

Continuous measurement of diffusive and ebullitive fluxes of methane in aquatic ecosystems by an open dynamic chamber method

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Supporting information

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S1. Literature review

Table S1. List of previous reports on the use of ODC (ordered chronologically).

Ecosystem	Gas analyzed	Author	Year
Soil	CO ₂	Edward & Sollins ¹	1973
Soil	N ₂ O	Denmead ²	1979
Soil	NH ₃	Svensson ³	1994
Soil	CO ₂	Fang & Moncrieff ⁴	1996
Soil	CH ₃ Br	Gao <i>et al.</i> ⁵	1997
Soil	CH ₂ Cl ₂	Gao & Yates ⁶	1998
Soil	VOC	Gao & Yates ⁷	1998
Soil	Hg	Carpi & Lindberg ⁸	1998
Soil	Hg	Zhang <i>et al.</i> ⁹	2002
Soil	Hg	Lindberg <i>et al.</i> ¹⁰	2002
Soil	NO	Tabachow <i>et al.</i> ¹¹	2002
Soil	VOC	Reichman & Rolston ¹²	2002
Soil	CO ₂	Pumpanen <i>et al.</i> ¹³	2004
Soil	CH ₄ , N ₂ O	Denmead ¹⁴	2008
Soil	CO ₂ , H ₂ O, NO, NO ₂ , O ₃	Pape <i>et al.</i> ¹⁵	2009
Plants	NO, NO ₂ , O ₃	Breuninger <i>et al.</i> ¹⁶	2012
Peatland	CH ₄	Yu <i>et al.</i> ¹⁷	2014
Soil	H ₂ O	Marasco <i>et al.</i> ¹⁸	2014
Soil	BTEX	Verginelli <i>et al.</i> ¹⁹	2018

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24 **S2. ODC characteristics**

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26 **Table S2.** Main power and gas requirements, autonomy and approximate cost of the ODC
 27 method, as used in the present work.

Power requirement (12 V)	60	W
Energy autonomy (50 Ah car battery)	8-10	h
CH ₄ -free nitrogen flow rate	7.7	L min ⁻¹
Autonomy "BT-80" cylinder (2300 L capacity)	5.0	h
UGGA datalogger autonomy	>100	d
UGGA (approximate cost)	40000	US\$
Mass flow controller (approximate cost)	1000	US\$
Battery, chamber, tubing and accessories (approximate cost)	300	US\$

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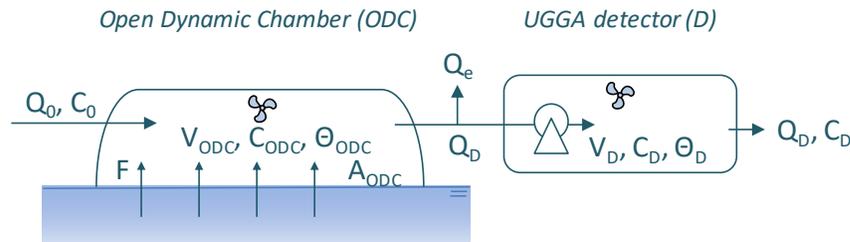
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30 **S3. Flowchart and mass balance equations**

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32 The flowchart of the ODC method is shown in Figure S1. Our model considers that the gas
 33 phase is fully mixed, both in the chamber and in the spectrometer cavity. Although in the
 34 present work we used CH₄- and CO₂-free nitrogen as carrier gas, the model considers an
 35 input concentration for increased versatility.

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39 **Figure S1:** Flowchart of the ODC method.

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41 Where:

42 F : CH₄ flux (g m⁻² s⁻¹),

43 C_0 : Influent carrier gas concentration (g m⁻³)

44 C_{ODC} : Gas concentration in the ODC (g m⁻³)

45 C_D : Measured concentration in the UGGA spectrometer (g m⁻³)

46 Q_0 : Carrier gas influent flow rate (m³ h⁻¹)

47 Q_e : Purge, excess flow rate (m³ h⁻¹)

48 Q_D : Flow rate extracted by the spectrometer (m³ h⁻¹)

49 A_{ODC} : Area of the dynamic chamber in contact with the ecosystem (m²)

50 V_{ODC} : Volume of the dynamic chamber (m³)

51 V_D : Volume of the spectrometer cavity (m³)

52 Θ_{ODC} : Gas residence time in the dynamic chamber (s)

53 Θ_D : Gas residence time in the spectrometer cavity (s)

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55 Assuming that the concentration in the spectrometer cavity (C_D) is homogeneous, the mass
 56 balance in that cavity can be described by:

$$57 \frac{dC_D}{dt} = \frac{Q_D}{V_D} \cdot C_{ODC} - \frac{Q_D}{V_D} \cdot C_D \quad (S1)$$

59 In turn, assuming that the cavity of the spectrometer is fully mixed, the residence time (θ_D)
 60 can be described by:

$$61 \frac{1}{\theta_D} = \frac{Q_D}{V_D} \quad (S2)$$

62 Such that we can write:

$$63 \frac{dC_D}{dt} = \frac{1}{\theta_D} \cdot (C_{ODC} - C_D) \quad (S3)$$

$$64 C_{ODC} = \theta_D \frac{dC_D}{dt} + C_D \quad (S4)$$

65 Equation S4 allows the determination of C_{ODC} , at any time, from spectrometer measurements.
 66 Similarly, assuming that the headspace of the Open Dynamic Chamber is fully mixed, the
 67 mass balance can be described by:

$$68 \frac{dC_{ODC}}{dt} = \frac{Q_0}{V_{ODC}} \cdot C_0 - \frac{Q_e}{V_{ODC}} \cdot C_{ODC} - \frac{Q_D}{V_{ODC}} \cdot C_{ODC} + F \cdot \frac{A_{ODC}}{V_{ODC}} \quad (S5)$$

69 Assuming mass conservation, the influent flow (Q_0) equals the sum of the effluent flows (Q_e
 70 + Q_D), and Eq. S5 can be rewritten as follows:

$$71 \frac{dC_{ODC}}{dt} = \frac{Q_0}{V_{ODC}} \cdot (C_0 - C_{ODC}) + F \cdot \frac{A_{ODC}}{V_{ODC}} \quad (S6)$$

72 The residence time in dynamic chamber (θ_{ODC}) can be described by:

$$73 \frac{1}{\theta_{ODC}} = \frac{Q_0}{V_{ODC}} \quad (S7)$$

74 By rearranging Eq. S6, we obtain:

$$75 F = \left(\frac{dC_{ODC}}{dt} + \frac{C_{ODC} - C_0}{\theta_{ODC}} \right) \cdot \frac{V_{ODC}}{A_{ODC}} \quad (S8)$$

76 Which, in combination with Eq. S4, results in:

$$77 F = \left(\frac{d(\theta_D \frac{dC_D}{dt} + C_D)}{dt} + \frac{\theta_D \frac{dC_D}{dt} + C_D - C_0}{\theta_{ODC}} \right) \cdot \frac{V_{ODC}}{A_{ODC}} \quad (S9)$$

94 Thus, Equation S9 allows the measurement of instantaneous flux from spectrometer
95 measurements. It is important to note that the UGGA detector measures gas concentrations
96 in ppm ($C_{D,ppm}$), whereas in Equation S9 the concentrations are expressed in $g\ m^{-3}$. To convert
97 $C_{D,ppm}$ we used Equations S10 and S11; where M_{CH_4} is the molar weight of CH_4 ($16\ g\ mol^{-1}$),
98 V_M is the molar volume of the gas of interest ($m^3\ mol^{-1}$; Eq. S11), 1,000,000 is the
99 conversion factor from ppm to vol./vol., P is the atmospheric pressure during measurements
100 (Atm), R is the universal gas constant ($8.2058\ 10^{-5}\ m^3\ Atm\ K^{-1}\ Mol^{-1}$), and T is the absolute
101 temperature during measurements (K):

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$$103\ C_D = \frac{C_{D,ppm}}{1,000,000} \cdot \frac{M_{CH_4}}{V_M} \quad (S10)$$

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$$105\ V_M = \frac{R \cdot T}{P} \quad (S11)$$

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108 **S4. Data processing.**

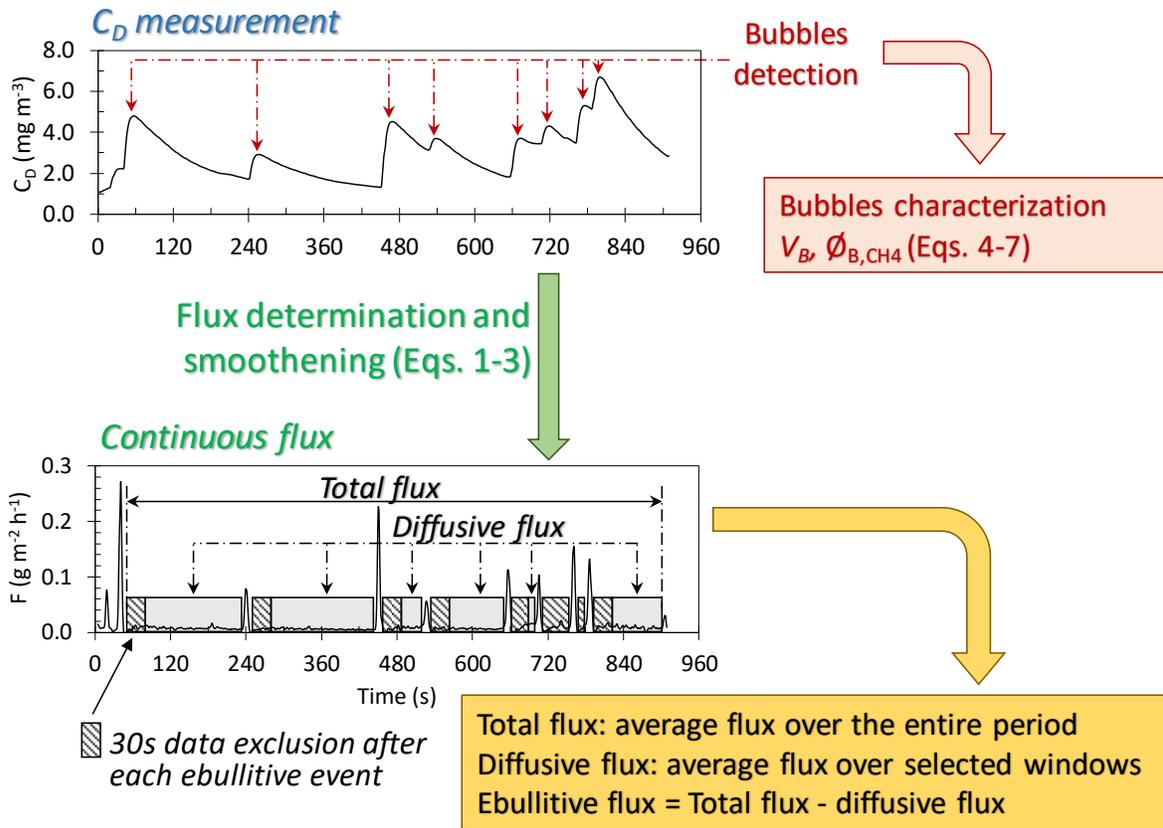
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110 **S4.1. Flux measurement**

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112 During field application of the ODC method, as described above, we established the best
113 method for data processing as a combination of flux determination and bubble
114 characterization. The strategy is described in Figure S4, on the basis of an actual
115 measurement made at a highly ebullitive region of LG. After calibration, the ODC was
116 positioned on the surface of the ecosystem. About 30 s were required for the system to
117 stabilize. Then, C_D was continuously acquired at a frequency of 1 Hz. Data treatment
118 consisted, first, in the detection of ebullitive event and the characterization of bubbles,
119 according to Equations 4–7. Then, C_D measurements were converted to flux (Eq. 1), which
120 were smoothed (Eqs. 1–3).

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Figure S2: Sequence of data processing.

As shown in the Results and Discussion section, just after an ebullition event, a significant instability was observed in flux data for 10–15 s, which corresponded to the period during which the C_D concentration was reaching the peak concentration and returning to a decreasing trend. Consequently, to reduce errors in diffusive flux calculations, we decided to discard 30 s of data after each ebullitive event. As shown in Figure S2, total flux was then determined as the average of the complete set of flux data (without exclusion periods), while diffusive flux was determined as the average of diffusive flux over the selected periods, and ebullitive flux was determined from the difference between total and diffusive flux.

S4.2. SNR measurement

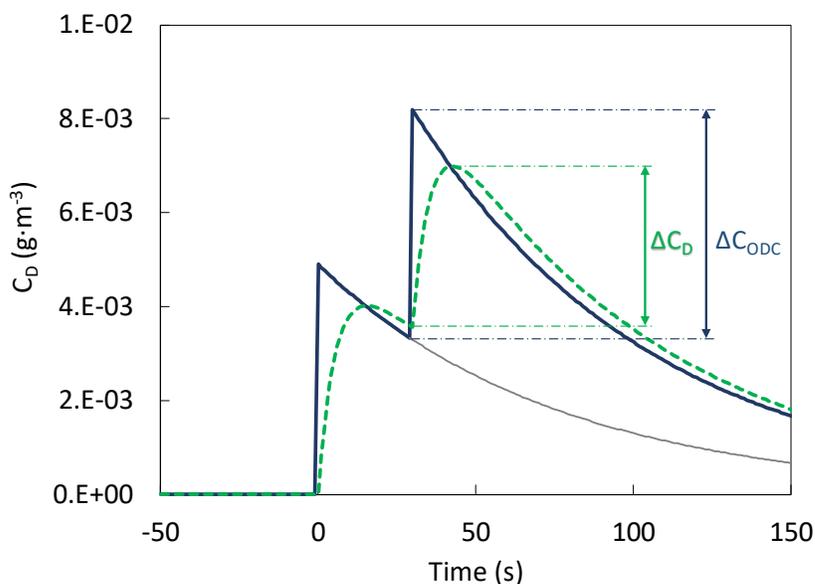
SNR of C_D was calculated as follows; first, we selected 10 min regions of the data sets obtained from the laboratory and field deployment where steady state was observed; i.e., absence of ebullition and visible trend, and final C_D values not significantly different than initial values. Second, the SNR was calculated from the standard deviations and means of 15 C_D data.

146 **S5. Bubbles characterization**

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148 As shown in the results sections, instantaneous total flux was successfully determined with
149 Equation S9. Ebullitive events were clearly detectable by an abrupt and short increase in flux.
150 However, a significant noise in the instantaneous flux determination was observed,
151 particularly when two ebullitive events were occurring in a short period of time, resulting in
152 a significant error and impeding the precise determination of the bubbles characteristics.
153 Therefore, a different strategy was implemented, which was based on the determination of
154 the maximum concentration observed after an ebullitive event, instead of the instantaneous
155 flux determination. Figure S2 illustrates the concentration observed in the ODC and in the
156 cavity of the UGGA detector during a double ebullitive event. This simulation was done,
157 using the mass balance equations (Eq. S1 and S6; Figure S3), to illustrate the strategy used
158 for the characterization of bubbles, presented in the main document.

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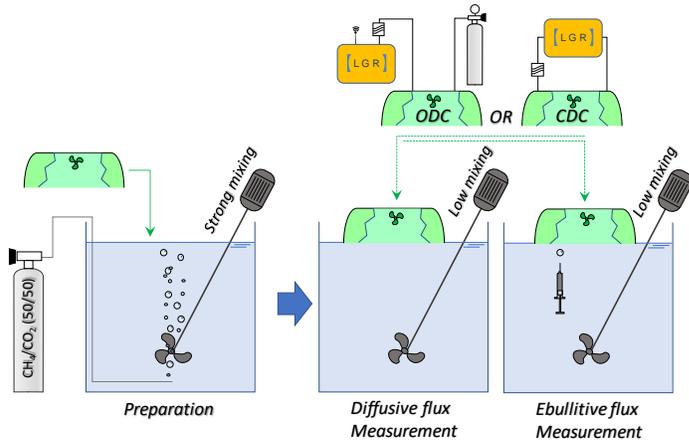
161 **Figure S3:** Theoretical simulation of the concentration observed in the Chamber (C_{ODC} ; —)
162 and in the UGGA detector (C_D ; ---), that would be observed during a double ebullitive event
163 with a 30 s interval. This simulation was obtained from mass balance equations, under the
164 following conditions; bubbles diameter, 5 mm; bubble composition, 100% CH₄; θ_{ODC} , 76 s;
165 θ_D , 6 s; T, 293.15 K, P, 1 atm.

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S6 Laboratory testing.

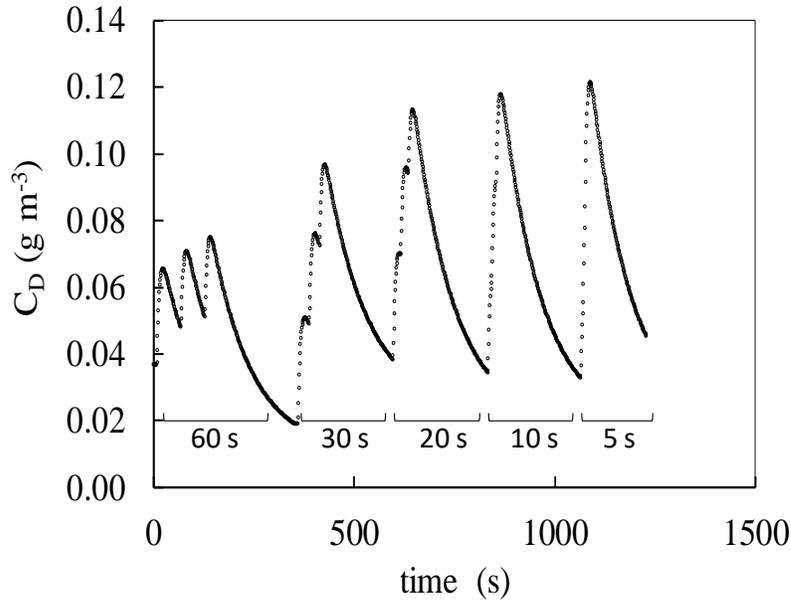


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Figure S4: Experimental set-up for laboratory testing.

S7 Results and discussion

S7.1 CH₄ flux measurements



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Figure S5: CH₄ concentration (C_D) observed during the injection of three 2 mL bubbles (CH₄/CO₂; 60/40% vol.) with an interval of 60, 30, 20, 10 and 5 s.

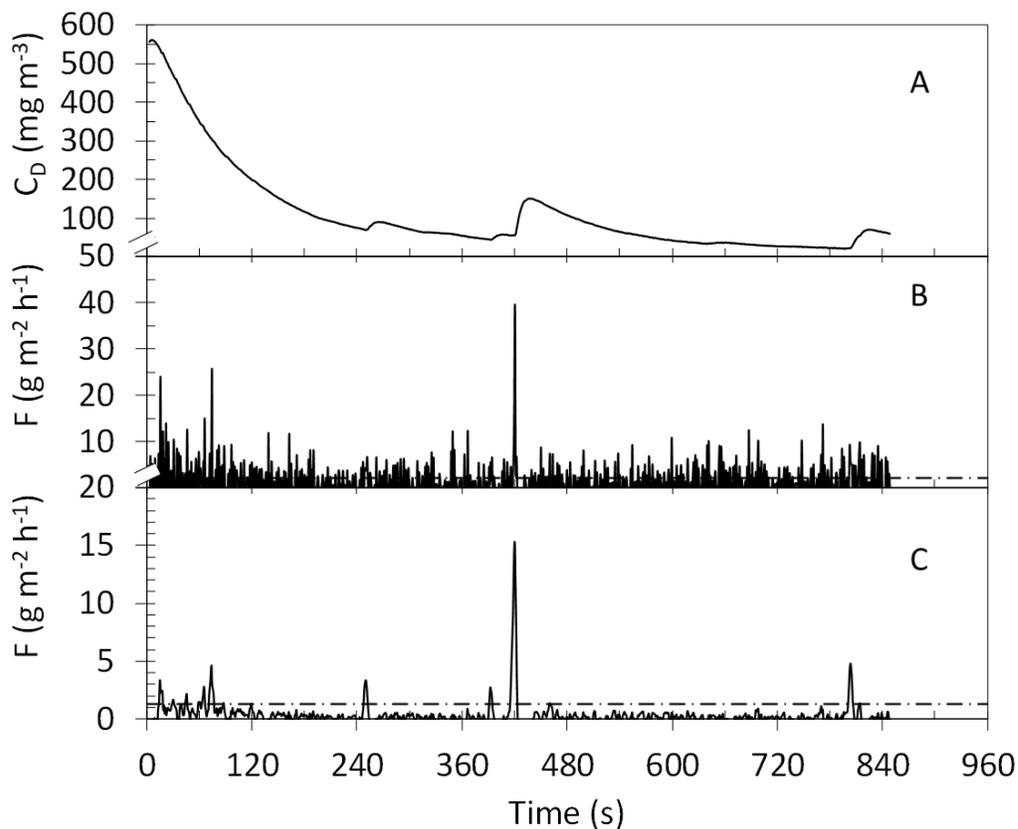
S7.2 CO₂ flux measurements

187 First, we tested the ODC method in the laboratory, for CO₂ diffusive fluxes and for the
188 characterization of bubbles, as previously done for CH₄ (Figs. 2 and 3). The results (not
189 shown) were similar to those observed with CH₄ (in both cases $R^2 > 0.96$). However, during
190 the field deployment, we observed noisier signals from the UGGA, which were quantified
191 through an SNR of 343 ± 82 , which was significantly lower than that observed with CH₄. In
192 LG, the correlation between diffusive fluxes measured with the standard CDC and the ODC
193 was significant (R^2 of 0.92, $p < .01$); in LL, however, erratic results were observed.
194 Undoubtedly, the failure of the method in LL was caused by the relatively low UGGA
195 sensitivity to CO₂ (reported by the manufacturer to be 1 ppm). The mean CO₂ diffusive flux
196 in LG and LL, measured with the standard CDC, was 70.9 ± 26 and 3.0 ± 2.5 mg m⁻² h⁻¹,
197 respectively. These CO₂ fluxes corresponded to a C_D of 28.3 ± 10.4 and 1.21 ± 1.00 ppm,
198 respectively. Thus, in LG, the CO₂ flux was in the region of the UGGA detection limit (1
199 ppm). From these results, we arbitrarily estimated that the limit of detection of the ODC
200 method was about 10 mg CO₂ m⁻² h⁻¹, which corresponded to a C_D of 4 ppm. This minimum
201 flux that can be measured is higher than the lower range of CO₂ fluxes reported by Rasilo et
202 al.²⁰ in 224 lakes.

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204 During the combined diffusive and ebullitive flux measurements in LG, we observed
205 abrupt increases in C_D and their corresponding CO₂ peaks in the flux data (Fig. S7). However,
206 in this case, to distinguish peak fluxes from the background noise, we needed a cut-off index
207 for an instantaneous flux of 2.15 g CO₂ m⁻² h⁻¹, which corresponds to a bubble content of
208 31.2×10^{-3} mL of CO₂ or to bubbles of 12.6–18.1 mm in diameter if they contain 3–1% CO₂.
209 This range is above the higher range of bubble diameters previously reported in lakes, 2.6–
210 11.4 mm,^{21–24} which indicates that the ODC method, under the experimental conditions
211 tested, presents serious limitations. The latter was confirmed by our field experiments, as
212 CO₂ was detected only in the largest bubbles. Supposing that the bubbles were only
213 composed of CH₄ and CO₂; i.e. the two gases detected by the UGGA, the smallest bubble
214 with CO₂ detected (0.29 mL total volume or 8.2 mm diameter) contained 17% CO₂. Only 23
215 bubbles (27%) containing CO₂ were detected out of the 84 with a volume >0.29 mL. These
216 bubbles contained, on average, 8.5% CO₂. Clearly, these percentages are not representative
217 of the entire bubble population, but only of the largest bubbles containing a mass of CO₂
218 above the detection limit of the method. Thus, the CO₂ percentages reported here are
219 significantly higher than most previous reports in the literature; i.e., 0.1–3%,^{25–27} although
220 Poissant et al.²⁸ reported a range of 5–15% CO₂ in a river.

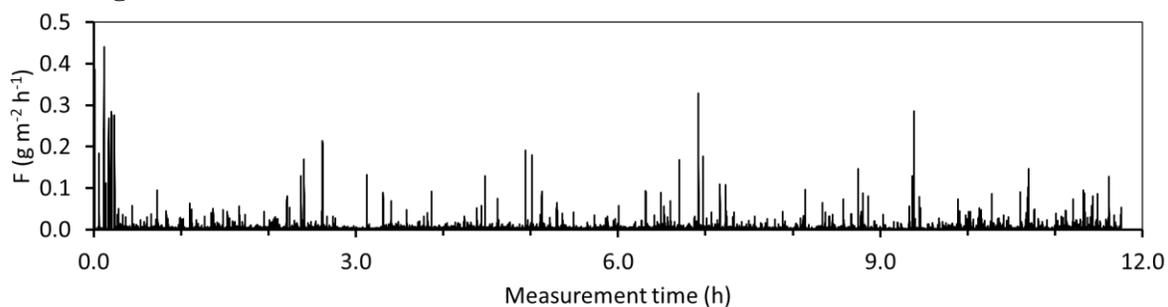
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Figure S6. Example of continuous CO₂ flux measurement by ODC in Lake Guadalupe; A. C_D as measured; B. unfiltered flux (Eq. 1); C. double smoothed flux (Eq. 2, 3). Horizontal line in panel B shows the cut-off index.

S7.3 Long term measurement of CH₄ emissions



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Figure S7. Total CH₄ flux observed over a 12-h continuous measurement. Time 0 corresponds to midnight in such manner that the time corresponds to the hour of the day.

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241 **Table S3:** Mean (± 1 standard deviation) of total, diffusive and ebullitive flux ($\text{mg m}^{-2} \text{h}^{-1}$)
242 observed over three distinctive periods of the day. Means and standard deviations were
243 determined by equally weighted running means of 15 minutes data subsets.

Time period	Time period	Total Flux	Diffusive flux	Ebullitive flux
Night	0:00-3:00	6.0 ± 17.4	3.4 ± 2.1	2.5 ± 17.6
Morning	9:00-12:00	5.2 ± 10.9	2.9 ± 2.5	2.3 ± 11.0
Afternoon	15:00-18:00	5.0 ± 8.4	3.5 ± 2.1	1.5 ± 8.7
	Total	5.3 ± 11.7	3.4 ± 2.2	1.9 ± 11.9

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245 **Table S4:** Comparison between ODC and other point flux measurement methods.

	Boundary Layer ²⁹	Bubbles traps ³⁰	Hydroacoustic bubble detection ²¹	Bubble survey ³¹	Sensor on robotic boat ³²	Closed static chamber ¹³	Closed dynamic chamber ¹³	Open dynamic chamber
Ebullitive flux	No	Yes	Yes	Yes	Yes	No	Yes	Yes
Diffusive flux	Yes	No	No	No	No	Yes	Yes	Yes
Distinction ebullitive/diffusive flux	No	No	No	No	No	No	Yes/No ^a	Yes
Continuous measurement	Yes	Yes/No	Yes	No	Yes	No	Yes/No	Yes
Headspace concentration build-up	No	Yes	No	No	No	Yes	Yes	Moderate ^b
Cost (excluding detector) ^c	Low	Low	-	Low	High	Low	Low	Low
Detector sensitivity ^d	High	Low	High	None	High	Moderate/high	Moderate/high	High
Ease of field deployment	High	Moderate	Moderate	High	High	High	High	High
Mobility	High ^e	Low	High ^e	High	High	Moderate	Moderate	Moderate
Direct measurement ^f	No	Yes	No	No	No	Yes	Yes	Yes
High throughput method	Yes	Yes/No	Yes	No	Yes	No	No	Yes
Ancillary measurements	Yes ^g	No	No	No	No	No	No	No
Requirement of carrier gas	No	No	No	No	No	No	No	Yes

246 ^a Distinction between diffusive and ebullitive flux is possible, only with relatively high frequency measurements.

247 ^b The continuous carrier gas flow reduces but does not avoid concentration build-up inside the chamber headspace.

248 ^c The experimental set-up, excluding the gas detector is usually of low cost.

249 ^d The detector sensitivity required depends on the level of emissions. Bubbles traps usually require a sensitivity in the % v/v range, while the other methods require a sensitivity in the ppb-ppm v/v range.

250 ^e These methods allow for measurements in motion.³³

251 ^f Direct measurement refers to the capture of the gas emitted, in an enclosure.

252 ^g The boundary layer method requires the precise measurement of wind speed.

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256 REFERENCES

257

258 (1) Edwards, N. T.; Sollins, P. Continuous measurement of carbon dioxide evolution
259 from partitioned forest floor components. *Ecol. Soc. Am.* **1973**, 54 (2), 406–412.

260 (2) Denmead, O. T. Chamber systems for measuring nitrous oxide emission from soils
261 in the field. *Soil Sci. Soc. Am. J.* **1979**, 43 (1), 89.

262 (3) Svensson, L. A new dynamic chamber technique for measuring ammonia emissions
263 from land- spread manure and fertilizers. *Acta Agric. Scand.* **1994**, 35–46.

264 (4) Fang, C.; Moncrieff, J. B. An improved dynamic chamber technique for measuring
265 CO₂ efflux from the surface of soil. *Funct. Ecol.* **1996**, 10 (2), 297–305.

266 (5) Gao, F.; Yates, S. R.; Yates, M. V.; Gan, J.; Ernst, F. F. Design, fabrication, and
267 application of a dynamic chamber for measuring gas emissions from soil. *Environ. Sci.*
268 *Technol.* **1997**, 31 (1), 148–153.

269 (6) Gao, F.; Yates, S. R. Simulation of enclosure-based methods for measuring gas
270 emissions from soil to the atmosphere. *J. Geophys. Res.* **1998**, 103, 126–129.

271 (7) Gao, F.; Yates, S. R. Laboratory study of closed and dynamic flux chambers:
272 experimental results and implications for field application. *J. Geophys. Res.* **1998**, 103, 115–
273 125.

274 (8) Carpi, A.; Lindberg, S. E. Application of a Teflon® dynamic flux chamber for
275 quantifying soil mercury flux: tests and results over background soil. *Atmos. Environ.* **1998**,
276 32 (5), 873–882.

277 (9) Zhang, H.; Lindberg, S. E.; Barnett, M. O.; Vette, A. F.; Gustin, M. S. Dynamic flux
278 chamber measurement of gaseous mercury emission fluxes over soils. Part 1: Simulation of
279 gaseous mercury emissions from soils using a two-resistance. *Atmospheric Environment.*
280 **2002**, pp 835–846.

281 (10) Lindberg, S. E.; Zhang, H.; Vette, A. F.; Gustin, M. S.; Barnett, M. O.; Kuiken, T.
282 Dynamic flux chamber measurement of gaseous mercury emission fluxes over soils: Part 2 -
283 Effect of flushing flow rate and verification of a two-resistance exchange interface simulation
284 model. *Atmospheric Environment.* **2002**, pp 847–859.

285 (11) Tabachow, R. M.; Roelle, P. A.; Peirce, J. J.; Aneja, V. P. Soil nitric oxide emissions:
286 Lab and field measurements and comparison. *Environ. Eng. Sci.* **2002**, 19 (4), 205–214.

287 (12) Reichman, R.; Rolston, D. E. Atmospheric pollutants and trace gases design and
288 performance of a dynamic gas flux chamber. *J. Environ. Qual.* **2002**, 31, 1774–1781.

289 (13) Pumpanen, J.; Kolari, P.; Ilvesniemi, H.; Minkkinen, K.; Vesala, T.; Niinistö, S.;
290 Lohila, A.; Larmola, T.; Morero, M.; Pihlatie, M.; Janssens, I.; Curiel Yuste, Jorge.;
291 Grünzweig, J. M.; Reth, S.; Subke, J.; Savage, K.; Kutsch, W.; Østreg, G.; Zeigler, W.;
292 Anthoni, P.; Lindroth, A.; Hari, P. Comparison of different chamber techniques for
293 measuring soil CO₂ efflux. *Agric. For. Meteorol.* **2004**, 123 (3–4), 159–176.

294 (14) Denmead, O. T. Approaches to measuring fluxes of methane and nitrous oxide
295 between landscapes and the atmosphere. *Plant Soil.* **2008**, 309 (1–2), 5–24.

- 296 (15) Pape, L.; Ammann, C.; Nyfeler-Brunner, A.; Spirig, C.; Hens, K.; Meixner, F. X. An
297 automated dynamic chamber system for surface exchange measurement of non-reactive and
298 reactive trace gases of grassland ecosystems. *Biogeosciences*. **2009**, 6 (3), 405–429.
- 299 (16) Breuninger, C.; Oswald, R.; Kesselmeier, J.; Meixner, F. X. The dynamic chamber
300 method: Trace gas exchange fluxes (NO, NO₂, O₃) between plants and the atmosphere in the
301 laboratory and in the field. *Atmos. Meas. Tech.* **2012**, 5 (5), 955–989.
- 302 (17) Yu, Z.; Slater, L. D.; Schäfer, K. V. R.; Reeve, A. S.; Varner, R. K. Dynamics of
303 methane ebullition from a peat monolith revealed from a dynamic flux chamber system. *J.*
304 *Geophys. Res. Biogeosciences*. **2014**, 119 (9), 1789–1806.
- 305 (18) Marasco, D. E.; Hunter, B. N.; Culligan, P. J.; Ga, S. R.; McGillis, W. R. Quantifying
306 evapotranspiration from urban green roofs: A comparison of chamber measurements with
307 commonly used predictive methods. *Environ. Sci. Technol.* **2014**, 48, 10273–10281.
- 308 (19) Verginelli, I.; Pecoraro, R.; Baciocchi, R. Using dynamic flux chambers to estimate
309 the natural attenuation rates in the subsurface at petroleum contaminated sites. *Sci. Total*
310 *Environ.* **2018**, 619–620, 470–479.
- 311 (20) Rasilo, T.; Prairie, Y. T.; Del Giorgio, P. A. Large-scale patterns in summer diffusive
312 CH₄ fluxes across boreal lakes, and contribution to diffusive C emissions. *Glob. Chang. Biol.*
313 **2015**, 21, 1124–1139.
- 314 (21) Ostrovsky, I.; McGinnis, D. F.; Lapidus, L.; Eckert, W. Quantifying gas ebullition
315 with echosounder: the role of methane transport by bubbles in a medium-sized lake. *Limnol.*
316 *Oceanogr. Methods*. **2008**, 6 (2), 105–118.
- 317 (22) Delwiche, K. B.; Hemond, H. F. Methane bubble size distributions, flux, and
318 dissolution in a freshwater lake. *Environ. Sci. Technol.* **2017**, 51 (23), 13733–13739.
- 319 (23) Delwiche, K.; Senft-Grupp, S.; Hemond, H. A novel optical sensor designed to
320 measure methane bubble sizes in situ. *Limnol. Oceanogr. Methods*. **2015**, 13 (12), 712–721.
- 321 (24) Delsontro, T.; McGinnis, D. F.; Wehrli, B.; Ostrovsky, I. Size does matter:
322 Importance of large bubbles and small-scale hot spots for methane transport. *Environ. Sci.*
323 *Technol.* **2015**, 49 (3), 1268–1276.
- 324 (25) Martinez, D.; Anderson, M. A. Methane production and ebullition in a shallow,
325 artificially aerated, eutrophic temperate lake (Lake Elsinore, CA). *Sci. Total Environ.* **2013**,
326 454–455, 457–465.
- 327 (26) Sepulveda-Jauregui, A.; Anthony Walter, K. M.; Martinez-Cruz, K.; Greene, S.;
328 Thalasso, F. Methane and carbon dioxide emissions from 40 lakes along a north – south
329 latitudinal transect in Alaska. *Biogeosciences*. **2015**, 12, 3197–3223.
- 330 (27) Casper, P.; Maberly, S. C.; Hall, G. H.; Finlay, B. J. Fluxes of methane and carbon
331 dioxide from a small productive lake to the atmosphere. *Biogeochemistry*. **2000**, 49 (1), 1–
332 19.
- 333 (28) Poissant, L.; Constant, P.; Pilote, M.; Canário, J.; O’Driscoll, N.; Ridal, J.; Lean, D.
334 The ebullition of hydrogen, carbon monoxide, methane, carbon dioxide and total gaseous
335 mercury from the Cornwall Area of Concern. *Sci. Total Environ.* **2007**, 381 (1–3), 256–262.
- 336 (29) Bartosiewicz, E.; Laurion, I.; MacIntyre, S. Greenhouse gas emission and storage in a
337 small shallow lake. *Hydrobiologia*. **2015**, 757(1), 101–115.

- 338 (30) Flury, S.; McGinnis, D.; Gessner, M. Methane emissions from a freshwater marsh in
339 response to experimentally simulated global warming and nitrogen enrichment. *J. Geophys.*
340 *Res.* **2010**, 115(G1), 1-9.
- 341 (31) Walter Anthony, K. M.; Vas, D. A.; Brosius, L.; Chapin III, F. S.; Zimov, S. A.;
342 Zhuang, Q. Estimating methane emissions from northern lakes using ice- bubble surveys.
343 *Limnol. Oceanogr.: Methods* **2010**, 8, 592–609.
- 344 (32) Grinham, A.; Dunbabin, M.; Gale, D.; Udy, J. Quantification of ebullitive and
345 diffusive methane release to atmosphere from a water storage. *Atmos. Environ.* **2011**, 45 (39),
346 7166–7173.
- 347 (33) Crawford, J.; Loken, L.; Casson, N.; Smith, C.; Stone, A.; Winslow, L. High-Speed
348 limnology: Using advanced sensors to investigate spatial variability in biogeochemistry and
349 hydrology. *Environ. Sci. Technol.* **2015**, 49, 442-450.
- 350