1 Supporting Information

| 2 | Long term in situ reduction in soil lead bioavailability measured in a |
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| 3 | mouse model |
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22 Table of contents

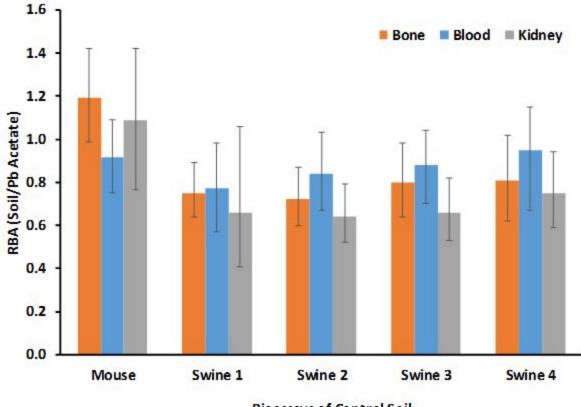
| 23 | Origin, collection | , and processir | ng of soil samples | 3 | | | Pa | age S2 |
|----------|--|-----------------|--------------------|---------|----------------|----------------|---------|---------|
| 24 | Comparison of soil Pb RBA estimates in mouse and juvenile swinePage S3 | | | | | age S3 | | |
| 25 | Figure S1: Pb RBA estimates for untreated soil in mouse and juvenile swine models Page S3 | | | | | | age S5 | |
| 26 | Figure S2: Pb RBA estimates for PA-treated soil in mouse and juvenile swine models .Page S | | | | | age S6 | | |
| 27 | Lead SpeciationPage S | | | | | age S 7 | | |
| 28 29 | Figure S3: Linear combination fitting results for Pb speciation in diet and feces samplesPage S9 | | | | | feces | | |
| 30 31 | Figure S4: Spec Page S10 | ctra of Pb star | ndards utilized i | n LCF | process | | | |
| 32 | Elemental | | analysis | | of | | | soils |
| 33 | | | | | Page | S10 | | |
| 34 35 | Figure S5: | | • | of | untreated | and | treated | soils. |
| 36 | Anodic stripping | voltammetric d | letermination of b | olood P | 'b | | Ра | age S11 |
| 37 38 | Table S1: QC Da S12 | ata for an anod | ic stripping volta | ammeti | ry blood Pb re | sults | | Page |
| 39 | References | | | | | | Pa | age 13 |
| 40 | | | | | | | | |
| 41 | | | | | | | | |
| 42 | | | | | | | | |
| 43 | | | | | | | | |
| 44 | | | | | | | | |
| 45 | | | | | | | | |

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- 49
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Origin, collection, and processing of soil samples - Soil samples used in this work were obtained 52 from field studies that evaluated the effectiveness and persistence of soil amendment treatments to 53 reduce Pb bioavailability. These field studies were conducted under the aegis of the Remediation 54 Technology Development Forum (RTDF), an organization created by the U.S. Environmental 55 Protection Agency (US EPA) in 1992 to foster a public-private sector collaboration to develop and 56 improve technologies for safe and cost-effective solutions to common types of environmental 57 contamination. Under the RTDF mandate, the US EPA National Risk Management Research 58 Laboratory and DuPont Co. formed the In-Place Inactivation and Natural Ecological Restoration 59 Technologies (IINERT) Soil-Metals Action Team in 1995. The IINERT Soil-Metals Action Team 60 included federal, state, and private organizations with a common interest in development and 61 62 validation of *in situ* techniques to reduce hazards posed by metals in soils. The team's efforts culminated in a field trial of various soil remediation procedures at the Joplin, Missouri, site. The 63 64 field study included a completely randomized design with four replicates of each treatment, 65 including control plots. A high-density polyethylene membrane was placed around the perimeter of each of the 2 x 4-m plots to reduce the potential of interplot contamination. In order to assure 66 that soils could not move between plots, plots were surrounded with heavy polyethylene sheet 67 68 (landfill liner grade) pre-installed to 30 cm in the soil and 10 cm above the soil. Amendments were

applied on the soil surface by hand to obtain even coverage. Composite samples from each of the 69 four replications were mixed and sieved. For these studies, phosphoric acid-treated plots were 70 installed by Dr. John Yang and Mr. David E. Mosby of the University of Missouri. Other plots 71 were installed by Drs. Sally L. Brown, Carrie E. Green, and William R. Berti. Amendments were 72 repeatedly rototilled with a tractor operated rototiller mixing the surface to a 0-15 cm depth. At 73 74 one month after initial treatment, all plots were treated with calcium hydroxide to raise soil pH, again rototilled repeatedly for thorough mixing, and seeded with tall fescue grass. Dr. Eton E. 75 Codling assisted in collecting the soils after 3 years for a feeding trial in juvenile swine. In 2013, 76 77 Dr. Green and Ms. Amy Poet assisted RLC in collecting soils for evaluation in the mouse model. Composite samples were prepared for each treatment by pooling together 100 g portions of 78 soil collected from each of four replicate plots. Pooled soil samples were sieved to 250 µm and 79 80 stored at room conditions until analyzed.

Comparison of soil Pb RBA estimates in mouse and juvenile swine - The Superfund site in Joplin, Missouri, and the soil remediation plan have been described.^{1,2} Estimates of Pb RBA for the untreated and PA-treated soil obtained in the juvenile swine have been reported.³ Figure S1 compares the Pb RBA estimates for untreated soil that were obtained in the mouse and the juvenile swine model. Figure S2 shows estimates of the Pb RBA obtained in mouse using soil collected about 3 years after application of phosphoric acid and in juvenile swine using soil collected 3 to 78 months after application of phosphoric acid.



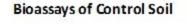


Figure S1 – Pb RBA estimates for untreated soil in mouse and juvenile swine models. Juvenile 89 90 swine assays of Pb RBA for untreated soil taken from unpublished reports describing results of 4 independent trials. Soil tested in Swine 1 and 2 collected less than one year after 91 remediation. Soils tested in Swine 3 collected less than two years after remediation. Soil 92 tested in Swine 4 collected about seven years after remediation. Mouse assays of Pb RBA for 93 untreated soil taken from a single multi-dose level assay. Error bars show 90% confidence 94 intervals on mean values. For both species, the bioavailability of the reference compound Pb 95 acetate was used in calculation of the RBA for the untreated soil. Each swine assay included 96 15 swine dosed with either soil or Pb acetate. Mouse assays included 27 mice dosed with 97 98 soil, 135 mice dosed with Pb acetate (multiple assays).

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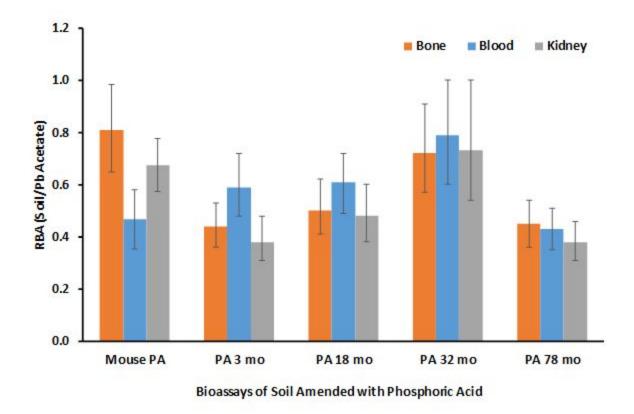




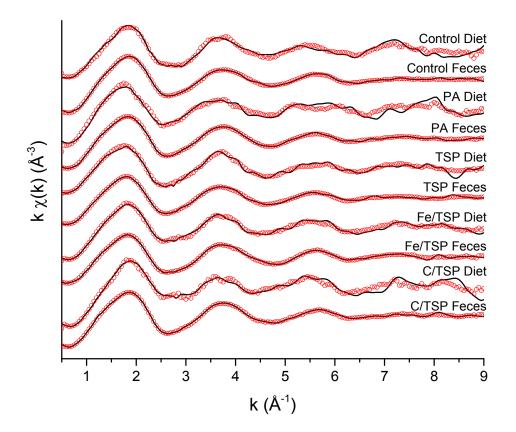
Figure S2 – Pb RBA estimates for PA-treated soil in mouse and juvenile swine models, Soils 101 assayed in juvenile swine were collected at 3, 18, 32 or 78 months after treatment. The soil 102 assayed in mice was a composite of soil samples collected about 3 years after treatment. 103 Error bars show 90% confidence intervals on mean values. For both species, the 104 bioavailability of the reference compound Pb acetate was used in calculation of the RBA for 105 the untreated soil. Each swine assay included 15 swine dosed with either soil or Pb acetate. 106 107 Mouse assays included 27 mice dosed with soil, 135 mice dosed with Pb acetate (multiple assays). R = rototilled. 108

Lead Speciation - X-ray absorption spectra were collected at the DuPont-Northwestern-Dow 110 Collaborative Access Team (DND-CAT) Sector 5, beam line 5BM-D, at the Advanced Photon 111 Source (APS) of the Argonne National Laboratory (ANL), U.S. The storage ring operated at 7 112 GeV in top-up mode. A water cooled double crystal Si(111) monochromator was used to select 113 the incident photon energies. The incident X-ray beam intensity was detuned to 60 % of its 114 115 maximum. Calibration was performed by assigning the first derivative inflection point of the absorption L_{III}-edge of Pb metal (13035 eV), and each sample scan was collected simultaneously 116 with a Pb metal foil. The samples were ground and pressed into pellets, affixed to a multiport 117 sample holder, and mounted for analysis without any further modifications. Data collection was 118 conducted in fluorescence with two Vortex-ME4 four-element silicon drift detectors and 119 transmission modes for the samples. For most samples, the transmission data were unusable for 120 analysis. Various Pb standards were used as reference spectra, including mineral sorbed Pb [Pb-121 ferrihydrite, Pb-kaolinite, Pb-goethite, Pb-gibbsite, Pb-birnessite, and Pb-montmorillonite in 122 which each mineral was equilibrated with Pb(NO₃)₂ at pH 6 for a target surface loading of 2500 123 mg kg⁻¹ after dialysis], organic bound Pb [Pb-fulvic acid and Pb-humic acid as reagent grade 124 organic acids equilibrated with Pb(NO₃)₂ at pH 6 for a target loading of 1500 mg kg⁻¹ after dialysis, 125 126 and reagent grade Pb acetate, Pb cysteine, and Pb citrate], Pb carbonate [Smithsonian Natural History Minerals Collection specimens of cerussite, hydrocerussite, and plumbonacrite with X-ray 127 128 diffraction verification], PbO [massicot and litharge], Pb-phosphates (chloropyromorphite, 129 hydroxypyromorphite, tertiary Pb phosphate $(Pb_3(PO_4)_2)$, PbHPO₄, and Pb sorbed to apatite at pH 6 and surface loading of 2000 mg kg⁻¹], and other Pb minerals [leadhillite, magnetoplumbite, 130 131 plumboferrite, plumbogummite, plumbojarosite, anglesite, and galena from the Smithsonian 132 Natural History Minerals Collection with X-ray diffraction verification]. All reference spectra were collected in transmission mode with dilution calculations determined by XAFSMass mixed
 in binder and pressed into a pellet.⁴

All sample and standard spectra were calibrated to a Pb foil on the same energy grid, averaged, 135 and normalized, and the background was removed by spline fitting using IFEFFIT.⁵ Principal 136 components analyses were performed in Sixpack on the k-weighted χ functions of the normalized 137 scans, and target factor analyses of each Pb standard were performed to determine the most 138 appropriate standards to be used for linear combination fits (LCF) analyses.⁶ Pb standards with 139 140 SPOIL values <3.0 were used in the LCF analyses, which included mineral sorbed Pb [sum of Pb-141 ferrihydrite, Pb-goethite, and Pb-birnessite], organic bound Pb [sum of Pb-fulvic acid and Pbhumic acid], Pb carbonate [sum of cerussite and hydrocerussite], PbO [sum of massicot and 142 litharge], Pb-phosphates [chloropyromorphite, hydroxypyromorphite, Pb₃(PO₄)₂, and Pb sorbed to 143 apatite], and other Pb minerals [plumbojarosite, anglesite, and galena]. The k-weighted γ functions 144 of the standards and samples were used for all linear combination fitting. Levenberg-Marguardt 145 least squares algorithm was applied to a fit range of 0.6 to 9.0 Å⁻¹. Best-fit scenarios, defined as 146 having the smallest residual error, also had sums of all fractions close to 1. To fully describe any 147 particular sample within 1% reproducible error, a minimum of two components was necessary, 148 and results have a $\pm 10\%$ accuracy. 149

Figure S3 and S4 show linear combination fitting (LCF) results for Pb speciation in diet and feces samples and for Pb standards, respectively.

S8



152 153 Figure S3 – Linear combination fitting (LCF) results for Pb speciation in diet and feces

samples.. 154

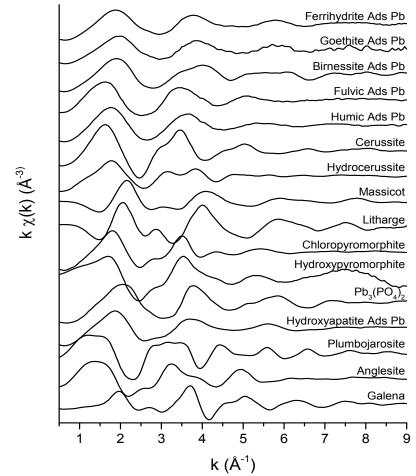
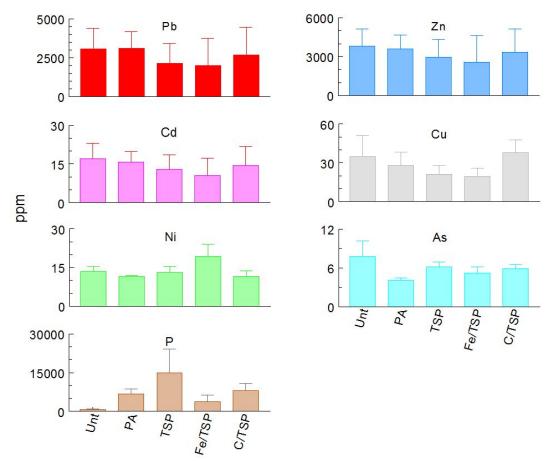


Figure S4 – Spectra for Pb standards utilized in LCF process

Elemental analysis of soils – Composited soil samples used for the studies reported here were prepared by aqua regia digestion and were analyzed by inductively coupled plasma-atomic emission spectrometry. Supplemental Figure S5 shows the results of these analyses.



162

Figure S5. Elemental composition of untreated and treated soils. Concentrations of elements shown in parts per million (ppm). Samples prepared by aqua regia digestion and analyzed by inductively coupled plasma-atomic emission spectrometry. Treatments are phosphoric acid (PA), triple super phosphate (TSP) alone or combined with iron oxide (Fe) or biosolids compost (C). Error bar show 95% confidence limits for mean values.

Anodic stripping voltammetric determination of blood Pb – Pb concentrations were
determined by anodic stripping voltammetry (ASV) using of NIST SRM 955c (Toxic Metals in
Caprine Blood) and blood form mice fed control diet that as spiked with Pb at three concentrations.
Table S1 summarizes the performance of ASV for Pb determination in blood.

- **Table S1.** Expected and observed Pb levels and recoveries for NIST SRM 955C (Toxic Metals in
- 174 Caprine Blood) and blood collected from mice fed control diet that spiked with Pb at three different

| 175 concentrations, as determined by anodic stripping voltammetry | 175 | concentrations, as determined b | y anodic | stripping | voltammetry. |
|---|-----|---------------------------------|----------|-----------|--------------|
|---|-----|---------------------------------|----------|-----------|--------------|

| Analyte | NIST Certified Value (µg/dL) | Observed Values (µg/dL) (n=3) | % Recovery |
|--------------------------|---------------------------------|----------------------------------|---------------------|
| NIST SRM 955C Level 2 | 13.95 | 14.9, 13.8, 13.2 | 106.8, 98.9, 94.6 |
| Pb-spiked mouse blood | Target value (µg/dL) | Observed Values (µg/dL) (n=3) | % Recovery |
| Low | 2.8 | 2.9, 3.0, 2.8 | 103.6, 107.1, 100.0 |
| Medium | 13.5 | 13.8,12.9, 12.2 | 102.2, 95.6, 90.4 |
| High | 54.0 | 55.2, 54.8, 51.3 | 102.2, 101.5, 95.0 |

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