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## MAERIALS AND METHODS

### *Triclosan adsorption and P25 photocatalysis experiments*

Adsorption and photocatalysis experiments were conducted in the dark and visible light conditions respectively, at room temperature ( $23 \pm 1$  °C) by mixing 5 mg/L P25 with different concentrations of triclosan ranging from 312.5 to 5 000 µg/L. The solution concentrations were determined at three time intervals including 12h, 24, and 72 h. Hydrodynamic diameter and zeta potential of the P25 in aqueous phase in the presence of triclosan were measured by a Zetasizer Nano ZS (Malvern, United Kindom). For the quantification of residual triclosan, the aqueous phase was separated using centrifugation at 16128g for 20 min. The triclosan in supernatant was analyzed using Agilent 1260 liquid chromatograph equipped with a diode array detector (Santa Clara, CA, USA). A ZORBAZ XDB-C18 column ( $250 \times 4.6$  mm, 5 µm, Agilent) was used for the analysis, with the oven temperature of 40 °C. The injection volume was 50 µL. The mobile phase consisting of 70:30 v:v% acetonitrile : water with flow rate of 0.8 mL/min. The wavelength used for detection was 214 nm. The material in the pellet was used to determine the adsorption efficiency of triclosan by P25 through detecting the functional groups of triclosan on the surface of P25 using a FT-IR Bruker Tensor 27 spectroscopy (Bruker, Etlinger, Germany) equipped with an attenuated total reflectance.

### *Determination of algal dry weight and Chlorophyll a/b concentrations*

50 Biomass dry weight was obtained by filtering a known volume of algal cells on a pre-  
51 weighed 0.45  $\mu\text{m}$  membrane filter. Filters with algal cells were dried for 24 hours at  
52 60 °C and weighed to determine cell mass per volume of culture. The dry weight of  
53 cells was obtained by the difference between weight before and after drying.

54

55 Chlorophyll a (Chl a) and chlorophyll b (Chl b) were measured according to a  
56 modified method.<sup>1</sup> Briefly, the algae were exposed to target components for 120 h.  
57 8 mL of algal suspension was centrifuged at 2268g for 15 min, after which the  
58 supernatant was discarded. The pellet was resuspended in 8 mL 90% ethanol with 1%  
59  $\text{MgCO}_3$ , incubated at 75 °C for 5 min in the dark. 1%  $\text{MgCO}_3$  was added to prevent  
60 the conversion of chlorophylls to phaeophytin.<sup>2</sup> The extract was centrifuged at 2268g  
61 for 15 min, and the pellet was discarded. The absorbance of chlorophyll in the  
62 supernatant of the extract at 645 and 663 nm were analyzed by a Cary-300 double  
63 beam UV-visible spectrophotometer (Agilent Technologies, CA, USA). The content  
64 of Chl a and Chl b were calculated according to Jeffrey and Humphrey equation.<sup>3</sup>

65

#### 66 *Cell morphology and ultrastructure observations*

67

68 Surface morphology of cells grown for 120h was investigated using scanning electron  
69 microscopy (SEM). Samples were centrifuged at 10,000g for 15 min, after which the  
70 supernatants were removed. The pellets were fixed in 2% glutaraldehyde in 50mM  
71 sodium cacodylate buffer (NaCac), pH 7.2 at room temperature for three hours, and  
72 then stored overnight in a refrigerator. Samples were subsequently washed in 50mM  
73 NaCac three times and further fixed in 1% Osmium tetroxide in 50mM NaCac for 60  
74 min. Samples were then washed in 50mM NaCac three times and dehydrated in an

75 ethanol gradient. After drying through a critical point drying apparatus (Polaron  
76 E3000, Canada), samples were coated with gold in a sputter coater (Quorum Q150T  
77 ES, Canada). SEM images were obtained using a FE-SEM-cold field emission  
78 scanning electron microscope (Hitachi SU8010, Japan).  
79  
80 Transmission electron microscopy (TEM) was used to investigate cellular  
81 ultrastructure. Algal cells grown for 120 h were thoroughly washed using deionized  
82 water, prefixed in 2% glutaraldehyde, postfixed in 1% osmium tetroxide for 1 h, and  
83 dehydrated through graded ethanol series. Samples were infiltrated with LR white  
84 resin 1:1 100% EtOH for 3 hours, and 100% LR white resin overnight. Samples were  
85 blocked into BEEM capsule and polymerized for 24 hours at 60 degrees. Ultrathin  
86 sections (90nm) were cut using an ultramicrotome (Leica, Canada). TEM images  
87 were obtained on a transmission electron microscopy (Hitachi HT7700, Japan).

88

89 *Elemental distribution for single cells through Synchrotron-based X-ray*  
90 *fluorescence imaging*

91

92 For the present XFI measurement, single algal cells were scanned using a  
93 polychromatic (pink) beam with a spot size of  $3 \times 3 \mu\text{m}$ , a step size of  $5 \mu\text{m}$  and a  
94 dwell time of 5 s per data point. The cell deposited on the metal-free Lexan<sup>TM</sup>  
95 polycarbonate sheets were mounted on a sample holder and attached to a motorized  
96 stage. Synchrotron X-ray beam (pink) was incident at an angle of  $90^\circ$  to the cell,  
97 while a Vortex<sup>®</sup> silicon drift detector (SDD), positioned 50 mm away from the cell at  
98 an angle of  $45^\circ$  to the incident beam, was used for simultaneous collection of the  
99 emitted X-ray fluorescence signals of tested elements. The map size of approximately

100 160  $\mu\text{m}$   $\times$  160  $\mu\text{m}$  was selected for the elemental mapping of all algae samples. Three  
101 replicates of algae cells from each treatment were used for XFI to ensure  
102 reproducibility of the collected data.

103

#### 104 *Oxidative stress*

105

106 Intracellular ROS was measured using 2,7-dichlorodihydrofluorescein diacetate  
107 (DCFH-DA, BioVision Incorporated, USA), which is an oxidation-sensitive  
108 fluorescent probe dye.<sup>4</sup> Briefly, algal cells were grown for 120 h in the presence of 0-  
109 4 mg/L triclosan with or without 5 mg/L P25, harvested (12,000  $\times$  g, 10 min), washed  
110 and suspended in phosphate-buffered saline (PBS) containing 10 $\mu\text{M}$  DCFH-DA in  
111 dimethyl sulfoxide (DMSO), and then incubated for 30 min in the dark at 30 °C. Algal  
112 cells were harvested, washed and lysed by sonication on ice for 2min in 30 s/min  
113 on/off cycles in the dark. 200  $\mu\text{L}$  of the supernatant was transferred to a 96 well plate  
114 and the fluorescence measured using a BioTek microplate reader (Winooski, VT, US;  
115  $\lambda_{\text{ex}}$  = 485 nm;  $\lambda_{\text{em}}$  = 528 nm). The assay was repeated a minimum of three times.

116

117 The mitochondrial dysfunction of exposed algal cells was measured by alteration of  
118 mitochondrial membrane potential (MMP,  $\Delta\psi\text{m}$ ). Briefly, cells were collected by  
119 centrifugation (12,000  $\times$  g, 10 min) and washed by assay buffer. MMP was measured  
120 by the incorporation of a cationic fluorescent dye, tetramethylrhodamine, ethyl ester  
121 (TMRE) dye. Aliquots of approximately  $1 \times 10^5$  -  $5 \times 10^5$  cells/mL were stained with  
122 TMRE dye to a final concentration of 200 nM at 37 °C for 30 min. The dyed algal  
123 cells were washed three times by assay buffer to eliminate the excess dye. After

124 washing, 100 $\mu$ L assay buffer was added and the fluorescence was measured using a  
125 BioTek microplate reader (Winooski, VT, US;  $\lambda_{ex}$  = 549 nm;  $\lambda_{em}$  = 575 nm).  
126  
127 The decomposition rate of hydrogen peroxide by catalase (CAT) was used to assay  
128 the enzyme activity according to manufacturer's instruction (EnzyChrom Catalase  
129 Assay Kit, BioAssay Systems Co, CA, USA). Briefly, cells were are homogenized in  
130 cold PBS, and centrifuged for 10 min at 12, 000g. A reaction mixture of 100  $\mu$ L  
131 containing 90  $\mu$ L of 50  $\mu$ M H<sub>2</sub>O<sub>2</sub>, and 10  $\mu$ L algae supernatant was reacted for 30 min  
132 at room temperature. 100  $\mu$ L detection reagent containing horse radish peroxidase  
133 enzyme and dye reagent was added and incubated for 10 min. a mixture of 100  $\mu$ L  
134 was transferred to a 96 well plate and the fluorescence was measured using a BioTek  
135 microplate reader (Winooski, VT, US;  $\lambda_{ex}$  = 530 nm;  $\lambda_{em}$  = 585 nm). Catalase  
136 activity was normalized to the unit amount of total protein to compare results between  
137 different treatments.

138

139

## 140 RESULTS AND DISCUSSIONS

141

### 142 *Verification of triclosan adsorption and P25 photocatalysis*

143

144 P25 is a mixture with the crystallographic forms of 78 % anatase and 14 % rutile.<sup>5</sup>

145 TEM micrographs showed that P25 was spherical with diameters of 15-25 nm (Figure

146 S3). The mean hydrodynamic diameter of P25 after 48 h was determined to be 417.6

147  $\pm 22.23$  nm when distributed in LEW and aggregation was observed. The zeta

148 potential of P25 was  $-9.29 \pm 0.54$  mV. The pH of LEW was 6.8, higher than 6.3 ( $p_{zc}$  of

149 P25), which led to higher degree of deprotonation of surface hydroxyl groups, causing

150 negatively charged surface of P25. The relatively low negative zeta potential may lead

151 to the disequilibrium of chemicals in the surrounding water environment and result in

152 charge heterogeneity, which was favorable for particle aggregation.<sup>6</sup>

153

154 In dark adsorption experiments, the average size of aggregated P25 showed an

155 increasing trend with increasing concentrations of triclosan at 48h (Figure S4). As

156 well, there was a decreasing trend of zeta potential with increasing triclosan

157 concentrations (Figure S5). Lower zeta potential corresponded to larger

158 hydrodynamic diameter, which was favorable for the formation of aggregation. Thus,

159 higher concentrations of triclosan resulted in the aggregation of P25 into larger

160 aggregates. Since larger particle aggregates may be more difficult to enter the cells, it

161 is hypothesized that larger P25 aggregates with higher levels of triclosan might not

162 cause enhanced toxicity compared to that with triclosan-only exposed cells. But that

163 one more thing should be considered is molecular concentration of triclosan also

164 played a key role in toxicity. Because non-ionized triclosan is known to be more toxic

165 than its ionized form,<sup>7</sup> the toxicity would be more serious when non-ionized triclosan  
166 is more than ionized triclosan. Since triclosan has a pKa of 8.1,<sup>8</sup> most triclosan was in  
167 its molecular form at pH 6.8 in LEW. Thus, triclosan concentration seriously matters  
168 using LEW in our study.

169

170 Through adsorption studies in dark, the results showed triclosan adsorption onto P25  
171 occurred from 0 to 24h, and it reached equilibrium after 24h. However, the adsorption  
172 of triclosan on P25 was so mild and the adsorption rate was less than 6% (data not  
173 shown). In Figure S6, it does not show any difference in terms of band intensities and  
174 frequencies when comparing ATR-FTIR spectra of P25 in the absence and in the  
175 presence of triclosan at both 5000 µg/L and 2500 µg/L. It might be because the  
176 detection limit of ATR-FTIR was not enough to verify the adsorption of triclosan on  
177 P25 surface. The mild adsorption occurred; but it was too weak to be observed by  
178 ATR-FTIR.

179

180 Through triclosan photodegradation study (Figure S7), it can be seen that triclosan  
181 degradation under illumination was a slow process no matter it was due to photolysis  
182 (in the absence of P25) or photocatalysis (in the presence of P25) in visible light.  
183 Triclosan degradation rate in the presence of P25 was from 4 to 7 %, higher than that  
184 without P25 in 120h. The highest degradation rate was 6.47% in the presence of P25  
185 in 120h. However, triclosan degradation rate in the presence of P25 was lower than  
186 that without P25 in 60h. It suggested that triclosan can be phototransformed by P25 in  
187 visible light, but with a low degradation rate. Normally, there are two principal  
188 catalytic phases of P25, anatase and rutile. Anatase P25 has a band gap of 3.2 eV (385  
189 nm), and can not be active in visible light.<sup>9</sup> Rutile has a smaller band gap of 3.0 eV



190 with excitation wavelengths extending into visible wavelength at 410 nm.<sup>9</sup> Thus, P25  
191 can extend the photoactivity into visible wavelengths due to the existence of 14%  
192 rutile. It is also the reason that triclosan photodegradation could be enhanced in the  
193 presence of P25. However, only 14% rutile contributed to triclosan degradation, with  
194 hydroxyl radicals as the main species of ROS.<sup>10</sup> Therefore, the degradation rate of  
195 triclosan in visible light is low.

196

### 197 ***Dry weight analysis***

198

199 The size of *Eremosphaera* is too large to be counted by a hemocytometer. In order to  
200 estimate the size of the cultured algal population, dry weight measurement was  
201 performed. Figure S8 shows that dry weight exhibited dose-response relationship and  
202 had significant decreases in cells exposed to 4000 and 1000 µg/L triclosan. In the  
203 presence of P25, there was a significant decrease in cells exposed to 250 µg/L  
204 triclosan, while there were no significant changes of cells exposed to the same dose of  
205 triclosan in the absence of P25. Thus, the interaction can cause a significant reduction  
206 in dry weight at 250 µg/L triclosan, while pure triclosan can not cause a significant  
207 reduction until reaching 1000 µg/L. The dry weight in the presence of P25 was higher  
208 than that in the absence of P25. Hartmann's study<sup>6</sup> had the similar results that the  
209 growth of *Pseudokirchneriella subcapitata* was stimulated by 2 mg/L P25. This effect  
210 is considered to be the hormesis response, which is a stimulatory effect caused by  
211 short-term exposure to a low dose of a toxicant.<sup>11</sup> Interestingly, the dry weight of cells  
212 co-exposed to 15.625 µg/L triclosan and P25 was significantly higher than that of  
213 unexposed cells and cells only exposed to P25 and triclosan. It showed a stimulation  
214 of cell growth under such an interaction, with larger cell density than any individual

215 exposure. Since the absolute value of dry weight exposed to the combinations was  
216 higher than that exposed to pure triclosan, P25 may alleviate the toxic effects of  
217 triclosan on algal biomass.

218

#### 219 ***Chlorophyll a/b analysis***

220

221 The chlorophyll content is an indicator of photosynthetic ability. In Figure S9, in the  
222 absence of P25, there were significant decreases in chlorophyll a content to cells  
223 exposed to 1000 and 4000 µg/L triclosan, and significant decreases in chlorophyll b  
224 content to cells exposed to 4000 µg/L triclosan. Our previous study presented an  
225 inhibition in chlorophyll content of *Chlorococcum sp.* with increasing triclosan  
226 concentrations.<sup>12</sup> Pan<sup>13</sup> also reported a significant decrease in chlorophyll content of  
227 *Chlamydomonas reinhardtii* at 405.3 µg/L triclosan due to the damages in  
228 photopigments. Moreover, chlorophyll a/b content of cells only exposed to 5 mg/L  
229 P25 was much higher than that of unexposed cells. Similarly, in Middepogu's study,<sup>14</sup>  
230 chlorophyll a content of *Chlorella pyrenoidosa* after 96 h exposure to 20 mg/L nano-  
231 TiO<sub>2</sub> was significant higher than that in no-exposed cells, and chlorophyll a content  
232 also increased from 20 to 45 nmol/10<sup>5</sup> cells with increasing nano-TiO<sub>2</sub> concentration  
233 from 0.1 to 10 mg/L.

234

235 In the presence of P25, there were significant decreases in chlorophyll a/b contents of  
236 cells exposed to triclosan in a range from 250 to 4000 µg/L, while a significant  
237 increase was observed for cells exposed to 15.625 µg/L triclosan. Our previous study  
238 also found that there was a significant stimulation of chlorophyll contents in  
239 *Asterococcus superbus* and *Eremosphaera viridis*, when exposed to much lower

240 triclosan concentrations of 0.56 or 0.0862 µg/L.<sup>15</sup> Such a significant stimulation in  
241 chlorophyll a/b is probably attributed to the hormesis caused by the interaction of low  
242 levels of triclosan and P25. The hormesis is the consequence of regulatory  
243 overcorrections by biosynthetic control mechanisms to low levels of inhibiting  
244 challenge, causing greater index than normal.<sup>16</sup>

245

#### 246 ***Variation of triclosan concentrations***

247

248 The concentration of triclosan was measured during the exposure period and the  
249 results are given in Figure S11. pH value was stable at 6.8 before and after exposure  
250 experiment. There was a trend that triclosan decreased rapidly at the first 3 days, and  
251 then decreased gradually at the rest 2 days in the absence and presence of P25. The  
252 final average concentrations of triclosan for all treatments were in the range of 80 –  
253 90% of initial concentrations, suggesting that triclosan was relatively stable during the  
254 exposure period.

255

256 The ratios of triclosan concentration variation in the presence of P25 to that in the  
257 absence of P25 in nonaqueous phase were shown in Table S2. On day 0, the ratio was  
258 between 1 and 2, indicating the adsorption of triclosan in the presence of P25 was  
259 higher than that in the absence of P25. Triclosan reduction in the presence of P25 was  
260 attributed to the adsorption of both cells and P25 particles. On day 5, the ratio all  
261 decreased, indicating triclosan decline in the presence of P25 became lower than that  
262 in the absence of P25. At 4000 µg/L triclosan, the difference between two-day's ratios  
263 was the smallest, whereas at 15.625 µg/L triclosan, the difference was the largest.

264 When triclosan concentration decreased, the growth became more stimulated, and led  
265 to the enlargement of cell population, increasing triclosan reduction.

266

### 267 ***Responses of mitochondrion membrane potential***

268

269 Mitochondria is important not only in bioenergetics and metabolism, but also in  
270 cellular processes to environmental stressors.<sup>17</sup> Mitochondrial membrane potential is  
271 often used for assessing mitochondrial function, as it relates to cells' capability to  
272 generate ATO by oxidative phosphorylation.<sup>18</sup> Thus,  $\Delta\psi_m$  is a key indicator of cell  
273 health or injury.

274

275 Figure S13 shows the changes on mitochondrial membrane potential of treated cells.

276 The absolute value of the mitochondrial depolarization in co-occurrence of P25 and  
277 triclosan was higher than that of only-triclosan-exposed treatments (data not shown).

278 Triclosan at 15.625 and 62.5  $\mu\text{g/L}$  with or without P25 had provoked mitochondrial  
279 depolarization but not significant. Thus, both P25 and low level triclosan with or  
280 without P25 could cause hormetic response in membrane potential depolarization.

281

282 At 250  $\mu\text{g/L}$  triclosan, mitochondrial membrane potential decreased with or without  
283 P25, indicating deleterious effects on mitochondrion. But co-exposed cells did not  
284 have such significant decrease as triclosan-only exposed cells in this situation. At

285 1000  $\mu\text{g/L}$  triclosan, co-exposed cells had a significant lower mitochondrial

286 membrane potential than triclosan-only exposed cells, indicating the interactive

287 effects beyond a certain amount of triclosan enhanced the inhibition in mitochondrial

288 function. It has been proved that triclosan may impair the function of mitochondria,

289 through uncoupling oxidative phosphorylation and disrupting mitochondrial  
290 membrane fluidity.<sup>19, 20</sup> However, P25 alleviated damages on mitochondrium at 250  
291  $\mu\text{g/L}$  triclosan and intensified damages at 1000  $\mu\text{g/L}$  triclosan. Thus, the interaction of  
292 P25 and a specific lower dose of triclosan led to antagonism in mitochondrial  
293 membrane potential. Once the cell membrane was damaged with elevating triclosan  
294 dose, P25 may enter the cells and attach to mitochondria, resulting in protein  
295 denaturation and cytoplasm flow-out.<sup>21</sup> Hence, the interaction of P25 and a specific  
296 higher dose of triclosan caused synergetic damage on mitochondrial function.

297

298 At 4000  $\mu\text{g/L}$  triclosan, there was significant lower mitochondrial membrane potential  
299 for cells in the presence and in the absence of P25, but there was no significant  
300 difference between those two situations. It indicated pure triclosan at such  
301 concentration caused serious cell apoptosis so that P25 had little additional effects on  
302 cells.

303

#### 304 ***Responses of catalase activity***

305

306 An increase in ROS production will eventually improve the production of antioxidant  
307 enzymes present inside the cells, such as SOD (acts on  $\text{O}_2^-$ ), catalase (on  $\text{H}_2\text{O}_2$ ), and  
308 GST (on the conjugation of glutathione).<sup>22</sup>  $\text{H}_2\text{O}_2$  is generated through the catalysis of  
309  $\text{O}_2^-$  during the photo respiration and  $\beta$ -oxidation of fatty acids.<sup>23</sup> Catalase is a  
310 manganese or heme-containing enzyme, catalyzing the decomposition of  $\text{H}_2\text{O}_2$  to  
311 water and oxygen.<sup>24</sup> Catalase plays an important role in the detoxification of active  
312 oxygen species generated by various types of environmental stress.<sup>23</sup> Catalase activity

313 has been frequently described in the mitochondria, and also been detected in  
314 chloroplasts.<sup>25</sup>

315

316 Figure S14 presents the catalase content of triclosan-exposed cells in the absence and  
317 presence of P25. Significant catalase changes in *Eremosphaera* were observed when  
318 cells were exposed to triclosan at 15.625 µg/L and beyond 1000 µg/L. No matter with  
319 or without P25, cells had a significant increase in catalase at 15.625 µg/L triclosan,  
320 but a significant decrease at 4000 µg/L triclosan. At 15.625 µg/L triclosan in the  
321 absence and presence of P25, catalase content had been stimulated by 133 and 155%,  
322 respectively, compared to its corresponding control. At 4000 µg/L triclosan in the  
323 absence and presence of P25, catalase content had been inhibited by 20 and 15 %,  
324 respectively, compared to its corresponding control. What's more, there was no  
325 significant difference between P25-only exposed cells and untreated cells, indicating  
326 P25 did not induce excess catalase. However, the distinctions between cells in the  
327 absence and presence of P25 at 250 and 1000 µg/L triclosan were significant. Thus,  
328 although 5 mg/L P25 did not cause significant changes in catalase content, its  
329 interaction with triclosan did. The main reason was attributed to triclosan and its  
330 concentrations.

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338 **List of figures and tables**

339 **Figure S1.** The microscopic image of freshwater green alga *Eremosphaera*.

340 **Figure S2.** (A) The Canadian light source, (B) Bruker Vertex 70v Interferometer /

341 Hyperion 3000 IR Microscope on Mid-IR beamline, (C) Workstation on VESPERS  
342 beamline.

343 **Figure S3.** TEM micrograph of P25.

344 **Figure S4.** Hydrodynamic diameter for 5 mg/L P25 in the presence of triclosan (n=3).

345 **Figure S5.** Zeta potential for 5 mg/L P25 in the presence of triclosan (n=3).

346 **Figure S6.** ATR-FTIR spectra of pure triclosan, P25, P25 – 5000 µg/L triclosan, and  
347 P25 – 2500 µg/L triclosan.

348 **Figure S7.** Degradation rate of triclosan in the absence and presence of 5 mg/L P25 in  
349 LEW.

350 **Figure S8.** Dry weight of *Eremosphaera* exposed to triclosan in the absence and  
351 presence of 5 mg/L P25 (n=3).

352 **Figure S9.** The variation of chlorophyll content in cultures exposed to triclosan in the  
353 absence and presence of 5 mg/L P25 (n=3). (A) chlorophyll a, (B) chlorophyll b.

354 **Figure S10.** TEM images of *Eremosphaera*. (A) (B) Cells exposed to 1000 µg/L  
355 triclosan; (C) (D) Cells co-exposed to 5 mg/L P25 and 1000 µg/L triclosan.

356 **Figure S11.** Observed triclosan concentration in algal media.

357 **Figure S12.** Partial investigation of the distribution of multi-elements in an individual  
358 algal cell. (A) Cells exposed to 250 µg/L triclosan; (B) Cells exposed to 5 mg/L P25;  
359 (C) Cells co-exposed to 5 mg/L P25 and 62.5 µg/L triclosan; (D) Cells co-exposed to  
360 5 mg/L P25 and 250 µg/L triclosan.

361 **Figure S13.** Mitochondrial depolarization of algal cells exposed to triclosan in the  
362 absence and presence of 5 mg/L P25 (n=3).

363 **Figure S14.** Catalase of algal cells exposed to triclosan in the absence and presence of  
364 5 mg/L P25 (n=3).

365 **Table S1.** Triclosan given with its molecular information, water solubility and  
366 application area.

367 **Table S2.** The ratios of triclosan concentration variation in the presence of P25 to that  
368 in the absence of P25 in nonaqueous phase.

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370



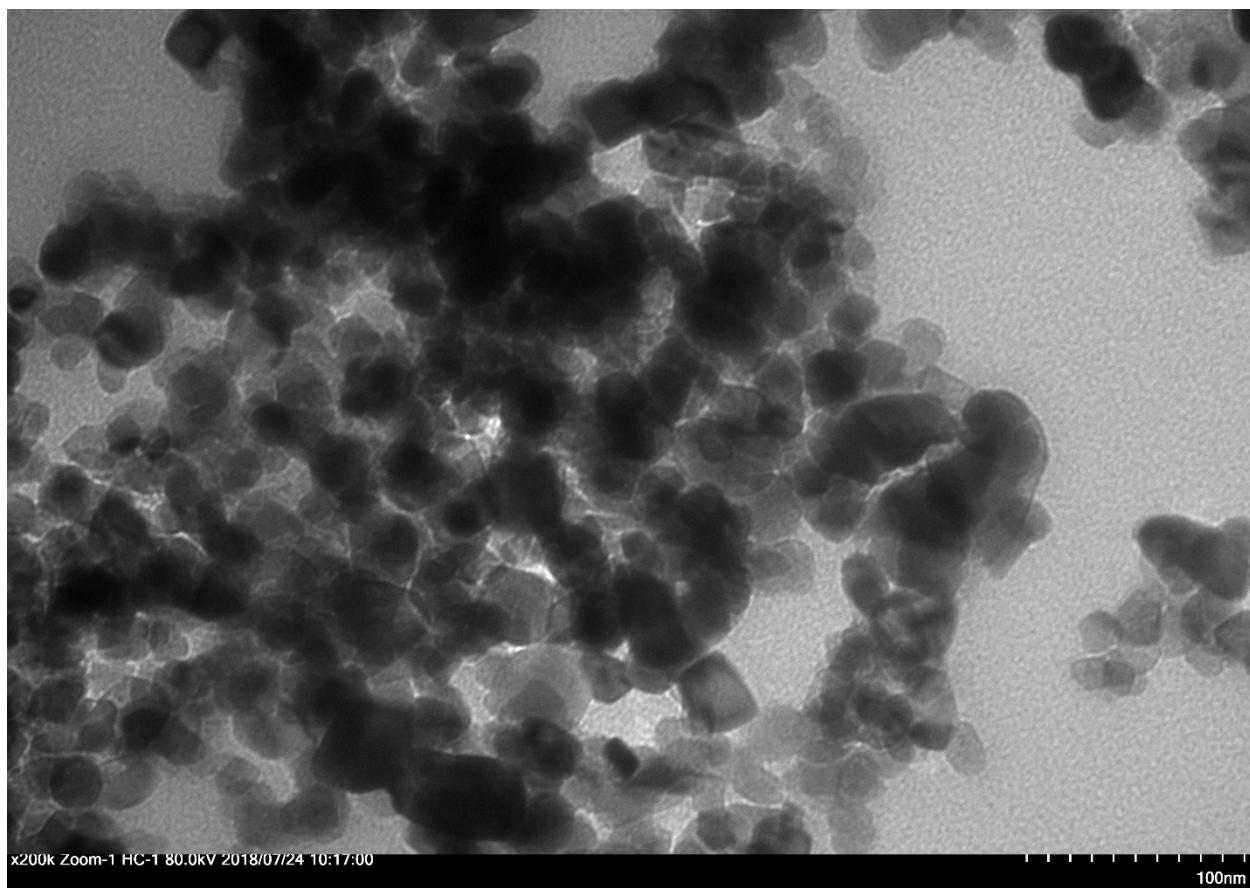


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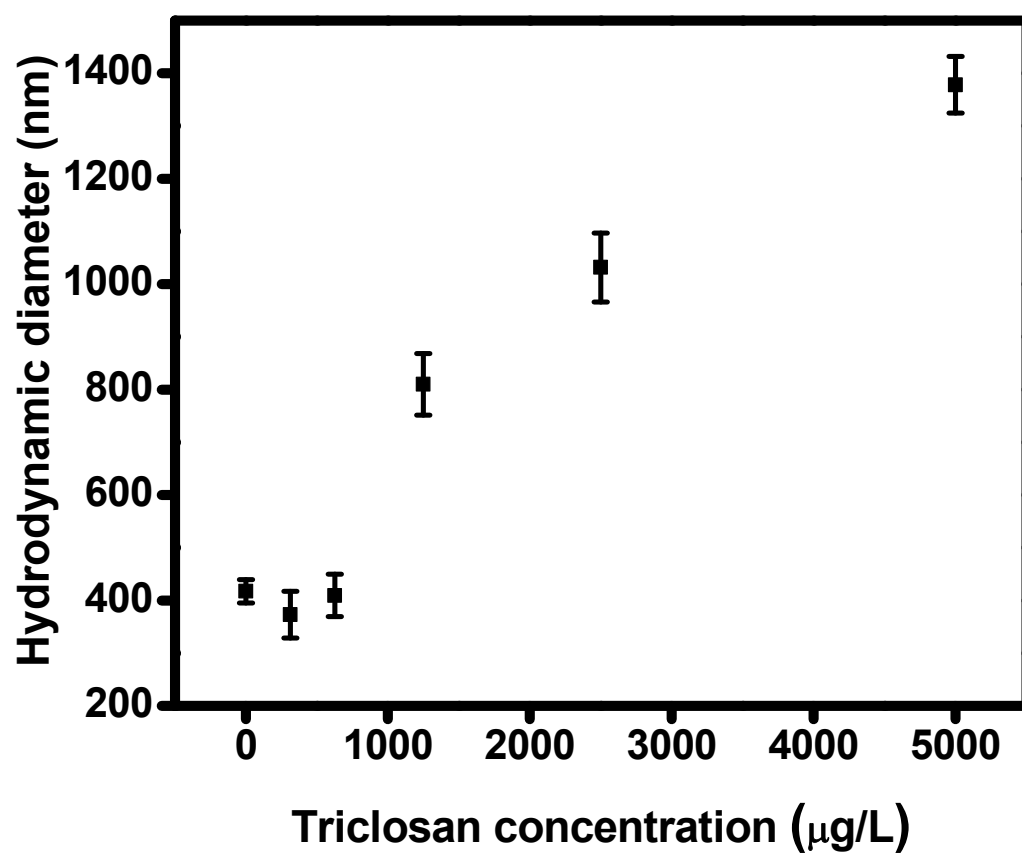
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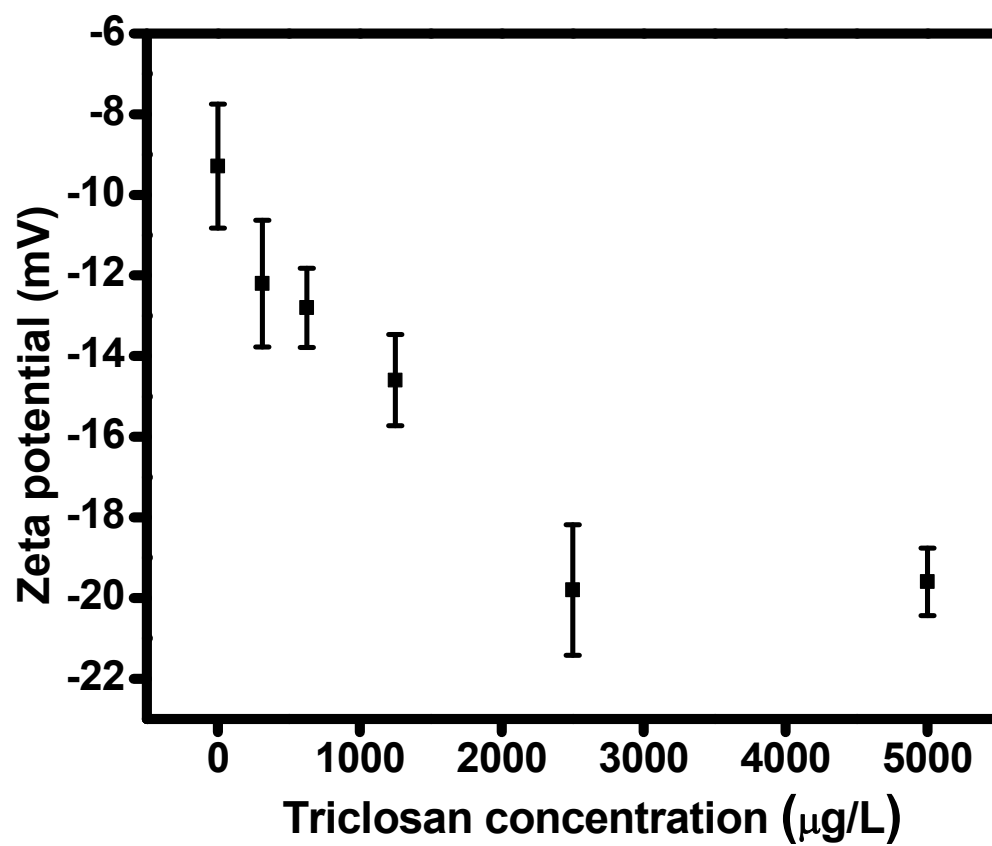
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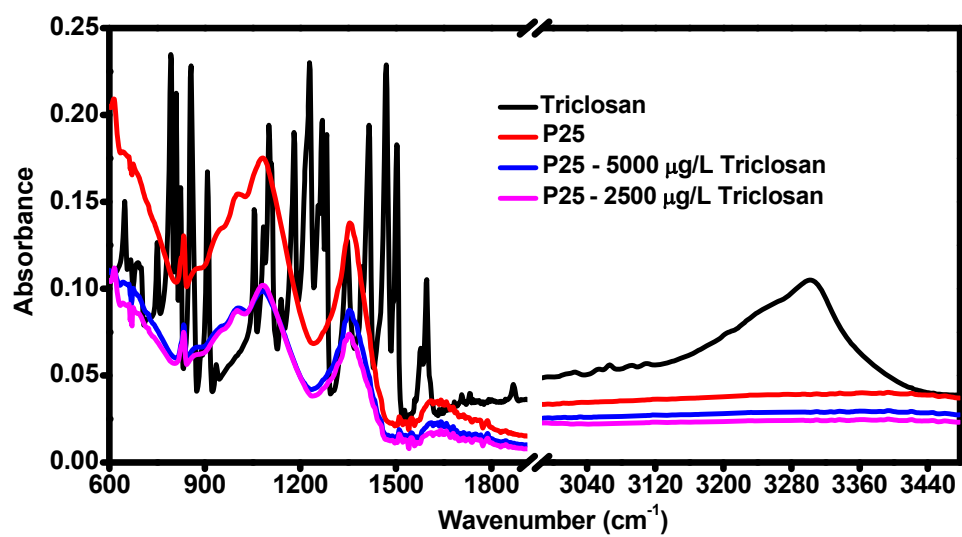
**Figure S3.** TEM micrograph of P25.



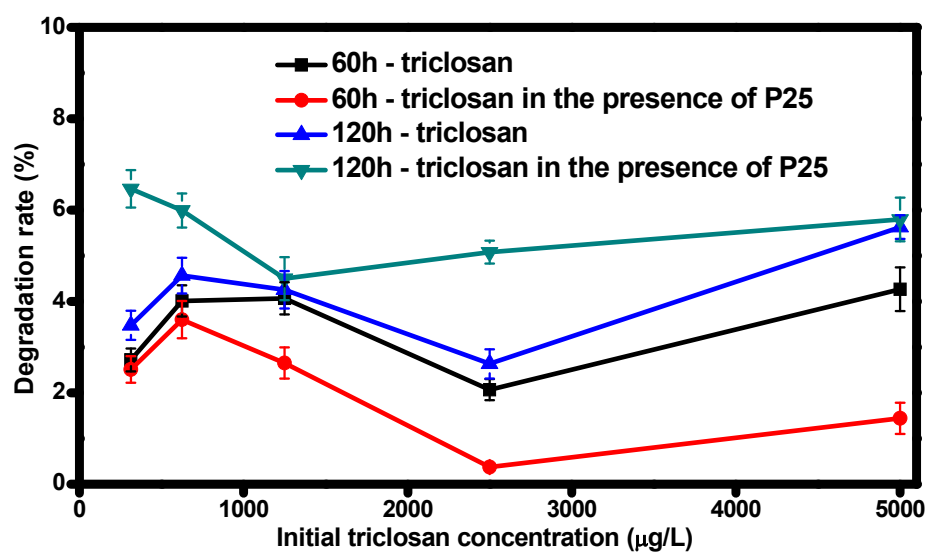
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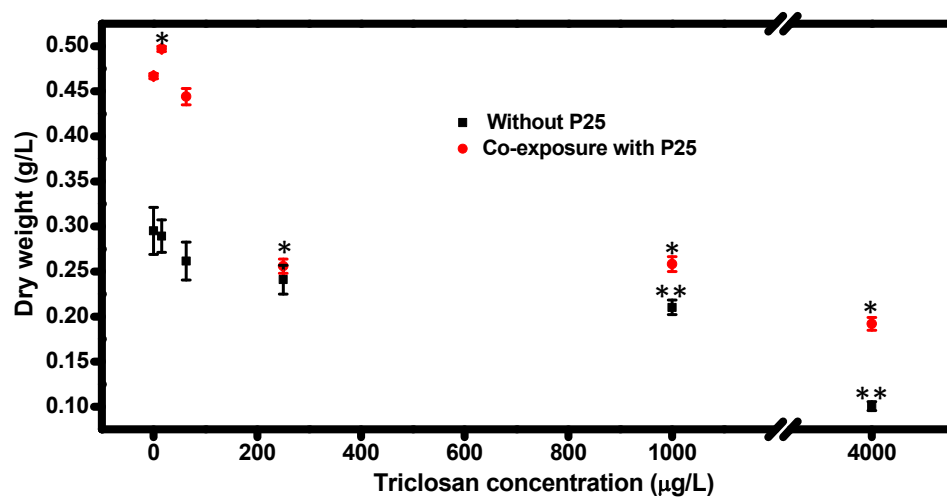
**Figure S5.** Zeta potential for 5 mg/L P25 in the presence of triclosan (n=3).



**Figure S6.** ATR-FTIR spectra of pure triclosan, P25, P25 – 5000 μg/L triclosan, and P25 – 2500 μg/L triclosan.



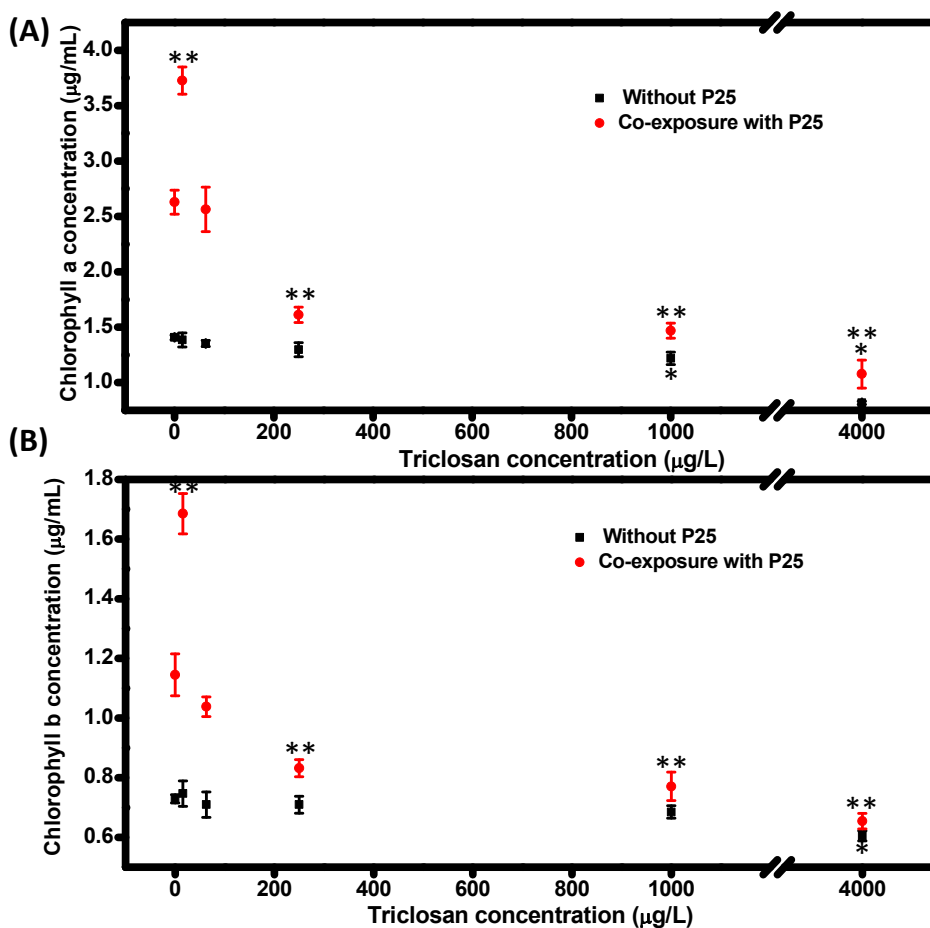
**Figure S7.** Degradation rate of triclosan in the absence and presence of 5 mg/L P25 in LEW.



**Figure S8.** Dry weight of *Eremosphaera* exposed to triclosan in the absence and presence of 5 mg/L P25 (n=3).

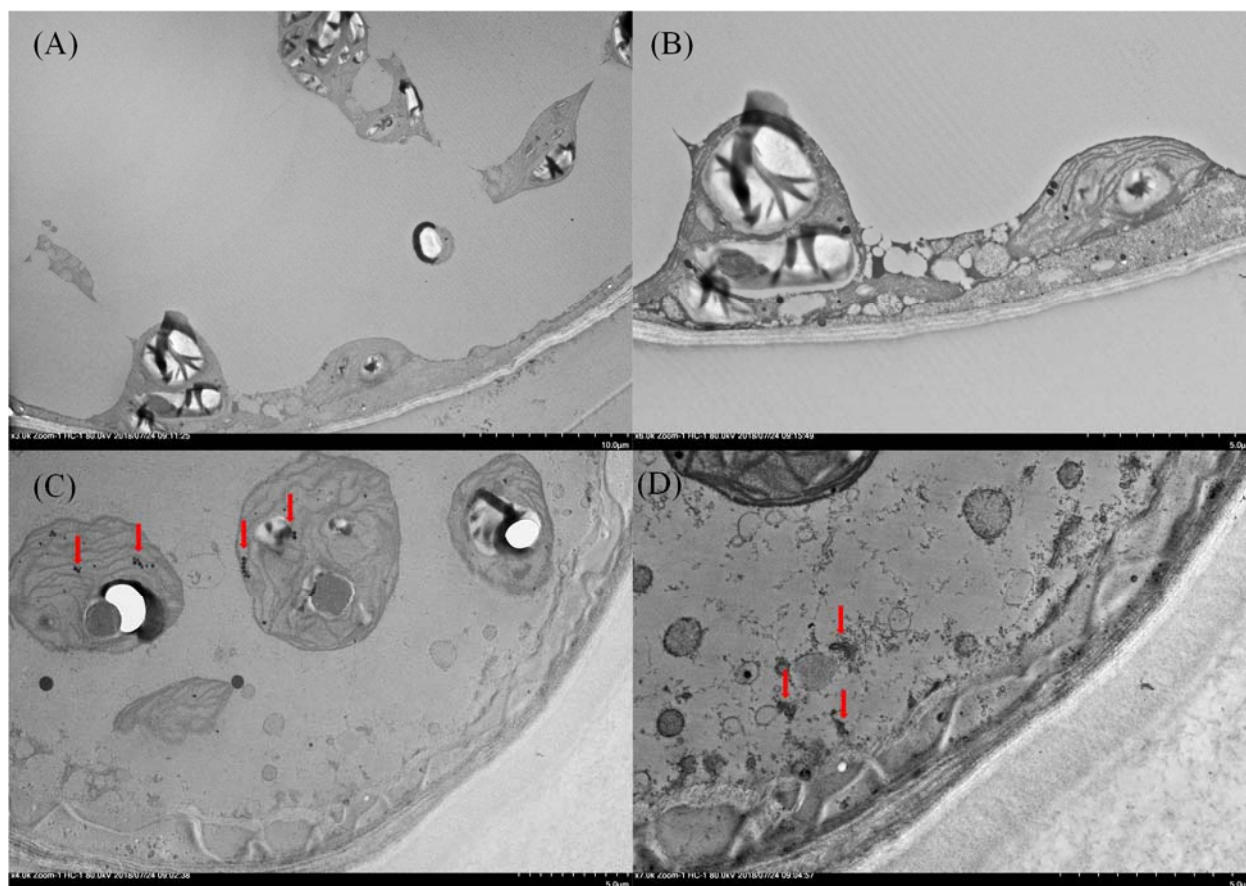
Note: “\*” represents results that were significantly different from the untreated control ( $p < 0.05$ ); “\*\*” represents results that were significantly different from the control only exposed to 5 mg/L P25 ( $p < 0.05$ ).





**Figure S9.** The variation of chlorophyll content in cultures exposed to triclosan in the absence and presence of 5 mg/L P25 (n=3). (A) chlorophyll a, (B) chlorophyll b.

Note: “\*” represents results that were significantly different from the untreated control ( $p < 0.05$ ); “\*\*” represents results that were significantly different from the control only exposed to 5 mg/L P25 ( $p < 0.05$ ).



**Figure S10.** TEM images of *Eremosphaera*. (A) (B) Cells exposed to 1000 µg/L triclosan; (C) (D) Cells co-exposed to 5 mg/L P25 and 1000 µg/L triclosan.

Note: Red arrow indicated P25 inside the cells.

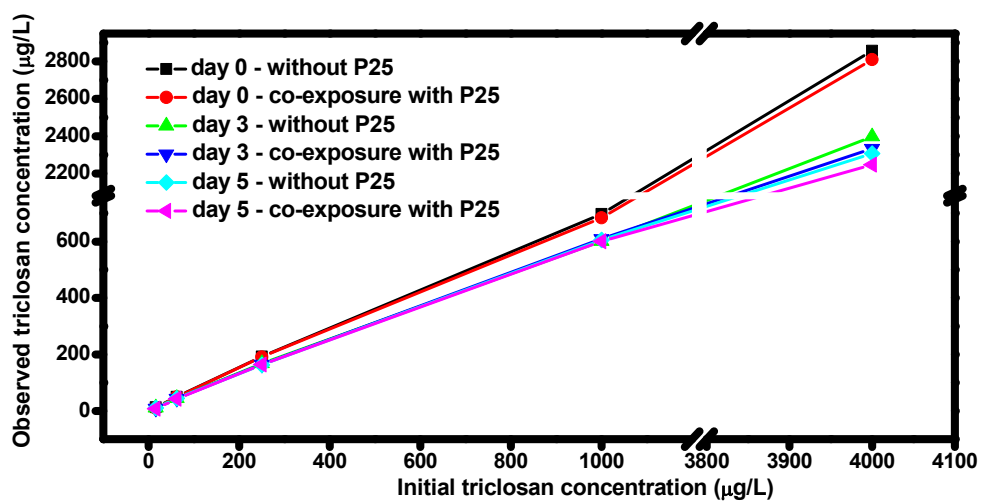
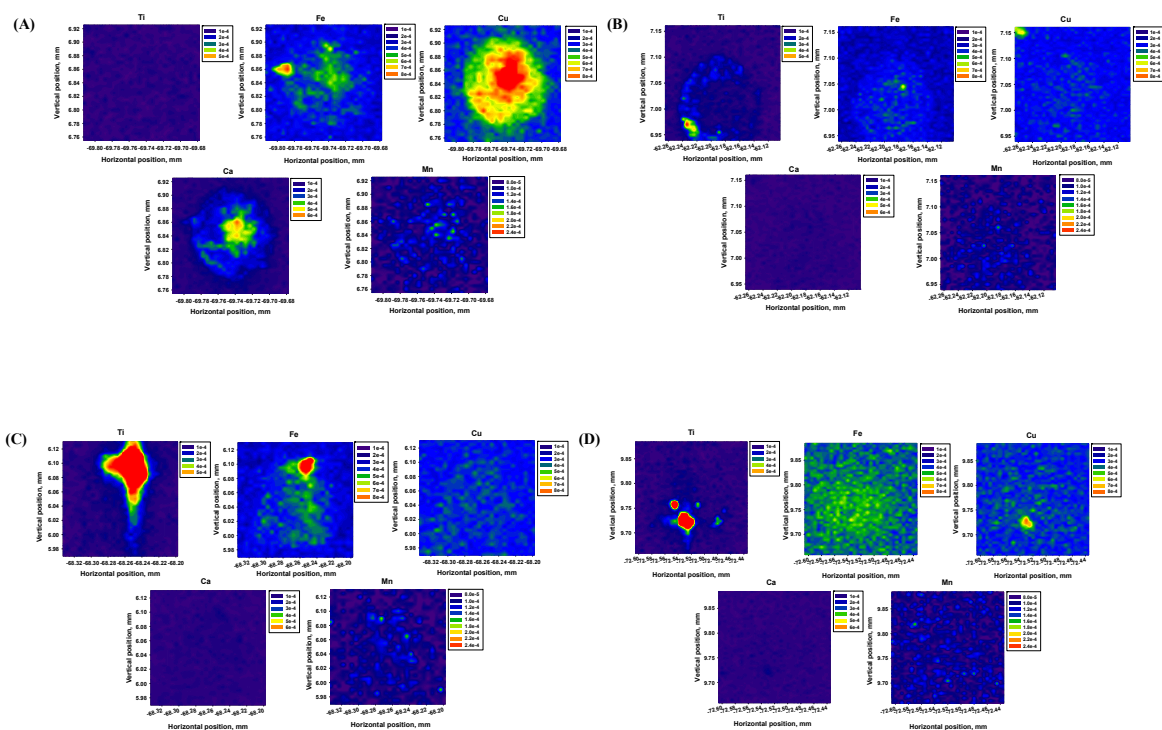
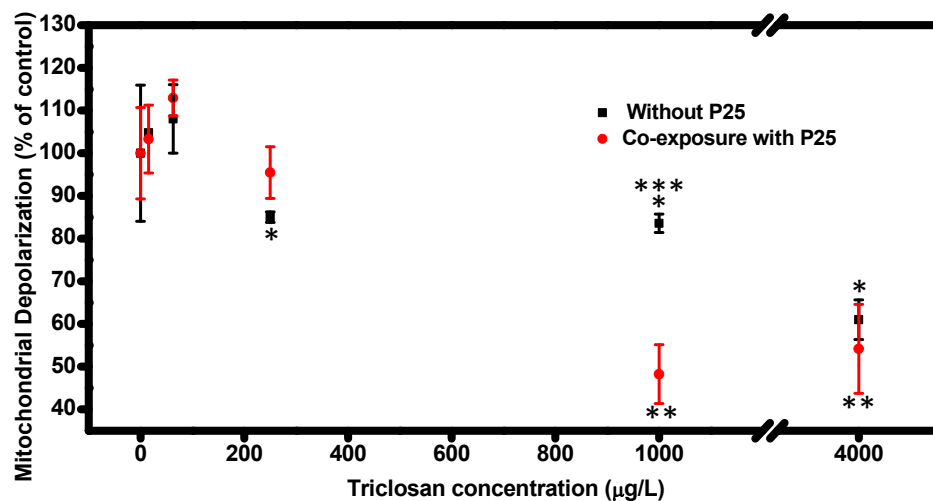


Figure S11. Observed triclosan concentration in algal media.



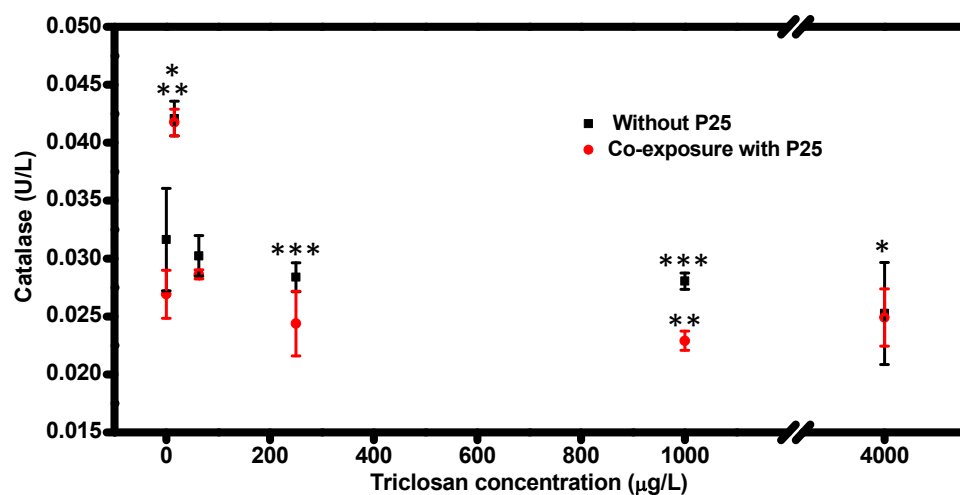
**Figure S12.** Partial investigation of the distribution of multi-elements in an individual algal cell.

(A) Cells exposed to 250 µg/L triclosan; (B) Cells exposed to 5 mg/L P25; (C) Cells co-exposed to 5 mg/L P25 and 62.5 µg/L triclosan; (D) Cells co-exposed to 5 mg/L P25 and 250 µg/L triclosan.



**Figure S13.** Mitochondrial depolarization of algal cells exposed to triclosan in the absence and presence of 5 mg/L P25 (n=3).

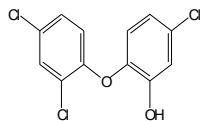
Note: “\*” represents results that were significantly different from the untreated control ( $p < 0.05$ ); “\*\*\*” represents results that were significantly different from the control only exposed to 5 mg/L P25 ( $p < 0.05$ ). “\*\*\*\*” represents results that were significantly different in treatments between in the absence and in the presence of 5 mg/L P25 ( $p < 0.05$ ).



**Figure S14.** Catalase of algal cells exposed to triclosan in the absence and presence of 5 mg/L P25 (n=3).

Note: “\*” represents results that were significantly different from the untreated control ( $p < 0.05$ ); “\*\*\*” represents results that were significantly different from the control only exposed to 5 mg/L P25 ( $p < 0.05$ ). “\*\*\*\*” represents results that were significantly different in treatments between in the absence and in the presence of 5 mg/L P25 ( $p < 0.05$ ).

**Table S1.** Triclosan given with its molecular information, water solubility and application area.<sup>a</sup>

Characteristics	Triclosan
Molecular Formula	C <sub>12</sub> H <sub>7</sub> Cl <sub>3</sub> O <sub>2</sub>
Molecular Structure	
MW (g/mol)	289.54
Solubility (mg/L)	10
pKa	8.1
Application	Antimicrobial agent

<sup>a</sup>This table has been published in Xin et al. 2018.<sup>12</sup>

Reprinted from Environmental Pollution, Volume 226, Xiaying Xin, Guohe Huang, Xia Liu, Chunjiang An, Yao Yao, Harold Weger, Peng Zhang, Xiujuan Chen, Molecular toxicity of triclosan and carbamazepine to green algae *Chlorococcum* sp.: A single cell view using synchrotron-based Fourier transform infrared spectromicroscopy, 12-20, Copyright (2017), with permission from Elsevier.

**Table S2.** The ratios of triclosan concentration variation in the presence of P25 to that in the absence of P25 in nonaqueous phase.

Triclosan concentration ( $\mu\text{g/L}$ )	Day 0	Day 5
4000	1.04180	1.03523
1000	1.04679	1.02336
250	1.05003	1.02900
62.5	1.05307	1.01713
15.625	2.20821	1.47657



## References

- (1) Wen, Y.; Chen, H.; Shen, C.; Zhao, M.; Liu, W. Enantioselectivity tuning of chiral herbicide dichlorprop by copper: roles of reactive oxygen species. *Environ. Sci. Technol.* **2011**, *45*, (11), 4778-4784.
- (2) White, E.; Payne, G. Chlorophyll production, in response to nutrient additions, by the algae in Lake Rotorua water. *New Zeal. J. Mar. Fresh.* **1978**, *12*, (2), 131-138.
- (3) Jeffrey, S. t.; Humphrey, G. New spectrophotometric equations for determining chlorophylls a, b, c 1 and c 2 in higher plants, algae and natural phytoplankton. *Biochem. Physiol. Pflanz.* **1975**, *167*, (2), 191-194.
- (4) Zhao, J.; Cao, X.; Liu, X.; Wang, Z.; Zhang, C.; White, J. C.; Xing, B. Interactions of CuO nanoparticles with the algae *Chlorella pyrenoidosa*: adhesion, uptake, and toxicity. *Nanotoxicology* **2016**, *10*, (9), 1297-1305.
- (5) Bickley, R. I.; Gonzalez-Carreno, T.; Lees, J. S.; Palmisano, L.; Tilley, R. J. A structural investigation of titanium dioxide photocatalysts. *J. Solid State Chem.* **1991**, *92*, (1), 178-190.
- (6) Hartmann, N.; Von der Kammer, F.; Hofmann, T.; Baalousha, M.; Ottofuelling, S.; Baun, A. Algal testing of titanium dioxide nanoparticles—testing considerations, inhibitory effects and modification of cadmium bioavailability. *Toxicology* **2010**, *269*, (2-3), 190-197.
- (7) Orvos, D. R.; Versteeg, D. J.; Inauen, J.; Capdevielle, M.; Rothenstein, A.; Cunningham, V. Aquatic toxicity of triclosan. *Environ. Toxicol. Chem.* **2002**, *21*, (7), 1338-1349.
- (8) Singer, H.; Müller, S.; Tixier, C.; Pillonel, L. Triclosan: occurrence and fate of a widely used biocide in the aquatic environment: field measurements in wastewater treatment plants, surface waters, and lake sediments. *Environ. Sci. Technol.* **2002**, *36*, (23), 4998-5004.
- (9) Hurum, D. C.; Agrios, A. G.; Gray, K. A.; Rajh, T.; Thurnauer, M. C. Explaining the enhanced photocatalytic activity of Degussa P25 mixed-phase TiO<sub>2</sub> using EPR. *J. Phys. Chem. B* **2003**, *107*, (19), 4545-4549.
- (10) Rafqah, S.; Wong-Wah-Chung, P.; Nelieu, S.; Einhorn, J.; Sarakha, M. Phototransformation of triclosan in the presence of TiO<sub>2</sub> in aqueous suspension: Mechanistic approach. *Appl. Catal. B. Environ.* **2006**, *66*, (1-2), 119-125.
- (11) Calabrese, E. J.; Iavicoli, I.; Calabrese, V. Hormesis: why it is important to biogerontologists. *Biogerontology* **2012**, *13*, (3), 215-235.
- (12) Xin, X.; Huang, G.; Liu, X.; An, C.; Yao, Y.; Weger, H.; Zhang, P.; Chen, X. Molecular toxicity of triclosan and carbamazepine to green algae *Chlorococcum* sp.: A single cell view using synchrotron-based Fourier transform infrared spectromicroscopy. *Environ. Pollut.* **2017**, *226*, 12-20.
- (13) Pan, C.-G.; Peng, F.-J.; Shi, W.-J.; Hu, L.-X.; Wei, X.-D.; Ying, G.-G. Triclosan-induced transcriptional and biochemical alterations in the freshwater green algae *Chlamydomonas reinhardtii*. *Ecotoxicol. Environ. Safety* **2018**, *148*, 393-401.
- (14) Middepogu, A.; Hou, J.; Gao, X.; Lin, D. Effect and mechanism of TiO<sub>2</sub> nanoparticles on the photosynthesis of *Chlorella pyrenoidosa*. *Ecotoxicol. Environ. Safety* **2018**, *161*, 497-506.
- (15) Xin, X.; Huang, G.; An, C.; Raina-Fulton, R.; Weger, H. Insights into Long-term Toxicity of Triclosan to Freshwater Green Algae in Lake Erie. *Environ. Sci. Technol.* **2019**.
- (16) Stebbing, A. Hormesis—the stimulation of growth by low levels of inhibitors. *Sci. Total Environ.* **1982**, *22*, (3), 213-234.
- (17) González-Pleiter, M.; Rioboo, C.; Reguera, M.; Abreu, I.; Leganés, F.; Cid, Á.; Fernández-Piñas, F. Calcium mediates the cellular response of *Chlamydomonas reinhardtii* to the emerging aquatic pollutant Triclosan. *Aquat. Toxicol.* **2017**, *186*, 50-66.
- (18) Perry, S. W.; Norman, J. P.; Barbieri, J.; Brown, E. B.; Gelbard, H. A. Mitochondrial membrane potential probes and the proton gradient: a practical usage guide. *Biotechniques* **2011**, *50*, (2), 98-115.
- (19) Franz, S.; Altenburger, R.; Heilmeier, H.; Schmitt-Jansen, M. What contributes to the sensitivity of microalgae to triclosan? *Aquat. Toxicol.* **2008**, *90*, (2), 102-108.

- (20) Newton, A. P. N.; Cadena, S. M. S.; Rocha, M. E. M.; Carnieri, E. G. S.; De Oliveira, M. B. M. Effect of triclosan (TRN) on energy-linked functions of rat liver mitochondria. *Toxicol. Lett.* **2005**, *160*, (1), 49-59.
- (21) Hou, J.; Wang, L.; Wang, C.; Zhang, S.; Liu, H.; Li, S.; Wang, X. Toxicity and mechanisms of action of titanium dioxide nanoparticles in living organisms. *J. Environ. Sci.* **2018**.
- (22) Roy, B.; Chandrasekaran, H.; Krishnan, S. P.; Chandrasekaran, N.; Mukherjee, A. UVA pre-irradiation to P25 titanium dioxide nanoparticles enhanced its toxicity towards freshwater algae *Scenedesmus obliquus*. *Environ. Sci. Pollut. Res.* **2018**, 1-14.
- (23) Cohen, G.; Heikkila, R. E. The generation of hydrogen peroxide, superoxide radical, and hydroxyl radical by 6-hydroxydopamine, dialuric acid, and related cytotoxic agents. *J. Biol. Chem.* **1974**, *249*, (8), 2447-2452.
- (24) Peng, Y.; Luo, Y.; Nie, X.-P.; Liao, W.; Yang, Y.-F.; Ying, G.-G. Toxic effects of Triclosan on the detoxification system and breeding of *Daphnia magna*. *Ecotoxicology* **2013**, *22*, (9), 1384-1394.
- (25) Mhamdi, A.; Queval, G.; Chaouch, S.; Vanderauwera, S.; Van Breusegem, F.; Noctor, G. Catalase function in plants: a focus on Arabidopsis mutants as stress-mimic models. *J. Exp. Bot.* **2010**, *61*, (15), 4197-4220.