)** Concentration-effect curves in the cytotoxicity assay; **(b)** Cytotoxicity equivalents; **(c)** Concentration-effect curves in the genotoxicity assay; **(d)** Genotoxicity equivalents.

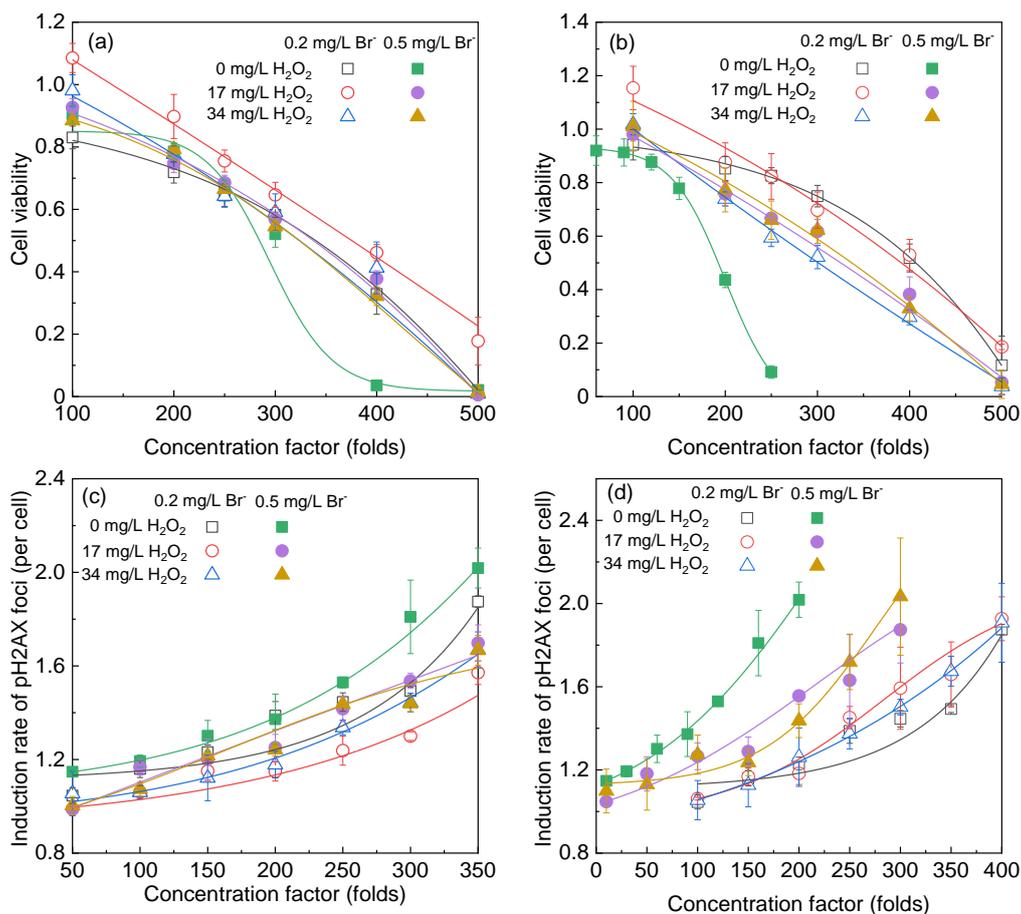


Figure S5 Influence of hydrogen peroxide (H_2O_2) on the toxicity of organic byproducts formed during ozonation in the presence of bromide: **(a)** Concentration-effect curves in the cytotoxicity assay (O_3 5 mg/L); **(b)** Concentration-effect curves in the cytotoxicity assay (O_3 10 mg/L); **(c)** Concentration-effect curves in the genotoxicity assay (O_3 5 mg/L); **(d)** Concentration-effect curves in the genotoxicity assay (O_3 10 mg/L)

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