

Supplementary Materials

**Decoupling livestock from land use through industrial feed
production pathways**

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45 - Pages

15 - Figures

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30 **Materials and methods**

31

32 **General description of the Model of Agricultural Production and its Impact on the**

33 **Environment (MAgPIE)**

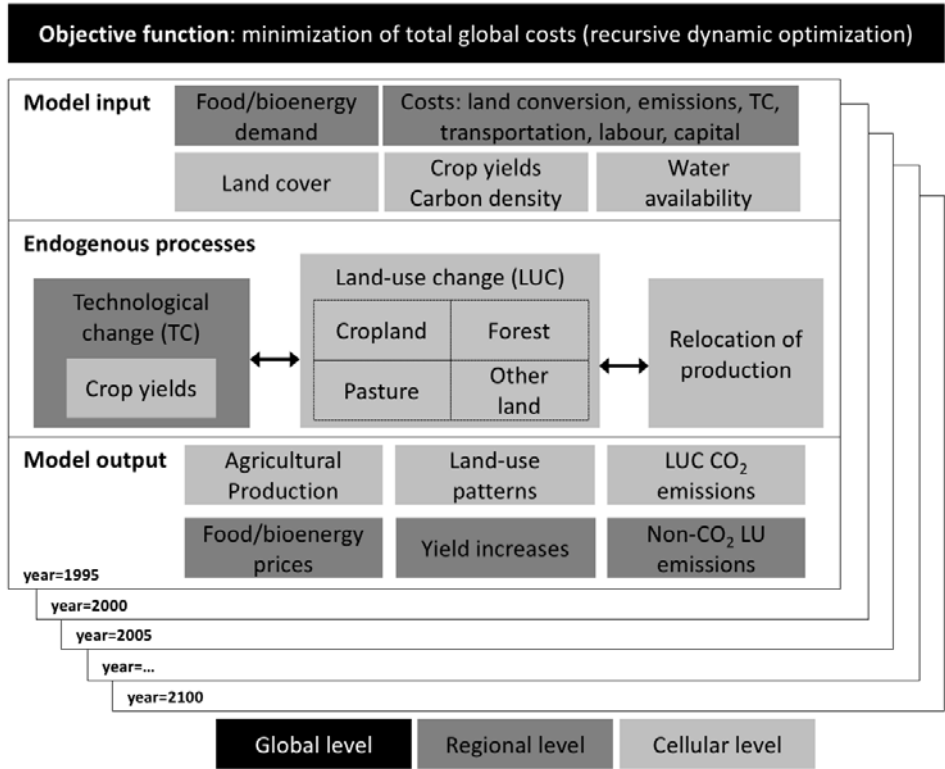
34 The MAgPIE model is a mathematical programming model that is used to create long-term

35 scenarios and assessments of global agriculture and land use.¹⁻⁷ It integrates several spatial scales

36 (see Figure S1 for an overview of the model structure) and simulates the full food supply chain

37 from nutrient, land and water use over crop and livestock production, food processing up to final

38 consumption.



39 **Figure S1.** Schematic representation of the MAgPIE model.

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42 For a given food demand (see Fig. S3 the model estimates cost-optimal production patterns, and

43 simulates major dynamics of the agricultural sector, like trade, technological progress and land

44 allocation according to the scarcity of suitable land, water and economic resources.⁶

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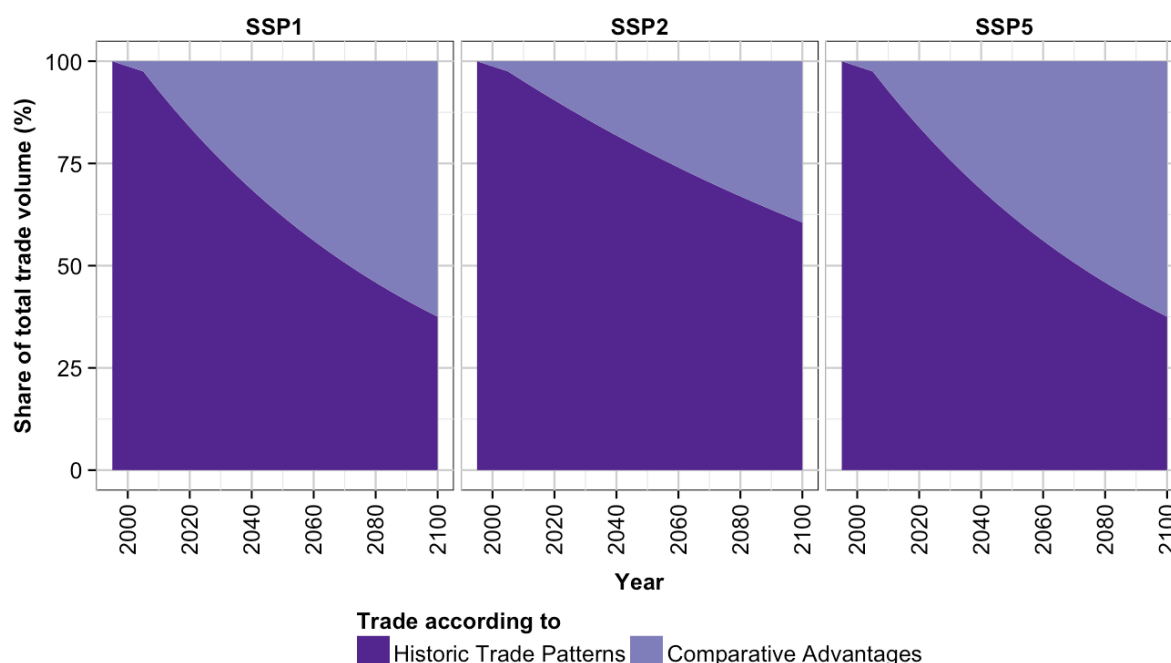
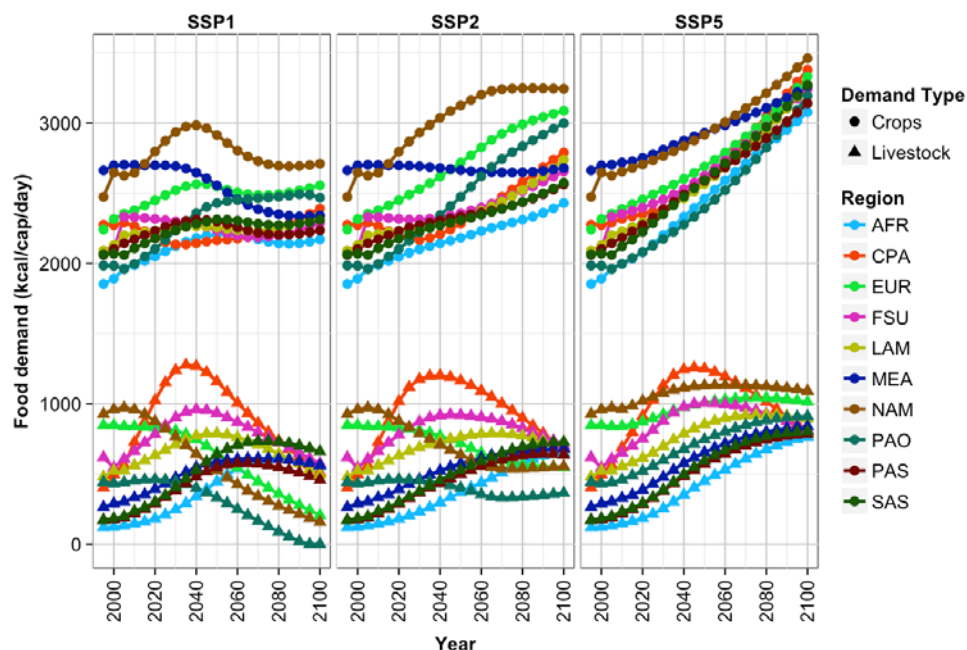


Figure S2. Agricultural Trade scenarios based on Schmitz et al., (2012).⁸ There are two trade pools in the MAgPIE model, one with trade according to historic trade patterns, another one with free trade according to comparative advantages. In SSP1 and SSP5 barriers for international trade decline by 1% per year (Policy scenario as described in Schmitz et al., (2012), reflecting globalized agricultural markets in these scenarios. In SSP2, trade barriers decline only by 0.5% per year, reflecting rather regionalized trade of agricultural goods.

As it treats agricultural products not only as economic values but also as physical goods, MAgPIE can perform analysis of material flows. The products in the model comprise 17 crop groups, each with individual above – and belowground crop residues, 5 livestock production types (i.e. beef cattle, dairy cattle, pigs, broiler and laying hen), 8 types of conversion by-products originating from food processing, as well as grazed pasture and scavenging. Products can be used for food, feed, other use (comprising material use and waste in the production chain) and seed, where applicable. Crop residues can be recycled to soils, burned in the fields, used as feed, or used as material.² The demand for food enters the model as an exogenous trajectory.^{2,9}



63

64 **Figure S3.** Time-series of regional per capita food demand in SSP1, SSP2 and SSP5. The values
 65 are estimated using the methodology described in Bodirsky et al ¹⁰ based on SSP population and
 66 GDP projections. SSP1 and SSP2 use a different demand system than SSP5, leading to higher
 67 demand for the latter. Additionally, it was assumed that per-capita calories in SSP1 fall to a
 68 maximum of 3000 kcal per capita per day in 2100, and that the share of livestock products in rich
 69 countries does not fall below 15% in SSP2 and SSP5.

70

71 Demand for material consumption and production waste is assumed to grow over time in
 72 proportion to food demand, while the demand for seed is a fixed share of crop production. ² The
 73 demand for feed depends on the quantity of livestock production, as well as regional and livestock
 74 system-specific feed baskets. ¹¹ Supply of livestock products (ruminant meat, whole-milk, pork,
 75 poultry meat and eggs) is provided by five animal food systems (beef cattle, dairy cattle, pigs,
 76 broiler and laying hen). The regional and system-specific feed baskets were derived as follows: In
 77 a first step, regional feed energy requirements were calculated based on specific feed energy
 78 requirements per product ¹² for each animal food system and animal group (i.e. reproducers,
 79 producers and replacing animals) that entail the minimum requirements for maintenance, growth,
 80 lactation, reproduction and other basic biological functions of the animals. In addition, they

comprise a general allowance for basic activity, temperature effects and expenditures for grazing. In a second step, these regional feed energy requirements are settled by regional feed energy supply comprising concentrate feed, conversion by-products (both in accordance with the Commodity Balance Sheets (CBS) from ¹³), crop residues ², and grazed biomass. As there are little data available on the amount of grazed biomass, grazed biomass was assumed to provide the remainder to fulfill the energy requirements of ruminants. The feed was distributed between the animal groups based on constraints regarding the nutrient density of resulting feed baskets. Future development of livestock productivity is driven by exogenous projections (Figure S4), and determines feed energy requirements and feed basket composition, leading to a higher share of concentrates with higher yields.

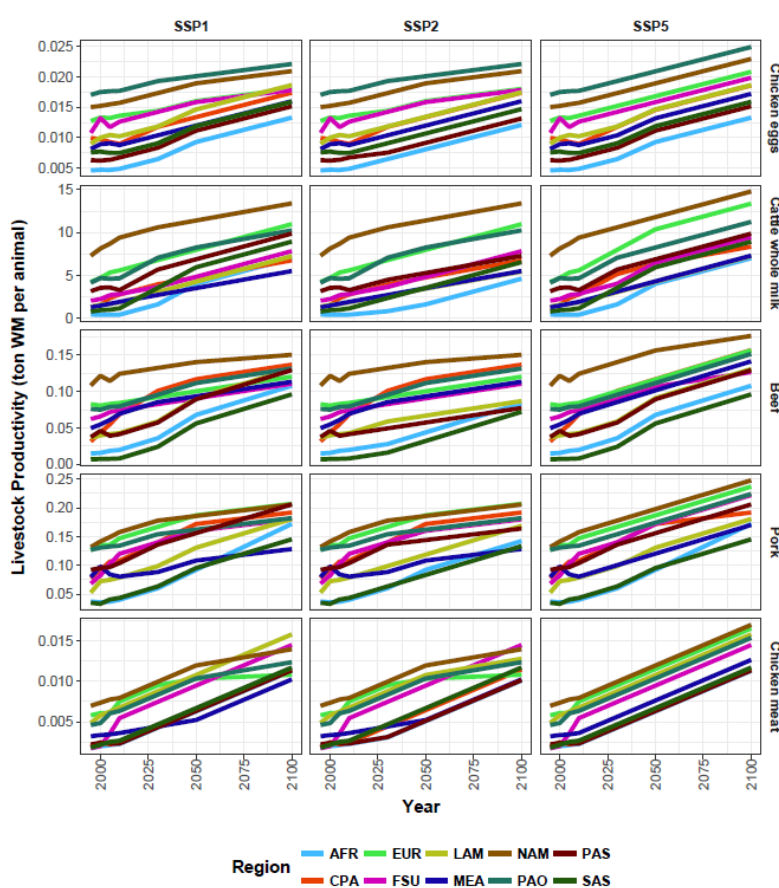


Figure S4. Time-series of regional livestock productivity in SSP1, SSP2 and SSP5. Future scenarios of livestock productivity are derived based on informed guesses, taking into account past productivity improvements, GDP projections, cultural particularities, and the general SSP

story line. Livestock productivity scenarios determine feed energy requirements and feed basket composition.

All demand categories are calculated on the level of ten world regions (Figure S5), and can be settled through domestic production or international trade.

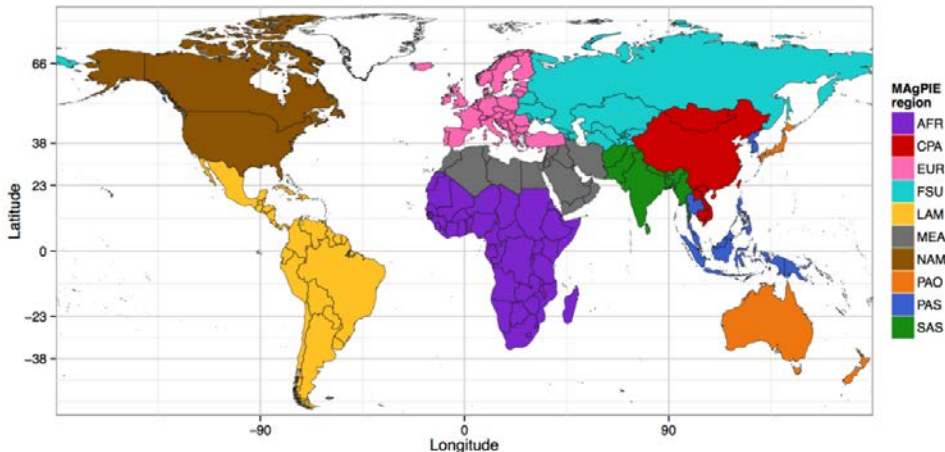


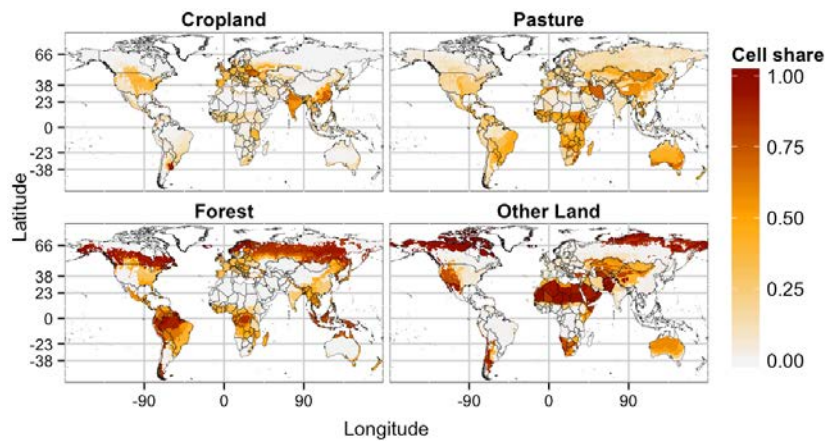
Figure S5. Map of economic world regions in the MAgPIE model.

To account for trade barriers, the regions have to produce a certain share of their demand domestically.⁸ The production of crops requires financial resources as well as land, water and fertilizer.¹⁴ Also, spatial production patterns have to account for good-practice crop rotations, which are enforced by crop-specific constraints. Cropland expansion leads to additional costs and is limited by biophysical constraints as well as by competing land use activities. Additionally to land-expansion, the model can also invest into yield-increasing research and technology.¹⁵ Crop growth functions connect crop harvest to the production of above – and belowground residues.² Similarly, the production of conversion by-products depends upon fixed conversion factors multiplied with the regional crop supply.² Finally, livestock production requires financial resources, feed and water. For meeting the demand, the model endogenously decides, based on cost-effectiveness, about the level of intensification (i.e. yield-increasing technological change), land expansion (i.e. land-use change) and production relocation (i.e. intra-regionally and inter-

115 regionally through international trade).^{8, 15-17} The decision criterion is cost minimization. Major
 116 cost types in MAgPIE are factor requirement costs (e.g. capital, labour and fertilizer), land
 117 conversion costs, transportation costs to the closest market and investment costs for yield-
 118 increasing technological change (TC). The model is solved in a recursive dynamic mode with a
 119 time step length of five years on a timescale from 1995 to 2100.

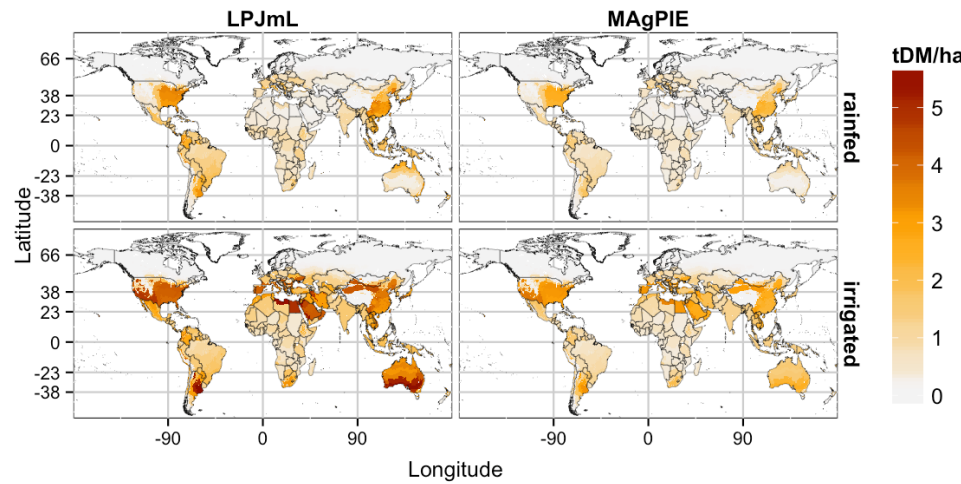
120 Biophysical information on initial land-cover, crop yields, carbon densities and water availability
 121 is spatially explicit (0.5 degree longitude/latitude). Due to computational constraints, spatially
 122 explicit data is aggregated to 200 simulation units for the optimization process. For each of the ten
 123 world regions in MAgPIE, the clustering algorithm combines grid cells to simulation units based
 124 on the similarity of data.¹⁸ Land types in MAgPIE comprise cropland, pasture, forest, other land
 125 (e.g. non-forest natural vegetation, abandoned agricultural land, deserts) and urban land (static
 126 over time).¹⁹ In 2005, the global land area (12907 Mha in total) consists of 1471 Mha cropland,
 127 3129 Mha pasture, 4208 Mha forest and 4135 Mha other land (e.g. urban areas) (Fig. S6,
 128 Supplementary Materials excel spreadsheet).

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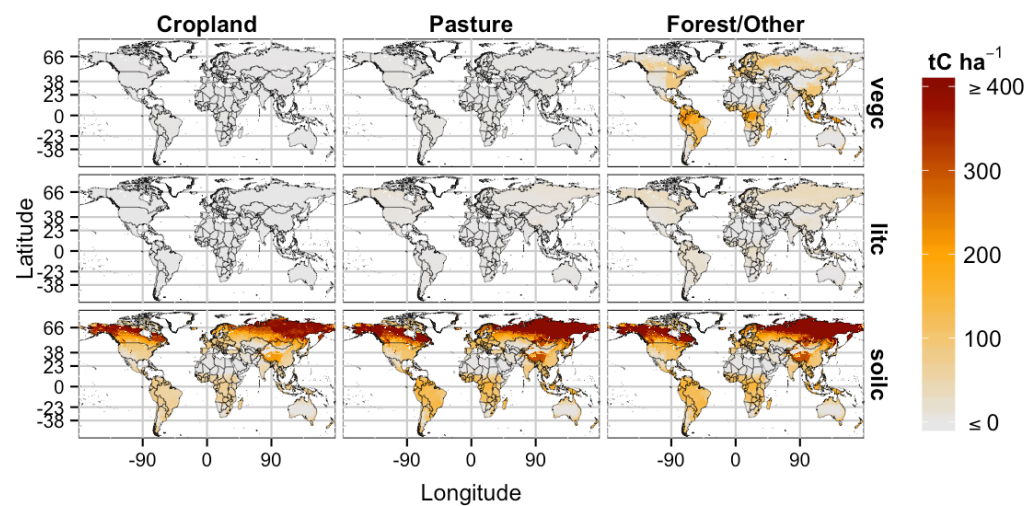


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 131 **Figure S6.** Spatially explicit land use / land cover in MAgPIE in 2005. For each grid cell, the
 132 cell shares of the four land types add up to 1.
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135 Cropland covers cultivation of 16 food/feed crop types (e.g. temperate and tropical cereals, maize,
 136 rice, oilseeds, roots), both rainfed and irrigated systems, and two 2nd generation bioenergy crop
 137 types (i.e. grassy and woody). Biophysical yields for these crop types as well as carbon densities
 138 of natural vegetation and water availability for irrigation are derived by a dynamic global
 139 vegetation, hydrology and crop growth model, the Lund-Potsdam-Jena model for managed Land
 140 (LPJmL) ^{20, 21}, and calibrated to observed yields from. ¹³



141
 142 **Figure S7.** Tropical cereal yields in rain fed and irrigated production systems as derived by
 143 LPJmL, and how these yields are adjusted before entering the MAgPIE model.



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 145 **Figure S8.** Spatially explicit carbon density (tC ha^{-1}) based on LPJmL for four land types and
 146 the three carbon pools vegetation, litter and soil (veg, litc, soil).
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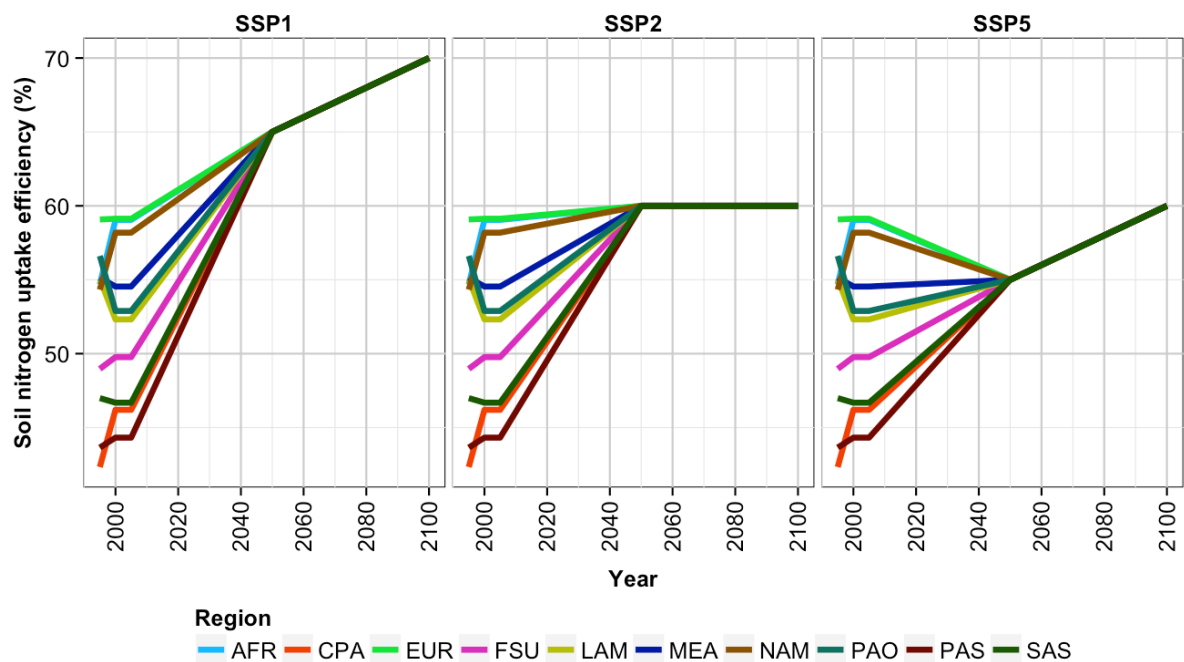
148 LPJmL simulations of crop yields assume that all crops are grown in all grid cells to assess the
 149 possible crop productivity also in areas currently not under cultivation to inform possible shifts in
 150 cropping areas. In seven individual LPJmL runs, crop yields are derived for seven different
 151 management intensity levels. LPJmL represents potential crop yields, while MAgPIE aims to
 152 represent actual crop yields. Due to endogenous investments in yield-increasing technological
 153 change in the MAgPIE model, MAgPIE yields can exceed LPJmL yields in future time steps. ¹⁵
 154 MAgPIE was extended by a Nr flow module ² that transforms all biomass flows in the model into
 155 reactive nitrogen (Nr) flows. The closed budget approach and a consistent connection of all Nr
 156 flows guarantees that the sum of Nr fixation, Nr release and inflows from other sectors corresponds
 157 to the sum of Nr losses and Nr flows to other sectors. Similarly, intermediary closed budgets are
 158 also used on a regional level for cropland soils, distribution and processing of agricultural
 159 products, livestock feeding, manure management, and the household sector. While certain
 160 processes in MAgPIE like crop production and water use are simulated spatially explicit on the
 161 level of clustered 0.5° grid cells, Nr flows and Nr budgets are calculated on the level of 10 world
 162 regions. ²
 163 Nr in harvested crops (*H*), aboveground (*RA*) and belowground (*RB*) crop residues are estimated
 164 based on crop-specific crop growth functions and Nr contents of the individual plant components.
 165 The Nr derived from biological fixation within plants (*B1*) is estimated based on plant-specific
 166 shares of plant Nr derived from atmospheric fixation, while biological fixation by free-living
 167 microorganisms (*B2*) is estimated based on typical rates of Nr fixation per cultivated area. All
 168 belowground crop residues are assumed to be recycled to the field (*RRB*). In contrast, aboveground
 169 crop residues have to settle first the demand for feed from the livestock sector; moreover, a fixed
 170 share of crop residues is assumed to be used for material purposes or is lost when residues are
 171 burned on the field. Only the remaining residues are recycled on fields (*RRA*). The Nr which is
 172 released on croplands when soil organic matter depletes after the opening of new cropland (*SOM*)

173 is calculated on the basis of model-endogenous land-use change activity, spatial-explicit carbon
174 contents of natural soils, typical climate-specific shares of lost soil carbon under cropland
175 management, and a fixed carbon to nitrogen (C:N) ratio in soils. Atmospheric deposition (*D*) is
176 estimated based on current deposition rates and is assumed to grow proportional to volatilized Nr
177 losses. The Nr input of seeds (*S*) to cropland soils is based on regional plant-specific shares of the
178 production that are used for seed. ²

179 To estimate the amount of manure recycled to croplands (*RM*), a budget approach is used. The Nr
180 in animal feed (*A*) is estimated by assigning Nr contents to all feed items. The Nr in livestock
181 products (*FA*) and slaughter waste (*SW*) is subtracted from the Nr in animal feed (*A*) to derive the
182 amount of excreted Nr. While the feed from pasture is assumed to be excreted back on pastureland
183 (*RPI*), other excrements are distributed between different animal waste management systems
184 according to regional shares that change over time with scenario assumptions. Depending on their
185 management, different shares of the nutrients are lost to volatilisation and denitrification (*LA*); the
186 remainder is recycled as organic fertilizers on cropland (*RM*) and pasture (*RP2*). ^{1, 2}

187 Central to the model are regional nutrient budgets for cropland soils, which are used to determine
188 the future requirements of fertilizers. Different Nr inputs to croplands have different fertilizer
189 equivalents, as they are to a different degree subject to losses on the field. Homogeneous Nr inputs
190 (e.g. Haber-Bosch fertilizer) that can be easily distributed over the field have high equivalence
191 values, while inhomogeneous Nr inputs, whose quantity, spatial distribution and timing cannot be
192 controlled (e.g. atmospheric deposition) have low equivalence values. Using a simplified
193 approach, we distinguish only two categories in MAgPIE: Firstly, Nr inputs from seed (*S*) and
194 biological fixation within the plant (*BI*) are not taken up from the soil, and hence are not subject
195 to losses by volatilisation, leaching or denitrification. Nr from these inputs is therefore assumed
196 to be fully incorporated into plant tissue. Second, we assume that all other Nr inputs are subject to
197 losses prior to plant uptake and that only a share of the applied Nr inputs are withdrawn by the

198 plant roots and incorporated into plant biomass. This share, named as soil Nr uptake efficiency
 199 (*SNUPE*), is a regional scenario parameter which reflects next to the climatic and biophysical
 200 conditions mostly the fertilization technology of farmers (Fig. S9).



201 **Figure S9.** Time-series of regional soil nitrogen uptake efficiency in SSP1, SSP2 and SSP5, based
 202 on ¹. The soil nitrogen uptake efficiency reflects the fertilization technology of farmers. Thus, the
 203 soil nitrogen uptake efficiency determines, together with the fertilization requirements for the
 204 production of crops, the application of inorganic nitrogen fertilizer. First, all available organic
 205 nitrogen fertilizer is used to fulfil the fertilization requirements. Subsequently, the remainder needs
 206 to be balanced out by the application of inorganic nitrogen fertilizer. Accordingly, the soil nitrogen
 207 uptake efficiency influences nitrogen pollution and N₂O emissions from soils. As the challenges
 208 for climate change mitigation increase from SSP1 to SSP2 to SSP5, the soil nitrogen uptake
 209 efficiency is high in SSP1 (low challenges), medium in SSP2 (medium challenges) and low in
 210 SSP5 (high challenges).
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 213 As we do not consider the differences in fertilizer equivalences between individual soil Nr inputs
 214 like manure and inorganic fertilizers, a high value for *SNUPE* also implies that inhomogeneous
 215 fertilizers can be better integrated into crop fertilization, e.g. by improved monitoring or by manure
 216 processing.

217 To calculate the application of inorganic fertilizers (F), we estimate the fertilization requirements
 218 to obtain a certain production under a given fertilization technology level. The Nr withdrawals
 219 from crop production ($H+RA+RB$) and the exogenous soil Nr uptake efficiency ($SNUPE$)
 220 determine the required Nr inputs from organic or inorganic fertilizers:

221

$$222 \quad H + RA + RB = SNUPE * (F + RM + RRA + RRB + D + B2 + SOM) + 1 * (S + B1) \quad (\text{Eq. 1})$$

223

224 Based on the availability of organic fertilizers, it is ultimately determined which amount of
 225 inorganic fertilizers is needed to balance out the budget.² As the cropland budget is aggregated
 226 over crop-types, nutrient transfers within crop rotations are accounted for. The difference between
 227 inputs and withdrawals represents the field losses (LF) by denitrification, leaching and
 228 volatilisation:

229

$$230 \quad LF = F + RM + RRA + RRB + D + B2 + SOM + RH + RS + S + B1 - H - RA - RB \quad (\text{Eq. 2})$$

231

232 Agricultural nitrous oxide (N_2O) emissions from agricultural soils and manure management are
 233 estimated in MAgPIE with the emission factors of the IPCC guidelines for national greenhouse
 234 gas inventories²². Methane (CH_4) emissions from rice, enteric fermentation and manure
 235 management are estimated based on the methodology of IPCC.²³

236 CO_2 emissions from land-use change in MAgPIE reflect differences in carbon stocks between time
 237 steps. If, for instance, forest is converted to cropland within the same simulation unit, the carbon
 238 stock of this unit decreases according to the difference in carbon density of forest and cropland
 239 (Fig. S10). In case agricultural land is abandoned (other land pool), ecological succession leads to
 240 regrowth of natural vegetation carbon stocks along sigmoid growth curves³. Regrowth of carbon
 241 stocks in MAgPIE is constrained by the LPJmL natural vegetation carbon density. If the vegetation

carbon density of re-growing vegetation passes a threshold of 20 tC/ha, the respective area is shifted to the forest land pool.^{24, 25}

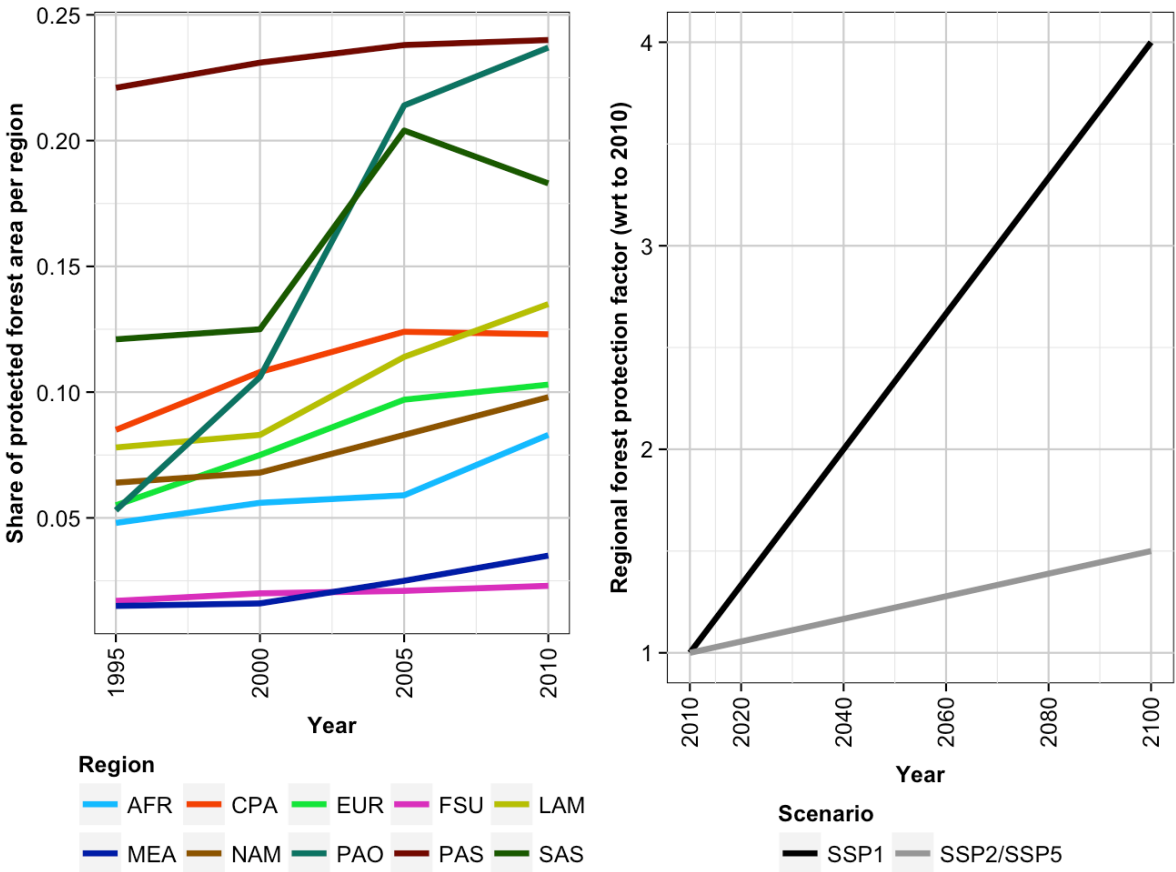


Figure S10. Left: Shares of protected forest for each region (in terms of regional forest area). Values are based on the global forest resources assessment report.²⁶ Right: Assumed increase of regional forest protection shares in SSP1, SSP2 and SSP5 with respect to 2010. Forest protection takes out land from the pool of available land for agricultural expansion. Thus, forest protection increases competition for land in MAgPIE.

Model results

Feed basket: The altered feed baskets for the five livestock categories, namely, beef cattle, dairy cattle, pigs, broiler chickens and laying hen are found in Figure S11-S15.

256 *Impacts on agricultural production, land-use dynamics and nitrogen flows:* A detailed resume of
257 the model results obtained for all 48 scenarios in terms of land use, nitrogen budget and greenhouse
258 gas emissions is available in the Supplementary Materials excel spreadsheet.

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260 **Economic assessment of the technological pathways for production of Microbial Protein**

261

262 *General assumptions*

263 To establish whether the production of MP can become economically competitive to crop-based
264 production of protein, the economic feasibility of five MP production pathways was investigated.
265 General assumptions for the process requirement are reported in Table S4.

266

267 *Process requirements and costs: capital investments and annuity*

268 The capital investments were calculated based on the annuity method assuming loan repayments
269 on an annual basis. The calculated annual repayment was subsequently normalized to the specific
270 costs per ton MP produced (\$/ton MP). The following assumptions were made:

- 271 • Capital investment of US\$75 million (US\$50 – US\$125 million);
- 272 • Equipment lifetime of 25 years;
- 273 • Repayment period of 25 years;
- 274 • Interest rate of 5% (3 – 7%);
- 275 • MP production of 25,000 ton MP per year (expressed in 100% dry weight).

276

277 Table S9 reports the annuity per ton MP produced above described assumptions (i.e. \$203/ton
278 MP). Shifting the required capital investment to US\$50, US\$100 and US\$125 million and interest
279 rate between 3 to 7%, would result in a minimum and maximum annuity of \$112 to \$401/ton MP,
280 respectively.

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Process requirements and operational costs: energy and carbon sources

Hydrogen

Table S5 shows the levelized costs per kg hydrogen produced as well as the hydrogen production costs per ton MP produced. An 80% hydrogen conversion into MP was assumed, with the other 20% presumed being lost due to gas transfer limitations. The reminder hydrogen could in theory be converted into heat and power using a Combined Heat Power (CHP) unit. The latter would provide sufficient energy for the drying step, an option that was not considered here however (table S9).

Methane

A low cost estimate (\$0.17 / Nm³) applicable to the USA market, and a high cost estimate (\$0.33 / Nm³), applicable to the European market were considered for natural gas requirements. An 80% methane conversion into MP was assumed similarly to hydrogen above and the potential to recover heat and power via a CHP unit was not considered (Table S9).

Raw sugar

Table S5 shows price of raw sugar per ton raw sugar purchased, whereas in Table S10 the costs are expressed as the cost per ton MP produced.

Other process requirements and related operational costs

Nitrogen and phosphate

305 The amount of nitrogen and phosphate required and their associated operational costs are
306 presented in table S6. The production of MP also requires the use of trace elements and micro-
307 nutrients and pH regulation dosing acid, which were found to be marginal and as such have been
308 implicitly included in the overhead costs (table S9).

309 310 Oxygen

311 Oxygen is required for microbial growth, independently of the carbon and energy sources. Table
312 S7 shows the estimated resources required to generate industry-grade pure oxygen; 80% oxygen
313 utilization efficiency was assumed while the reminder 20% was considered lost due to gas transfer
314 limitations, similar to the hydrogen and methane utilization rates.

315 316 Pumping and mixing

317 Fermentation in stirred tank reactor requires energy for gassing, mixing and pumping. Energy
318 requirements for gas fermentation are higher than for sugar substrate since gaseous substrate
319 necessitates strong mixing to facilitate gas-to-liquid transfer and obtain high volumetric
320 productivity. Pumping energy is comparable across substrates and an average mixing energy
321 between 0.2 – 3 kWh/m³, with 0.8 kWh/m³ considered in our assessment.²⁷ In our assessment, a
322 rather conservative value of 1.5 kWh/m³ for mixing and pumping was considered as energy
323 demand for all substrates.

324 325 Dewatering, drying and sterilization

326 After the fermentation step, produced undergoes heat-treatment to lyse cells and increase protein
327 accessibility, reduce nucleic acid content and obtain a dry sterilized product.²⁸ Prior to heat-
328 treatment, water is removed by centrifugation (or other conventional dewatering methods) leaving
329 around 25% dry solids content to reduce energy consumption for heat-treatment. Drying is
330 achieved via spray-drying with integrated fluidized bed, a common practice in the food processing

331 industry.²⁹ The operational costs of drying and sterilization are mainly related to the energy
332 consumption of the process (table S8).

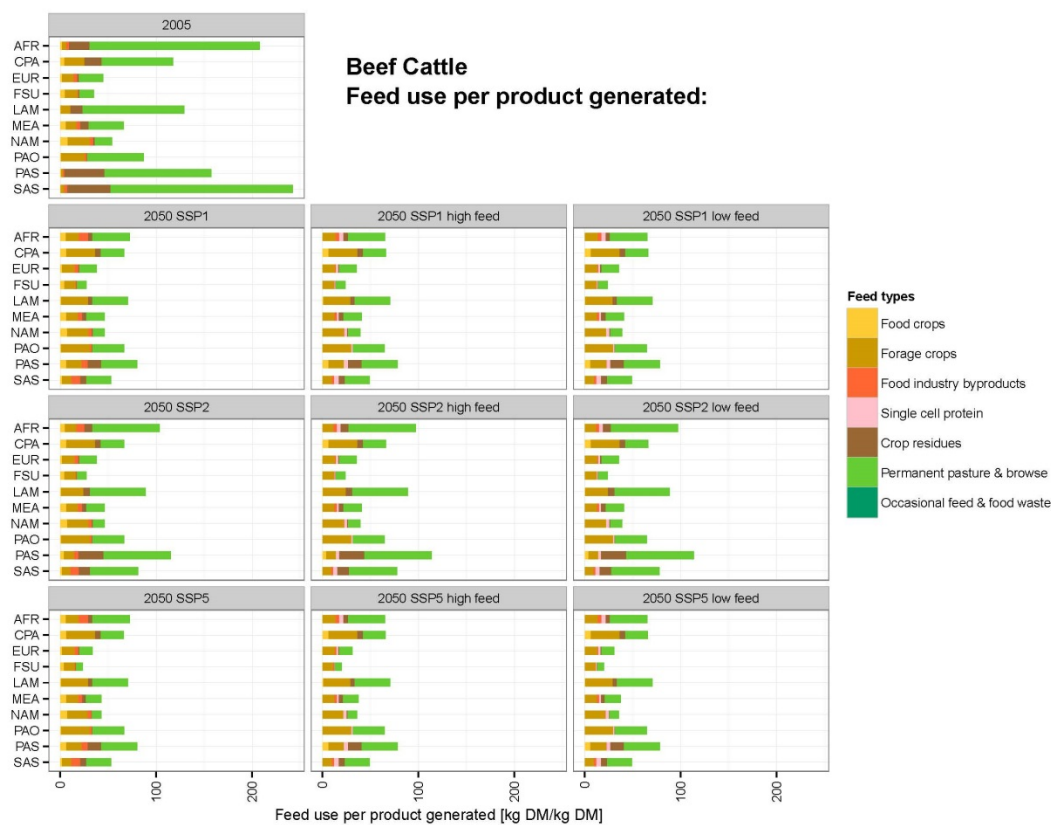
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334 Overhead

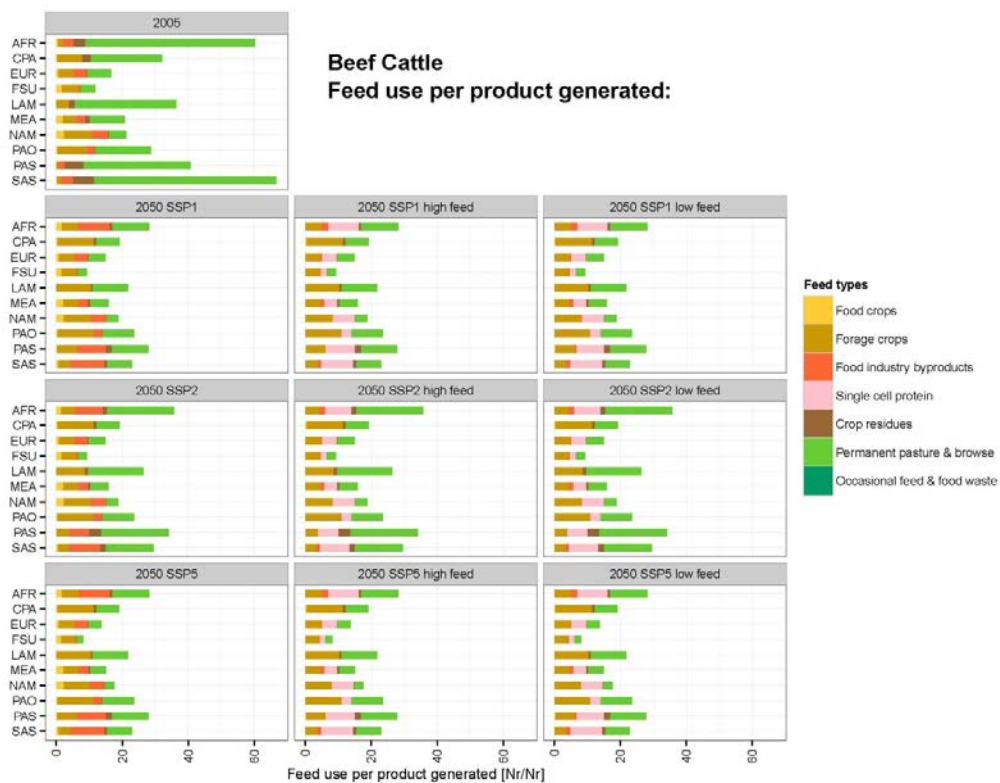
335 The overhead cost was assumed to be equal to the annuity of the production of 1 ton of MP,
336 corresponding to \$270 per year (i.e. ~15 – 20% of the total cost (except for hydrogen based
337 production through water electrolysis, i.e. 9%, Table S9), a conservative estimate. Overhead costs
338 include labor, maintenance, quality testing, packaging, depreciation and miscellaneous other
339 minor expenses. Note that the labor costs depend on the world's area where MP is produced as
340 well as on single countries governments (e.g. to support a particular industry with taxation,
341 assistance in the recruitment of highly skilled labor, etc.). Assessing the value of these differences
342 was beyond the scope of this paper and therefore a fixed (rather conservative value) was used.

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348 **Figure S11.** Feed basket for Beef Cattle. Top: Feed use per product generated expressed in kg
349 DM / kgDM feed

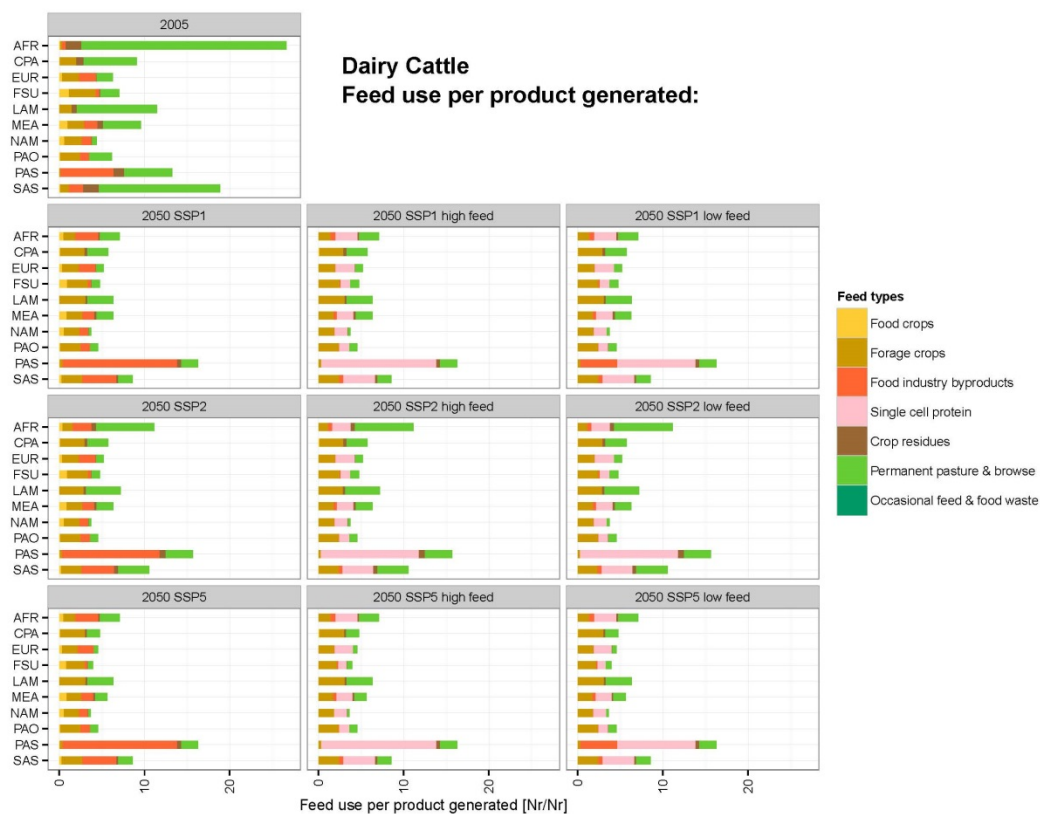
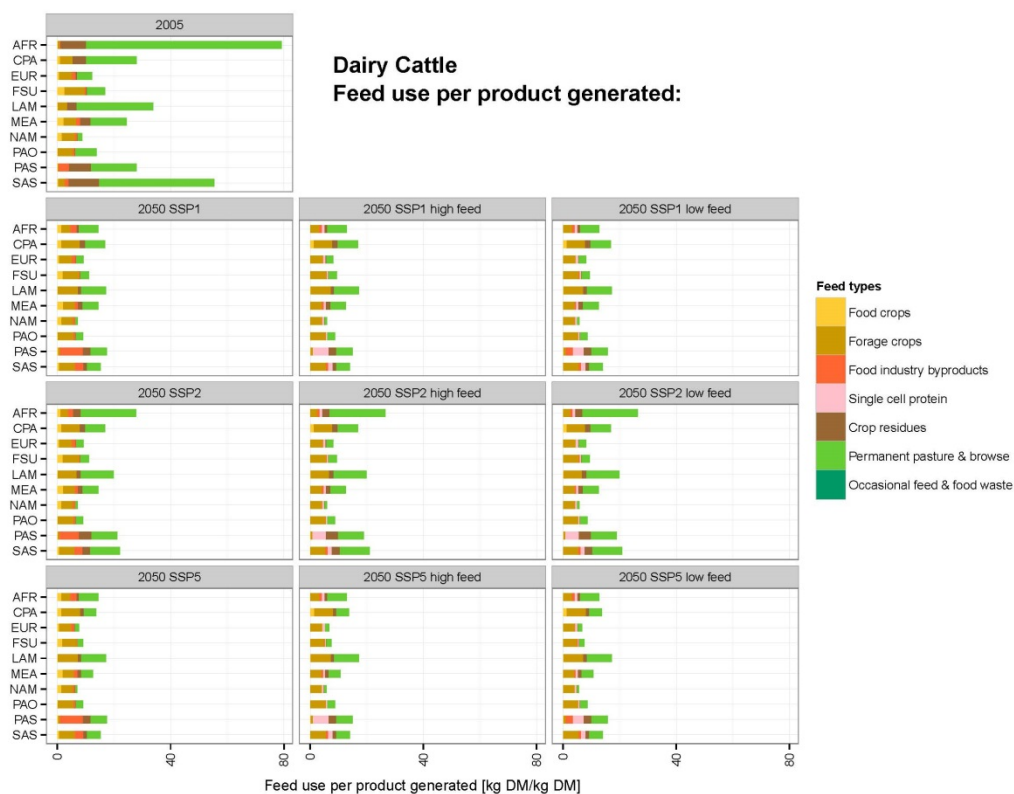


Figure S12. Feed basket for Dairy Cattle. Top: Feed use per product generated expressed in kg DM / kg DM feed

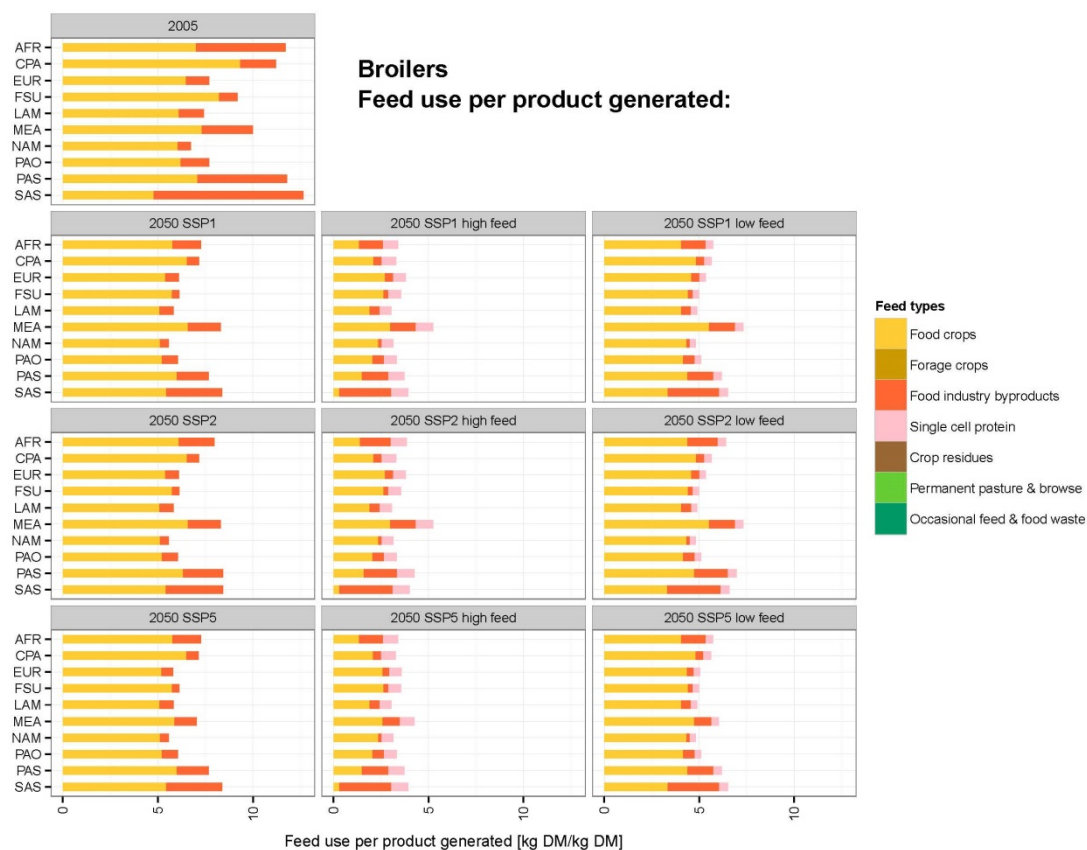
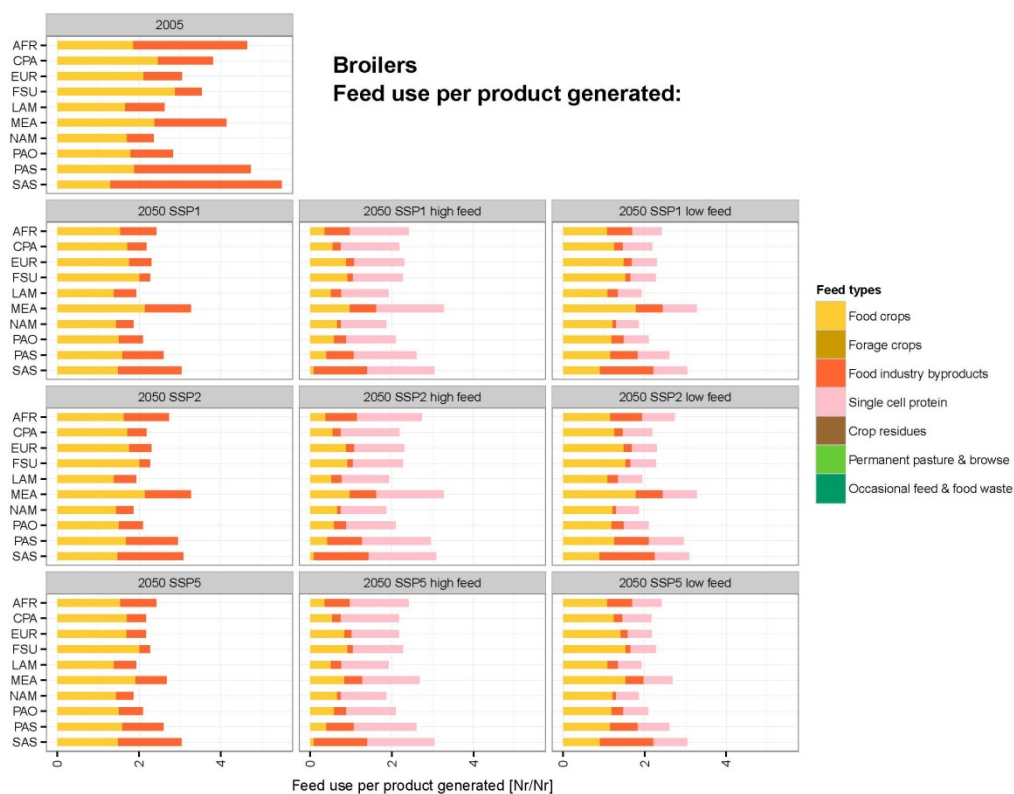


Figure S13. Feed basket for Broilers. Top: Feed use per product generated expressed in kg DM / kg DM feed

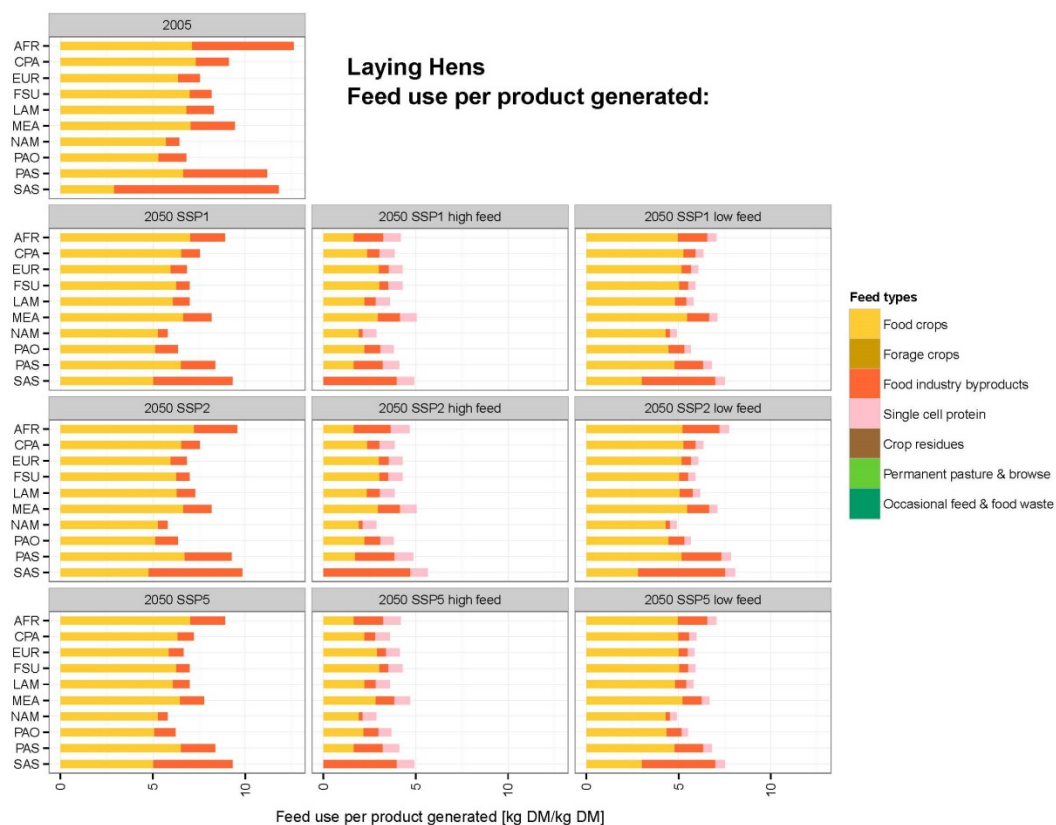
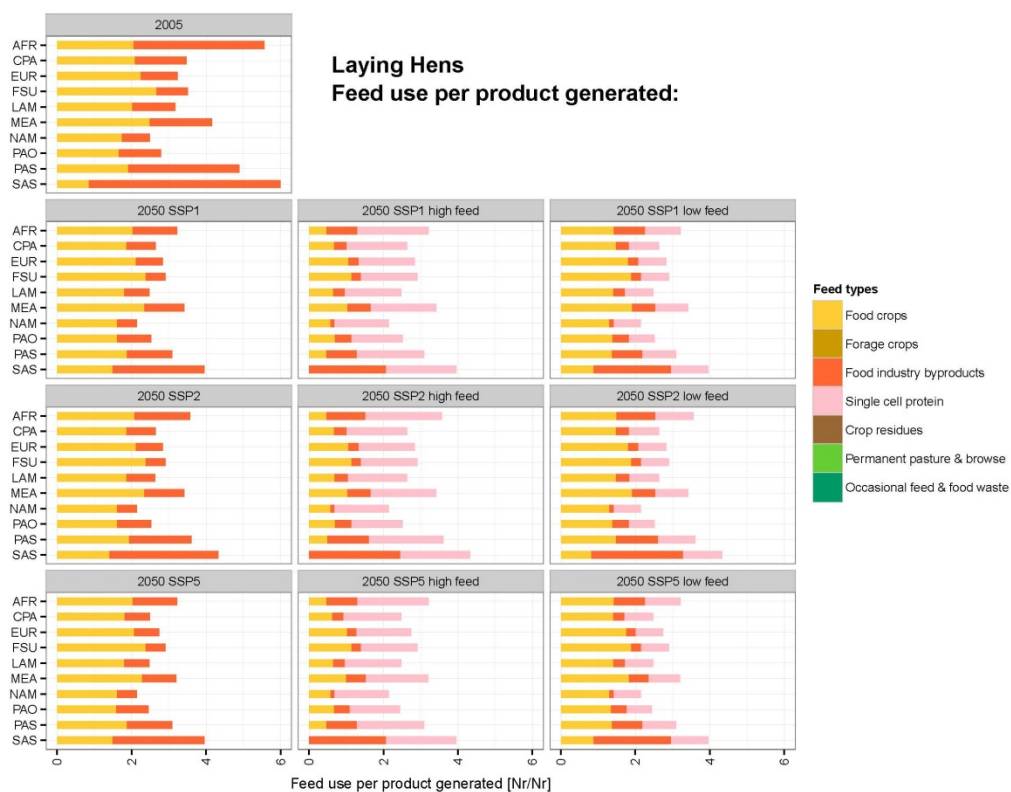
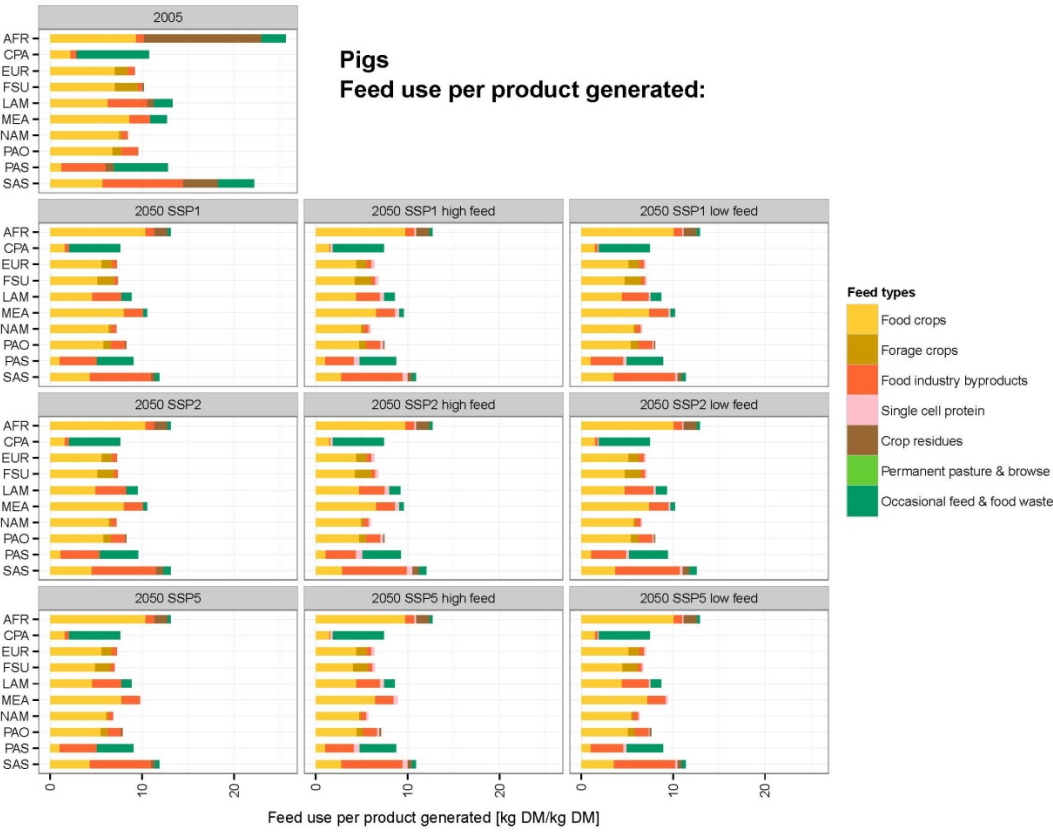
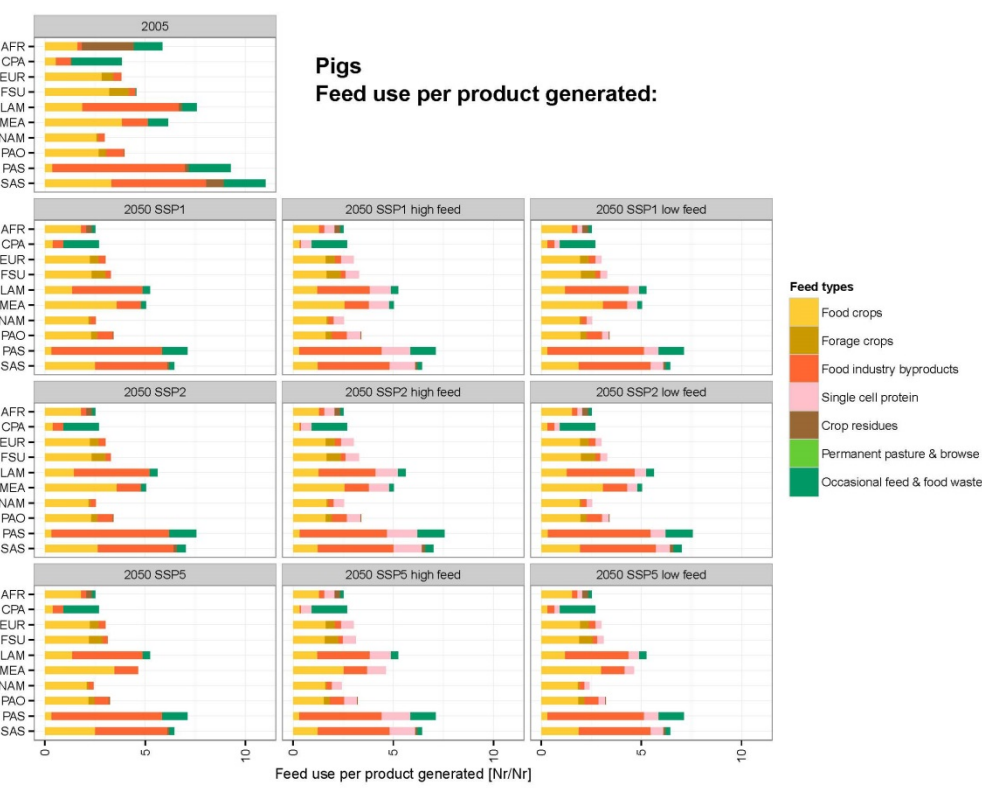


Figure S14. Feed basket for laying hens. Top: Feed use per product generated expressed in kg DM / kg DM feed

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367 **Figure S15.** Feed basket for Pigs. Top: Feed use per product generated expressed in kg DM /
368 kgDM feed

Table S1. Feed substitution potential of MP within livestock feed baskets.

Type of livestock	MP replacement in feed baskets	refs	Remarks
Pigs	21% (dry matter basis)	³⁰	In this study, feeding trials using 24 pigs (with a mean body weight 44 kg) were used to determine the digestibility of energy, nitrogen and amino acids. MP was added to the basal diet (21% dry matter basis) as such that the overall crude protein content of the feed basket remained constant at 18±0.5% crude protein on a DM basis. The inclusion of MP was compared with the addition of fish meal added to the basal diet at 24% (dry matter basis). Pigs fed with MP were found to have higher net protein utilization rates and apparent and true digestibility of amino acids than the pigs fed with fish meal.
	10% – 20% (dry matter basis)	³¹	In this study, the impact of 10% and 20% inclusion of MP in diets offered to 27 piglets from 21 – 42 days of age and compared with diets with fish meal. The diets were characterized by a slight increase in overall crude protein content of the feed baskets from 21±1% to 27±1% crude protein content on a dry matter basis when increasing the MP replacement share from 10 to 20%, respectively. No difference in growth rate was observed with the piglets converting equally efficient as piglets offered diets containing fish meal.
	5%, 10.5%, 16% (dry matter basis)	³²	In this study, soy bean meal was replaced in basal diet (which comprised of soy, wheat and barley) by MP (methane based production of bacterial protein) at three replacement rates (i.e. 5%, 10.5% and 16% of DM basis) as such that the overall crude protein content of the feed basket remained constant at 22±0.5% crude protein (DM basis). No increase in plasma concentrations was observed, indication the absence of any uricogenic effect.
	4%, 8%, 12% (dry matter basis)	³³	In this study, the effects of replacing fish meal, soy bean meal, and meat and bone meal in conventional diets with MP (methane based production of bacterial protein) at a replacement rate of 4%, 8% and 12% on the growth of weanling pigs examined. Similar growth performance to that obtained with a conventional diet was observed. Note that that the overall crude protein content of the feed basket within the different feed baskets remained constant at 23±0.5% crude protein on a DM basis.
	5%, 10%, 15% (dry matter basis)	³⁴	Feeding trials were conducted to investigate the effects of increasing the dietary content of MP (methane based production of bacterial protein) on the protein and energy metabolism of pigs from weaning to a weight of 80 kg. Soya-bean meal was replaced with MP at a replacement rate ranging from 0 – 15% (DM basis) as such that the overall crude protein content of the feed basket remained constant at 21.8±0.7% crude protein content on a DM basis. Both the overall protein and energy metabolism in growing pigs were not affected in all MP replacement rates used.
	17%, 20%, 35%, 40%, 52%, 60% (digestible N basis)	³⁵	In this study, feeding trials were conducted in which experimental diets contained increasing levels of MP (methane based production of bacterial protein) as replacement of soybean meal as such that the

			overall crude protein content of the feed basket remained constant at 38.3±0.5% crude protein content on a DM basis. It was found that MP was a suitable protein source without any significant differences compared to the soybean meal diet.
	6%, 12%, 15% (dry matter basis)	³⁶	In this study, it was found that replacement of soy beans by MP at replacement rates up to 12% (dry weight basis) in diets for pigs from 26 kg live weight until slaughter had no adverse effect on their overall growth performance as well as carcass lean or fat content. However, MP levels of up to 15% reduced growth rates during the piglet period and increased carcass fat content. The latter was found to be caused by marginal dietary lysine levels. Overall, it was concluded that the addition of MP achieved a dose dependent improvement in the utilization of total amino acids and lysine and the quality of back fat determined as fat firmness and fat color. Note that the overall crude protein content of the feed basket within the different feed baskets remained constant at 20±0.7% crude protein on a DM basis.
	5%, 10%, 15% (dry matter basis)	³⁷	In this study, MP replacement rates and dietary compositions were used according to. ^{34, 36} It was found that the inclusion up to 15% MP on a dry matter basis did not affect the metabolic function, as reflected in the measured blood parameters.
	5%, 10%, 15% (dry matter basis)	³⁸	In this study, feeding trials were conducted in which experimental diets contained increasing levels of MP (methane based production of bacterial protein) as replacement of soybean meal. The inclusion of MP was balanced by a reduction in amount of soybean in the diets in order to maintain constant crude protein levels within the different diets. Pigs fed diets containing MP had reduced thiobarbituric acid reactive substances (TBARS) value in backfat and muscle, reduced intensity of odor and rancid odor and taste in pork after short-time storage, and reduced off-odor and off-taste after intermediate-time storage. To conclude, adding MP to diets for pigs changed the fatty acid profile, improved the oxidative stability, and sensory quality of pork.
	23% (dry matter basis)	³⁹	Feeding trials were conducted to determine the digestibility of MP grown on methane. MP was added at a replacement rate of 23% to the basal diet. Total tract apparent digestibility and the ileal nitrogen digestibility were found to be 85.4% and 78.1%, respectively.
<i>In the MAGPIE model simulations, we used 4%, 8% and 15% of MP replacements shares to the feed baskets for pigs as low, default and ambitious feeding ratio, respectively. Replacement shares are indicated on dry matter basis of the total feed basket.</i>			
Beef and dairy cattle	0, 10, 20% (dry matter basis)	⁴⁰	During feeding trials over a period of 32 weeks it was found that replacing groundnut oil meal with up to 20% MP as protein source in the feed baskets had no detrimental effects on performance in calves.
	0%, 5%, 7.5% (dry matter basis)	⁴¹	Feeding trials using veal calves were conducted using MP at replacement rates of 2.5, 5 and 7.5% as a replacement for milk protein as such that the overall crude protein content of the feed basket remained constant at 25.5±0.7% crude protein content on a DM basis. Similar growth rates during the course of the fattening period (i.e.

			increase in weight of the veal calves from 60 – 150 kg) at all MP replacement shares.
	0, 5, 10, 15% (dry matter basis)	42	Feeding trials were carried out to determine the impact of yeast, grown on methanol, as a protein source using 4 – 5 months old fattening lambs. Similar performance was observed amongst the range of yeast inclusion tested.
	0%, 10.2% and 22.9% (dry matter basis)	43	In this study, the impact of inclusion of yeast derived MP from sugar cane, as a replacement of soy bean, on dry matter intake and digestibility, milk production and quality was examined in dairy goats. During the 90 day feeding trial, it was found that MP can be used as a protein supplement without any differences in milk production and quality observed. The different diets were characterized by a constant crude protein content of 23.8±0.9% on a dry matter basis.
	Not reported	44	The replacement of groundnut-cake protein with MP significantly (P<0.05) increased the milk yield in lactating dairy goats without observing any changes in the milk composition.
	8% (dry matter basis)	45	In 1995 the European Union approved the use of bacterial protein grown on natural gas as protein source in veal calves up to a replacement share of 8% (Council Directive No. 82/471/EEC).
<i>In the MAGPIE model simulations, we used 4%, 8% and 15% of MP replacements shares to the feed baskets for beef cattle and dairy cattle as low, default and ambitious feeding ratio, respectively. Replacement shares are indicated on dry matter basis of the total feed basket.</i>			
Broiler chicken and laying hens	2%, 4%, 6%, 8%, 10% (dry matter basis)	46	Overall, substitution of soybean meal protein with increasing levels of MP significantly lowered feed-to-gain ratio during the last part of the feeding period. Sensory analysis of thigh meat after 2 month of frozen storage showed that meat from 35-d-old chickens fed with inclusion of 6% and 10% MP in their diets had less odor intensity and less rancid flavor than meat from the control group. Other sensory attributes were not affected by treatment. Note that that the overall crude protein content of the different diets remained constant at 22.7±0.4% crude protein on a DM basis.
	6% (dry matter basis)	47	It was concluded that 6% of MP (methane based production of bacterial protein) can replace soybean meal in diets for broiler chickens without impairing growth performance. The overall crude protein content of the different diets remained constant at 26.3±1% crude protein on a DM basis.
	2%, 4%, 6% (dry matter basis)	48	The effects of replacing soybean meal or fish meal with 2, 4 or 6% MP on growth performance, digestibility of amino acids and sensory quality of meat, were examined using 630 broiler chickens. It was concluded that MP can replace soybean meal or fish meal protein in broiler chicken diets within the inclusion rates tested (i.e. up to 6%). Diets were characterized by a constant crude protein content of 23±1% on a dry matter basis.
	9.6%, 10%, 19.2%, 20%, 29% (dry matter basis)	49	It was found that the inclusion of 10% – 20% of MP as a replacement of soybean meal to 5 day-old male broiler chicks reduced growth rate and efficiency of food conversion over a 14 days period. In the same study, it was found that MP fed to chicks (with an age of 21 days) at a

			rate of 9.6% marginally improved growth rates, efficiencies of food utilization and nitrogen retention. Further increasing the MP inclusion rate up to 19.2 and 29% caused adverse effects. The different diets tested were characterized by constant crude protein contents on a dry matter basis.
	2%, 4%, 6% (dry matter basis)	50	In this study, the effect of the inclusion of MP to basal diet with increasing concentrations (0%, 4% and 6 %) on the energy metabolism and carcass composition was investigated. It was concluded that the overall protein and energy metabolism as well as carcass composition were not influenced by a dietary content of up to 6% MP. The latter corresponds to 20% of dietary nitrogen. The different diets tested were characterized by a constant crude protein content of 24.4±1% on a dry matter basis.
	4%, 8%, 12% (dry matter basis)	51	In this study, the effects of the inclusion of MP (methane based production of bacterial protein) to basal diet as a replacement of soybean meal with increasing concentrations (4%, 8% and 12%) on growth performance and carcass quality in broiler chickens were examined. reduced feed intake and improved gain, but did not affect weight gain compared to the soybean meal based control diet. The different diets were characterized by a constant crude protein content of 25.5±0.5% on a dry matter basis.
	4%, 8%, 12% (dry matter basis)	52	In this study, the effects of the inclusion of MP to basal diet with increasing concentrations (4%, 8% and 12% on fatty acid composition, the profile of volatiles by dynamic headspace gas chromatography-mass spectrometry, and sensory quality of frozen-stored broiler chicken thigh meat was examined. In the basal diet, containing e.g. wheat and maize), soy bean meal was used as protein source. Replacing soy bean by MP resulted in a reduction in lipid oxidation products in frozen-stored meat. The latter is important, as it is associated with quality deterioration and reduced consumer acceptance. The different diets tested were characterized by a constant crude protein content of 26.9±0.8% (DM basis).
	8% (dry matter basis)	53	In this study, the effects of bacterial protein meal at an 8% replacement rate (as a replacement of soybean meal) to broiler chickens (1–35 days of age) on the fatty acid composition, lipid oxidation and sensory quality on frozen thigh meat stored frozen for 6 month was examined. It was found that the inclusion of MP did not affect the sensory quality parameters, but had a positive effect in terms of reduced volatiles in frozen-stored broiler meat. The latter was hypothesized to be related to antioxidant properties of the bacterial autolysate. The different diets tested were characterized by a constant crude protein content of 26.5±0.85% (DM basis).
<i>In the MAGPIE model simulations, we used 3%, 6% and 12% of MP replacements shares to the feed baskets for broiler chicken and laying hen as low, default and ambitious feeding ratio, respectively. Replacement shares are indicated on dry matter basis of the total feed basket.</i>			

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372 **Table S2.** Environmental costs of reactive nitrogen losses to the biosphere.

Region	Annual costs	Costs (per kg N)	ref	Remarks
China	US\$9.5–31 billion	\$2.1–3.8 kg NO _x -N \$0.4–3.3 kg NH ₃ -N \$0.3 kg N ₂ O-N	⁵⁴	This study estimates that agriculture accounted for 95% of the NH ₃ and 51% of the N ₂ O in china in the year 2008. In the same study it was also estimated that the total atmospheric emissions of reactive nitrogen causing related health damage ranged US\$19–62 billion per year. Of this number, agricultural induced emissions accounted for more than 50% of the costs.
EU	€35–230 billion	€10–30kg NO _x -N €2–20kg NH ₃ -N €5–15kg N ₂ O-N	⁵⁵	This study revealed that the costs of agricultural induced reactive nitrogen losses exceed the economic benefit due to increase primary crop production by a factor 4. Overall, the annual costs associated with agricultural reactive nitrogen losses was estimated to range between €35–230 billion per year.
USA	US\$81–\$441 billion	No information given on costs per kg N	⁵⁶	This study is the first assessing the cost associated with reactive nitrogen losses to the biosphere from human activities in the United States. The study revealed that the total potential environmental and health economic impact of reactive nitrogen losses from anthropogenic nitrogen summed up to an average of US\$210 (\$81–\$441) billion per year in the beginning of the 21 th century. Of this, ~75% of the estimated costs were associated with agricultural induces losses.
World	US\$200–2000 billion	No information given on costs per kg N	⁵⁷	In this report, conducted by the European Nitrogen Assessment, a costing procedure based on the European situation was implemented aiming at calculating the global cost of nitrogen pollution. Taking into account that the global costs would be approximately a factor threefold of the European situation, resulting in an overall estimated costs associated with reactive nitrogen losses ranging between 200 to 2000 billion US dollars annually.
USA		\$900 (\$100 – \$59,400) ton NH ₃ 250 (\$20–\$1780) ton NO _x	⁵⁸	In this study, the Air Pollution Emission Experiments and Policy (APEEP) model (an integrated assessment model) was used to determine the economic impact of air pollution by means of air quality modeling, exposure, dose-response and valuation for a large range of point sources, based on data of more than 10,000 sources measured by the US EPA.

USA		\$3.03 (\$1.25 – \$4.80) kg NH ₃ 14.6 (\$2.0-\$27.27) kg NO _x	⁵⁹	This study aimed to determine the environmental and health externalities associated with the production of different agricultural crops such as corn and switch grass for the production of ethanol. While the purposed and crops used are different, the externalities are directly assessed based on the emissions of NH ₃ and NO _x .
USA and EU		€3.1 – €30 kg NH ₃ -N (to air) €13 – €43 kg NO _x -N (to air) €5 – €54 kg Nr (to water) €2-18kg N ₂ O-N (to air)	⁶⁰	In this study, the findings of several previous studies ⁶¹⁻⁶³ on the externalities of reactive nitrogen emissions in terms of health, ecosystems/coastal systems, crop decline and climate change were summarized.

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375 **Table S3.** Overview of the substrate requirements of the different technological pathways for
 376 production of Microbial Protein.

MP Substrate		ton substrate / ton MP-N	ton substrate / ton MP (@ 70% protein)	Microbial route
High C/N crops	Maize (dry weight)	49.6	5.6	Methane oxidation (via biogas)
	Miscanthus (dry weight)	50.2	5.5	Hydrogen oxidation (via biogas)
	Sugar cane (dry weight)	34.9	4.3	Organic carbon oxidation (fermentation)
Natural gas	CH ₄ gas	9.2	1.01	Methane oxidation (direct assimilation)
Hydrogen gas	H ₂ gas	4.0	0.46	Hydrogen oxidation (direct assimilation)

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378 **Table S4.** General assumptions for the economic assessment of the different technological
379 pathways for MP production

Parameter	Value	Unit	Refs.	Remarks
Electricity price	0.10	\$/kWh		Price may have regional fluctuations and may be as low as \$0.05/kWh. Note that levelized costs for wind energy are already as low as \$0.06/kWh ⁶⁴ .
Natural gas price	0.17-0.33	\$/Nm ³	⁶⁵	The prices corresponding to 5 and 10 \$/mmbtu, respectively. 10 \$/mmbtu is the current and predicted future price in Europe, whereas the current price and forecasted price in the US are lower than 5 \$/mmbtu.
Sugar price	250	\$/ton	⁶⁶	2005 prices. The same report, prepared by the US department of Agriculture, it is forecasted that raw sugar price will increase to around \$440 by 2025.
Soy price	400-600	\$/ton	⁶⁷	Note that soy has a protein content of around 32 – 40%, compared to 70-75% for MP. Prices substantially differ between regions.
Fishmeal price	1750	\$/ton	⁶⁸	The price steadily rising since 2000 and expected to double by 2030.
Market value MP	1750	\$/ton		The market value of MP is estimated to be the same as fishmeal as it has a similar protein and amino acid composition, however the protein MP protein content is higher than fishmeal (up to 75% versus 65% for MP and fishmeal respectively).

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Table S5. Levelized costs for hydrogen production including CAPEX and OPEX and subsequent impact on the operational expenditure.

Hydrogen production route	Costs (\$/kg H ₂)	Costs (\$/ton MP) ^{d)}	Costs (\$/ton MP-protein) ^{d,e)}	Remarks
Biomass gasification	\$1.61	902	\$1288	The cost for biomass gasification are based on study of the US department of Energy. ⁶⁹ The same study reports on future prediction of a further reduction in costs to \$1.47/kg H ₂ using the H2A production model. ⁷⁰ This model includes capital, operating, maintenance, feedstock, utility, transport and replacement costs.
Water electrolysis (current technology)	\$5.00	2800	\$4000 ^{c)}	The costs for hydrogen production by means of PEM electrolysis are based on an electricity price of \$0.05/kWh. ⁷¹
Water electrolysis (future predictions)	\$3.00	1680	2100	Future predicted levelized costs for hydrogen production using PEM electrolysis are predicted at \$3/kg H ₂ produced. ⁷¹ Note that this also includes compression, storage and dispensing, which are not needed for the production of MP.
Water electrolysis (off-peak electricity)	\$0.70	392	\$560 ^{c)}	The predicted future levelized cost are \$3/kg H ₂ produced. ⁷¹ Electricity price of \$0.05/kWh or \$2.3/kg, corresponding to \$0.7/kg H ₂ , including compression, storage and dispensing.

- a) Costs for delivery, compression, storage, and dispensing not required producing MP.
- b) Calculated costs per ton MP based on stoichiometric requirements of 4 kg H₂/kg N and conversion of N to protein of 6.25. ⁷²
- c) Additional economic benefits may derive from using CO₂ from industrial point source (e.g. reduction in carbon tax).
- d) We assumed 80% hydrogen uptake efficiencies for MP production.
- e) Final cost expressed per ton of protein (@70% protein content for the final MP product).

393 **Table S6.** Nitrogen and phosphorus requirements for MP production and associated OPEX.

Parameter	Value	Unit	Remarks
Haber-Bosch nitrogen requirements	0.112	ton NH ₃ -N/ton MP	Based on a conversion of nitrogen-N to protein of 6.25 ^{72, 73} , at an average protein content of MP of 70% on a dry weight basis.
Price of Haber-Bosch nitrogen	800	\$/ton N	Food-grade ammonia nitrogen is used.
Costs NH ₃ -N per ton MP produced	89.6	\$/ton MP	N/A
Phosphate required	32	kg PO ₄ /ton MP	N/A
Price phosphoric acid	1000	\$/ton	Food grade phosphoric acid used.
Costs phosphate per ton MP produced	32	\$/ton MP	N/A
Subtotal N and P costs	121.6	\$/ton MP	N/A

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Table S7. Oxygen requirements and generation of industrial grade oxygen for MP production and associated OPEX

Production platform	Oxygen Required		Oxygen Cost		Remarks
	Value	Unit	Value	Unit	
Hydrogen	2.05	kg O ₂ /kg MP	90	\$/ton MP	Based on equation 1, Table S10.
Methane	2.50	kg O ₂ /kg MP	109	\$/ton MP	Based on equation 2, Table S10.
Heterotrophic	0.85	kg O ₂ /kg MP	37	\$/ton MP	Based on equation 3, Table S10.
Industry grade oxygen generation	0.05 (0.04-0.06)	\$/m ³ O ₂	N/A	N/A	Based on ⁷⁴ and electricity price of \$0.10/kWh
	35 (28-42)	\$/ton O ₂			
Oxygen utilization efficiency	80	%	N/A	N/A	N/A

Table S8. Energy requirements dewatering, sterilization and drying and associated OPEX.

Operation	Parameter	Value	Unit
Centrifugation	Energy requirement for centrifugation	700	kWh/ton MP
	Electricity price	0.10	\$/kWh
	Dry solids content MP after centrifuge	25	wt %
	Energy costs per ton dry MP	70	\$/ton MP
Spray-drying	Energy requirements spray-drying ^{a)}	3500 ²⁹	MJ/ton H ₂ O
	Energy content natural gas	38.7	MJ/m ³
	Natural gas required	90	m ³ /ton H ₂ O
	Price natural gas	0.33 (0.17-0.33) ⁶⁵	\$/m ³
	Dry solids content final product	100%	wt%
	Energy requirements per ton dry MP	8167	MJ/ton MP
		211	m ³ CH ₄ /ton MP
	Energy costs	\$90	\$/ton MP

^{a)} Referred to spray drying integrated with a fluidized bed, a common method in the food industry. ²⁹

Table S9. Summary of operational expenditure (OPEX) for the MP production pathways. Table reports the production strategies (*Production technology*), direct expenses for production (Production *cost* also referred to as OPEX) and the total costs (*Total costs*) for producing MP from 3 different technologies (*production pathways*) using the reported carbon and energy sources (*Production pathway* and *type of substrate*); the cost section reports expenses for substrate (*Substrate*) and other demands such as secondary substrates (*Nitrogen & Phosphate*), gassing (*O₂* and *CO₂*), running costs for fermenting (*Mixing & pumping*) and final processing (*Dewatering* and *Drying/sterilization*). Nitrogen and Phosphate and overhead costs were assumed to be the same for all conditions.

Production technology		Production costs, OPEX									Total costs	
Production pathway	Type of Substrate	Annuity of capital expenditure	Substrate	O ₂	Nitrogen & Phosphate	CO ₂	Mixing & pumping	Dewatering	Drying / sterilization	Overhead	(70% protein) [\$/ton MP]	(100% protein) [\$/ton MP]
High C/N crops	Methane (after energy maize anaerobic digestion)	\$203	\$831	\$109	\$149	N/A	\$38	\$70	\$90	\$203	\$1,692	\$2,417
	Hydrogen (cellulose gasification - current price scenario)		\$886	\$90		N/A	\$38	\$70	\$90		\$1,727	\$2,466
	Hydrogen (biomass gasification - target price scenario)		\$666	\$90		N/A	\$38	\$70	\$90		\$1,507	\$2,153
	Sugar cane (dry weight)		\$504	\$37		N/A	\$38	\$70	\$90		\$1,293	\$1,848
Natural gas	CH ₄ gas (low price scenario)		\$300	\$109		N/A	\$38	\$70	\$90		\$1,162	\$1,660
	CH ₄ gas (high price scenario)		\$583	\$109		N/A	\$38	\$70	\$90		\$1,444	\$2,064
Hydrogen	H ₂ gas (water electrolysis - standard price)		\$1650	\$90		\$117	\$38	\$70	\$90		\$2,725	\$3,894
	H ₂ gas (water electrolysis - off peak energy)		\$385	\$90		\$117	\$38	\$70	\$90		\$1,460	\$2,086

Table S10. Stoichiometry of MP production normalized per mole of biomass. Same biomass composition and nitrogen assimilation efficiency were assumed for MP produced from hydrogen or methane oxidation.

Feedstock	Reaction stoichiometric	Anabolic products (mol)	Catabolic products (mol)	References
Hydrogen oxidation	$21.36H_2 + 6.21O_2 + 4.09CO_2 + 0.76NH_3$	$C_{4.09}H_{7.13}O_{1.89}N_{0.76}$	$18.70 H_2O$	⁷⁵
Methane oxidation	$7.59 O_2 + 6.13 CH_4 + 0.76 NH_3$	$C_{4.09}H_{7.13}O_{1.89}N_{0.76}^a$	$2.69 g CO_2$	⁷⁶
Organic carbon oxidation	$0.67 C_{12}H_{22}O_{11} + 3.00 O_2 + 1.00 NH_3$	$C_5H_7O_2N$	$3.00 CO_2 + 5.33 H_2O$	⁷⁷

^{a)} Assuming same biomass composition for autotrophic hydrogen MP production; an equivalent amount of nitrogen was included in the stoichiometry for biomass formation from methane.

References

1. Bodirsky, B. L.; Popp, A.; Lotze-Campen, H.; Dietrich, J. P.; Rolinski, S.; Weindl, I.; Schmitz, C.; Müller, C.; Bonsch, M.; Humpenöder, F.; Biewald, A.; Stevanovic, M., Reactive nitrogen requirements to feed the world in 2050 and potential to mitigate nitrogen pollution. *Nature Communications* **2014**, *5*.
2. Bodirsky, B. L.; Popp, A.; Weindl, I.; Dietrich, J. P.; Rolinski, S.; Scheffele, L.; Schmitz, C.; Lotze-Campen, H., N₂O emissions from the global agricultural nitrogen cycle-current state and future scenarios. *Biogeosciences* **2012**, *9*, (10), 4169-4197.
3. Humpenöder, F.; Popp, A.; Dietrich, J. P.; Klein, D.; Lotze-Campen, H.; Bonsch, M.; Bodirsky, B. L.; Weindl, I.; Stevanovic, M.; Müller, C., Investigating afforestation and bioenergy CCS as climate change mitigation strategies. *Environmental Research Letters* **2014**, *9*, (6).
4. Popp, A.; Humpenöder, F.; Weindl, I.; Bodirsky, B. L.; Bonsch, M.; Lotze-Campen, H.; Müller, C.; Biewald, A.; Rolinski, S.; Stevanovic, M.; Dietrich, J. P., Land-use protection for climate change mitigation. *Nature Clim. Change* **2014**, *4*, (12), 1095-1098.
5. Popp, A.; Lotze-Campen, H.; Bodirsky, B., Food consumption, diet shifts and associated non-CO₂ greenhouse gases from agricultural production. *Global Environmental Change* **2010**, *20*, (3), 451-462.
6. Lotze-Campen, H.; Müller, C.; Bondeau, A.; Rost, S.; Popp, A.; Lucht, W., Global food demand, productivity growth, and the scarcity of land and water resources: A spatially explicit mathematical programming approach. *Agricultural Economics* **2008**, *39*, (3), 325-338.
7. Popp, A.; Calvin, K.; Fujimori, S.; Havlik, P.; Humpenöder, F.; Stehfest, E.; Bodirsky, B. L.; Dietrich, J. P.; Doelmann, J. C.; Gusti, M.; Hasegawa, T.; Kyle, P.; Obersteiner, M.; Tabeau, A.; Takahashi, K.; Valin, H.; Waldhoff, S.; Weindl, I.; Wise, M.; Kriegler, E.; Lotze-Campen, H.; Fricko, O.; Riahi, K.; Vuuren, D. P. v., Land-use futures in the shared socio-economic pathways. *Global Environmental Change* **2017**, *42*, 331-345.
8. Schmitz, C.; Biewald, A.; Lotze-Campen, H.; Popp, A.; Dietrich, J. P.; Bodirsky, B.; Krause, M.; Weindl, I., Trading more food: Implications for land use, greenhouse gas emissions, and the food system. *Global Environmental Change* **2012**, *22*, (1), 189-209.
9. Valin, H.; Sands, R. D.; van der Mensbrugghe, D.; Nelson, G. C.; Ahammad, H.; Blanc, E.; Bodirsky, B.; Fujimori, S.; Hasegawa, T.; Havlik, P.; Heyhoe, E.; Kyle, P.; Mason-D'Croz, D.; Paltsev, S.; Rolinski, S.; Tabeau, A.; van Meijl, H.; von Lampe, M.; Willenbockel, D., The future of food demand: Understanding differences in global economic models. *Agricultural Economics (United Kingdom)* **2014**, *45*, (1), 51-67.
10. Bodirsky, B. L.; Rolinski, S.; Biewald, A.; Weindl, I.; Popp, A.; Lotze-Campen, H., Global Food Demand Scenarios for the 21st Century. *PLoS ONE* **2015**, *10*, (11).
11. Weindl, I.; Bodirsky, B. L.; Rolinski, S.; Biewald, A.; Lotze-Campen, H.; Müller, C.; Dietrich, J. P.; Humpenöder, F.; Stevanović, M.; Schaphoff, S.; Popp, A., Livestock production and the water challenge of future food supply: Implications of agricultural management and dietary choices. *Global Environmental Change* **2017**, *47*, 121-132.
12. Wirsenius, S. Human Use of Land and Organic Materials: Modeling the Turnover of Biomass in the Global Food System. Chalmers University of Technology, Gothenburg, Sweden, 2000.
13. <http://faostat.fao.org/>.
14. Schmitz, C.; Lotze-Campen, H.; Gerten, D.; Dietrich, J. P.; Bodirsky, B.; Biewald, A.; Popp, A., Blue water scarcity and the economic impacts of future agricultural trade and demand. *Water Resources Research* **2013**, *49*, (6), 3601-3617.

15. Dietrich, J. P.; Schmitz, C.; Lotze-Campen, H.; Popp, A.; Müller, C., Forecasting technological change in agriculture-An endogenous implementation in a global land use model. *Technological Forecasting and Social Change* **2014**, *81*, (1), 236-249.
16. Lotze-Campen, H.; Popp, A.; Beringer, T.; Müller, C.; Bondeau, A.; Rost, S.; Lucht, W., Scenarios of global bioenergy production: The trade-offs between agricultural expansion, intensification and trade. *Ecological Modelling* **2010**, *221*, (18), 2188-2196.
17. Popp, A.; Dietrich, J. P.; Lotze-Campen, H.; Klein, D.; Bauer, N.; Krause, M.; Beringer, T.; Gerten, D.; Edenhofer, O., The economic potential of bioenergy for climate change mitigation with special attention given to implications for the land system. *Environmental Research Letters* **2011**, *6*, (3).
18. Dietrich, J. P.; Popp, A.; Lotze-Campen, H., Reducing the loss of information and gaining accuracy with clustering methods in a global land-use model. *Ecological Modelling* **2013**, *263*, 233-243.
19. Krause, M.; Lotze-Campen, H.; Popp, A.; Dietrich, J. P.; Bonsch, M., Conservation of undisturbed natural forests and economic impacts on agriculture. *Land Use Policy* **2013**, *30*, (1), 344-354.
20. Bondeau, A.; Smith, P. C.; Zaehle, S.; Schaphoff, S.; Lucht, W.; Cramer, W.; Gerten, D.; Lotze-campen, H.; Müller, C.; Reichstein, M.; Smith, B., Modelling the role of agriculture for the 20th century global terrestrial carbon balance. *Global Change Biology* **2007**, *13*, (3), 679-706.
21. Müller, C.; Robertson, R. D., Projecting future crop productivity for global economic modeling. *Agricultural Economics (United Kingdom)* **2014**, *45*, (1), 37-50.
22. 2006, I., 2006 IPCC Guidelines for National Greenhouse Gas Inventories, Prepared by the National Greenhouse Gas Inventories Programme. In *2006 IPCC Guidelines for National Greenhouse Gas Inventories, Prepared by the National Greenhouse Gas Inventories Programme*, Eggleston H.S., B. L., Miwa K., Ngara T. and Tanabe K. (eds), Ed. The Institute for Global Environmental Strategies (IGES): Hayama, Japan, 2006 Vol. Volume 4.
23. (IPCC), I. P. o. C. C., Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories: Workbook. In *Institute for Global Environmental Strategies 1996; Vol. Volume 2*.
24. Hurtt, G. C.; Frohling, S.; Fearon, M. G.; Moore, B.; Shevliakova, E.; Malyshev, S.; Pacala, S. W.; Houghton, R. A., The underpinnings of land-use history: Three centuries of global gridded land-use transitions, wood-harvest activity, and resulting secondary lands. *Global Change Biology* **2006**, *12*, (7), 1208-1229.
25. Hurtt, G. C.; Chini, L. P.; Frohling, S.; Betts, R. A.; Feddema, J.; Fischer, G.; Fisk, J. P.; Hibbard, K.; Houghton, R. A.; Janetos, A.; Jones, C. D.; Kindermann, G.; Kinoshita, T.; Klein Goldewijk, K.; Riahi, K.; Shevliakova, E.; Smith, S.; Stehfest, E.; Thomson, A.; Thornton, P.; van Vuuren, D. P.; Wang, Y. P., Harmonization of land-use scenarios for the period 1500-2100: 600 years of global gridded annual land-use transitions, wood harvest, and resulting secondary lands. *Climatic Change* **2011**, *109*, (1), 117-161.
26. FAO *Global forest resources assessment 2010: Main report.* ; Food and Agriculture Organization of the United Nations: Rome, 2010.
27. Niazi, S. K.; Brown, J. L., *Fundamentals of Modern Bioprocessing*. CRC Press: 2015.
28. Anupama; Ravindra, P., Value-added food: Single cell protein. *Biotechnology Advances* **2000**, *18*, (6), 459-479.
29. Chen, X. D.; Mujumdar, A. S., *Drying Technologies in Food Processing*. In 1 ed.; Wiley: Hoboken, 2009.
30. D'Mello, J. P. F.; Peers, D. G.; Whittemore, C. T., Utilization of dried microbial cells grown on methanol in a semi purified diet for growing pigs. *British Journal of Nutrition* **1976**, *36*, (3), 403-410.

31. Whittemore, C. T.; Moffat, I. W.; Taylor, A. G., Evaluation by digestibility, growth and slaughter of microbial cells as a source of protein for young pigs. *Journal of the Science of Food and Agriculture* **1976**, 27, (12), 1163-1170.
32. Hellwing, A. L. F.; Tauson, A. H.; Skrede, A., Excretion of purine base derivatives after intake of bacterial protein meal in pigs. *Livestock Science* **2007**, 109, (1-3), 70-72.
33. Øverland, M.; Skrede, A.; Matre, T., Bacterial protein grown on natural gas as feed for pigs. *Acta Agriculturae Scandinavica - Section A: Animal Science* **2001**, 51, (2), 97-106.
34. Hellwing, A. L. F.; Tauson, A. H.; Kjos, N. P.; Skrede, A., Bacterial protein meal in diets for growing pigs: Effects on protein and energy metabolism. *Animal* **2007**, 1, (1), 45-54.
35. Hellwing, A. L. F.; Tauson, A. H.; Skrede, A.; Kjos, N. P.; Ahlstrøm, Ø., Bacterial protein meal in diets for pigs and minks: Comparative studies on protein turnover rate and urinary excretion of purine base derivatives. *Archives of Animal Nutrition* **2007**, 61, (6), 425-443.
36. Øverland, M.; Kjos, N. P.; Skrede, A., Effect of bacterial protein meal grown on natural gas on growth performance and carcass traits of pigs. *Italian Journal of Animal Science* **2004**, 3, (4), 323-336.
37. Hellwing, A. L. F.; Tauson, A. H.; Skrede, A., Blood parameters in growing pigs fed increasing levels of bacterial protein meal. *Acta Veterinaria Scandinavica* **2007**, 49, (1).
38. Øverland, M.; Kjos, N. P.; Olsen, E.; Skrede, A., Changes in fatty acid composition and improved sensory quality of backfat and meat of pigs fed bacterial protein meal. *Meat Science* **2005**, 71, (4), 719-729.
39. Skrede, A.; Berge, G. M.; Storebakken, T.; Herstad, O.; Aarstad, K. G.; Sundstøl, F., Digestibility of bacterial protein grown on natural gas in mink, pigs, chicken and Atlantic salmon. *Animal Feed Science and Technology* **1998**, 76, (1-2), 103-116.
40. Desai, H. B.; Shukla, P. C., Effect of feeding single-cell protein(SCP) on dry-matter intake, nutrients intake and efficiency of feed utilization. *Indian Journal of Animal Sciences* **1988**, 58, (4), 504-507.
41. Roth, F. X.; Kirchgessner, M., Methanol-fermentation protein in veal calf nutrition. *Animal Feed Science and Technology* **1976**, 1, (1), 33-44.
42. Rammo, S. a. J. S. *Study of the effect of single cell protein for lambs nutrition*; 1985, 1985.
43. de Lima, L. S.; Alcalde, C. R.; Freitas, H. S.; Molina, B. S. L.; de Macedo, F. A. F.; Horst, J. A., Performance of dairy goats fed diets with dry yeast from sugar cane as protein source. *Revista Brasileira de Zootecnia* **2012**, 41, (1), 232-236.
44. Mudgal, V. D.; Vijaygopal, P.; Singhal, Comparison of urea and single-cell protein as a protein supplement of lactating goats. *Indian Journal of Animal Sciences* **1986**.
45. Union, E., In *Council Directive No. 82/471/EEC*, Union, E., Ed. 2005.
46. Skrede, A.; Faaland Schøyen, H.; Svihus, B.; Storebakken, T., The effect of bacterial protein grown on natural gas on growth performance and sensory quality of broiler chickens. *Canadian Journal of Animal Science* **2003**, 83, (2), 229-237.
47. Schøyen, H. F.; Hetland, H.; Rouvinen-Watt, K.; Skrede, A., Growth performance and ileal and total tract amino acid digestibility in broiler chickens fed diets containing bacterial protein produced on natural gas. *Poultry Science* **2007**, 86, (1), 87-93.
48. Schøyen, H. F.; Svihus, B.; Storebakken, T.; Skrede, A., Bacterial protein meal produced on natural gas replacing soybean meal or fish meal in broiler chicken diets. *Archives of Animal Nutrition* **2007**, 61, (4), 276-291.
49. D'mello, J. P. F.; Acamovic, T., Evaluation of Methanol-Grown Bacteria as a Source of Protein and Energy for Young Chicks. *British Poultry Science* **1976**, 17, (4), 393-401.
50. Hellwing, A. L. F.; Tauson, A. H.; Skrede, A., Effect of bacterial protein meal on protein and energy metabolism in growing chickens. *Archives of Animal Nutrition* **2006**, 60, (5), 365-381.

51. Øverland, M.; Schøyen, H. F.; Skrede, A., Growth performance and carcass quality in broiler chickens fed on bacterial protein grown on natural gas. *British Poultry Science* **2010**, *51*, (5), 686-695.
52. Øverland, M.; Borge, G. I.; Vogt, G.; Schøyen, H. F.; Skrede, A., Oxidative stability and sensory quality of meat from broiler chickens fed a bacterial meal produced on natural gas. *Poultry Science* **2011**, *90*, (1), 201-210.
53. Øverland, M.; Skrede, A., Fatty acid composition, oxidative stability and sensory quality of meat from broiler chicken fed autolysate from bacteria grown on natural gas. *Journal of Animal Physiology and Animal Nutrition* **2012**, *96*, (4), 747-754.
54. Gu, B.; Ge, Y.; Ren, Y.; Xu, B.; Luo, W.; Jiang, H.; Gu, B.; Chang, J., Atmospheric Reactive Nitrogen in China: Sources, Recent Trends, and Damage Costs. *Environmental Science & Technology* **2012**, *46*, (17), 9420-9427.
55. Van Grinsven, H. J. M.; Holland, M.; Jacobsen, B. H.; Klimont, Z.; Sutton, M. A.; Jaap Willems, W., Costs and benefits of nitrogen for Europe and implications for mitigation. *Environmental Science and Technology* **2013**, *47*, (8), 3571-3579.
56. Sobota, D. J.; Compton, J. E.; McCrackin, M. L.; Singh, S., Cost of reactive nitrogen release from human activities to the environment in the United States. *Environmental Research Letters* **2015**, *10*, (2).
57. Sutton, M. A.; Bleeker, A.; Howard, C. M.; Bekunda, M.; Grizzetti, B.; de Vries, W.; van Grinsven, H. J. M.; Abrol, Y. P.; Adhya, T. K.; Billen, G.; Davidson, E. A.; Datta, A.; Diaz, R.; Erisman, J. W.; Liu, X. J.; Oenema, O.; Palm, C.; Raghuram, N.; Reis, S.; Scholz, R. W.; Sims, T.; Westhoek, H.; Zhang, F. S., *Our nutrient world: the challenge to produce more food and energy with less pollution*. Centre for Ecology & Hydrology on behalf of the Global Partnership on Nutrient Management (GPNM) and the International Nitrogen Initiative (INI): 2013.
58. Muller, N. Z.; Mendelsohn, R. O., Weighing the Value of a Ton of Pollution. *Regulation* **2010**, Vol. 33, (No. 2), p. 20.
59. Kusiima, J. M.; Powers, S. E., Monetary value of the environmental and health externalities associated with production of ethanol from biomass feedstocks. *Energy Policy* **2010**, *38*, (6), 2785-2796.
60. Erisman, J. W.; Galloway, J. N.; Seitzinger, S.; Bleeker, A.; Dise, N. B.; Petrescu, A. M.; Leach, A. M.; de Vries, W., Consequences of human modification of the global nitrogen cycle. *Philosophical transactions of the Royal Society of London. Series B, Biological sciences* **2013**, *368*, (1621), 20130116.
61. Compton, J. E.; Harrison, J. A.; Dennis, R. L.; Greaver, T. L.; Hill, B. H.; Jordan, S. J.; Walker, H.; Campbell, H. V., Ecosystem services altered by human changes in the nitrogen cycle: A new perspective for US decision making. *Ecology Letters* **2011**, *14*, (8), 804-815.
62. Birch, M. B. L.; Gramig, B. M.; Moomaw, W. R.; Doering, I. I. O. C.; Reeling, C. J., Why Metrics Matter: Evaluating Policy Choices for Reactive Nitrogen in the Chesapeake Bay Watershed. *Environmental Science & Technology* **2011**, *45*, (1), 168-174.
63. Brink, C.; van Grinsven, H. J. M., Costs and benefits of nitrogen in the environment. In *The European Nitrogen Assessment*, Sutton, M. A., Ed. Cambridge University Press: Cambridge, 2011.
64. Lantz, E.; Wiser, R.; Hand, M., IEA Wind Task 26: The Past And Future Cost Of Wind Energy. In Laboratory, N. R. E., Ed. U.S. Department of Energy, Office of Energy Efficiency & Renewable Energy: Colorado, 2012.
65. <http://knoema.com/ncszerf/natural-gas-prices-long-term-forecast-to-2020-data-and-charts>
66. Haley, S. *Projecting World Raw Sugar Prices*; SSSM-317-01; United States Department of Agriculture: 2015.

67. FAO, International Commodity Prices [In 2014.
68. Department, A. a. E. S. *FISH TO 2030 Prospects for Fisheries and Aquaculture*; World Bank: Washington DC, 2013.
69. Ramsden, T.; Steward, D.; Zuboy, J., Analyzing the Levelized Cost of Centralized and Distributed Hydrogen Production Using the H2A Production Model, Version 2. In Laboratory, N. R. E., Ed. U.S. Department of Energy: Oak Ridge, 2009.
70. Steward, D.; Ramsden, T.; Zuboy, J., H2A Production Model, Version 2 User Guide. In Laboratory, N. R. E., Ed. National Renewable Energy Laboratory: Golden, 2008.
71. Ayers, K. E.; Anderson, E. B.; Capuano, C. B.; Carter, B. D.; Dalton, L. T.; Hanlon, G.; Manco, J.; Niedzwiecki, M. In *Research advances towards low cost, high efficiency PEM electrolysis*, ECS Transactions, 2010; 2010; pp 3-15.
72. Vielma, J.; Mäkinen, T.; Ekholm, P.; Koskela, J., Influence of dietary soy and phytase levels on performance and body composition of large rainbow trout (*Oncorhynchus mykiss*) and algal availability of phosphorus load. *Aquaculture* **2000**, *183*, (3-4), 349-362.
73. Schulz, E.; Oslage, H. J., Composition and nutritive value of single-cell protein (SCP). *Animal Feed Science and Technology* **1976**, *1*, (1), 9-24.
74. <http://www.pci-intl.com/>
75. Ishizaki, A.; Tanaka, K., Batch culture of *Alcaligenes eutrophus* ATCC 17697T using recycled gas closed circuit culture system. *Journal of Fermentation and Bioengineering* **1990**, *69*, (3), 170-174.
76. Wilkinson, T. G.; Topiwala, H. H.; Hamer, G., Interactions in a mixed bacterial population growing on methane in continuous culture. *Biotechnology and Bioengineering* **1974**, *16*, (1), 41-59.
77. Park, D. H.-D.; Chang, I.-S.; Lee, K.-W., *Principles of Membrane Bioreactors for Wastewater Treatment*. CRC press: 2015.